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## Toward the Systematic Identification of Microplastics in the Environment

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# Toward the Systematic Identification of Microplastics in the Environment: Evaluation of a New Independent Software Tool (siMPle) for Spectroscopic Analysis

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## Abstract

Microplastics (MP) are ubiquitous within the environment, but the approaches to analysis of this contaminant are currently quite diverse, with a number of analytical methods available. The comparability of results is hindered as even for a single analytical method such as Fourier transform infrared spectroscopy (FT-IR) the different instruments currently available do not allow a harmonized analysis. To overcome this limitation, a new free of charge software tool, allowing the systematic identification of MP in the environment (siMPle) was developed. This software tool allows a rapid and harmonized analysis of MP across FT-IR systems from different manufacturers (Bruker Hyperion 3000, Agilent Cary 620/670, PerkinElmer Spotlight 400, and Thermo Fischer Scientific Nicolet iN10). Using the same database and the automated analysis pipeline in siMPle, MP were identified in samples that were analyzed with instruments with different detector systems as well as optical resolutions and the results discussed.

## Keywords

Microplastic, quantification, identification, FT-IR imaging, automated analysis, Raman, FT-IR, micro-FT-IR, wastewater

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## Introduction

Small plastic particles, called microplastics (MP) are currently recognized as a potential risk for environmental and human health.<sup>1,2</sup> The near-ubiquitous contamination of the environment, from terrestrial soils and air to the freshwater and marine environments has raised the profile of this topic in recent years, leading to a wealth of methods and approaches for sampling and analyzing MP in environmental matrices.<sup>3–11</sup> In general, these particles are defined as <5 mm in size while a lower size limit and a standard definition of MP has yet to be agreed on.<sup>7,12</sup> Subcategories distinguishing between large MP (5 mm–500 µm) and small MP (500–1 µm)<sup>13</sup> are often used, reflecting practical considerations during the full analytical procedure.

The analytical procedure to identify particles can be divided into three steps, starting with sampling for MP, followed by sample extraction and finally, identification and quantification. Each of these steps has its challenges (cf. Lusher et al.,<sup>3</sup> Brander et al.,<sup>4</sup> Primpke et al.,<sup>6</sup> and Cowger et al.<sup>14</sup>). Additionally, there is an increasing

awareness for quality assurance and quality control (QA/QC) to successfully and reliably identify MP in environmental samples.<sup>15,16</sup> Individual steps for QA/QC are currently discussed within this overarching special issue.<sup>4</sup> While QA/QC is important for the quality of the results, the

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intercomparison of studies is further hampered by a missing harmonization of the three steps: sampling, sample extraction, and identification. Especially for the chemical analysis, a plethora of different methods and software are used to interpret generated data.<sup>5,6,14</sup> Among the spectroscopy-based techniques, Fourier transform infrared spectroscopy (FT-IR) is considered most suitable to analyze MP from 5 mm to  $\sim 10\ \mu\text{m}$ , even if it requires different acquisition modes according to the particle size.<sup>6</sup> Micro-FT-IR ( $\mu\text{FT-IR}$ ) imaging<sup>17–20</sup> permits scanning of the whole sample, avoiding the manual sorting/point-and-shoot steps that are otherwise necessary, which are a source of bias as the analysis becomes operator dependent.<sup>21,22</sup> On the other hand, FT-IR imaging for MP analysis produces a huge amount of data (large areas are scanned at a very high spatial resolution) which are difficult to manage using the commercial software provided by the instrument manufacturers.<sup>4</sup> Challenges handling this data, together with the lack of a suitable software tool, not only impedes a reliable workflow but also makes it difficult, if not impossible, to compare between studies performed using FT-IR imaging.<sup>14</sup> While initial studies comparing various analytical methods have been conducted,<sup>23</sup> a comparison between studies using the same analytical technique but on different instruments is missing for MP research. One reason is the lack of a suitable software tool to perform such a comparison. Such a study, however, is essential to reproduce and compare results within meta-studies. Moreover, it allows the harmonized analysis of MP in the future, as the advantages and disadvantages of the currently available high-throughput analysis pipelines for individual systems can be determined.

To achieve this goal, a new software tool is presented and intercalibrated with existing studies.<sup>24</sup> Its performance is evaluated via the analysis against existing reference data sets and comparison of the achieved results. Furthermore, the tool allows the analysis of data generated by different FT-IR imaging techniques from various manufacturers. This is demonstrated through the application of the software to data sets from state of the art instruments from four major manufacturers, namely Agilent, Bruker, PerkinElmer, and ThermoFisher Scientific, with different types of detection modes ranging from focal plane array (FPA), linear arrays to single-element detectors. This analysis is followed by a short performance evaluation of the assigned spectra. The software and corresponding reference spectra databases<sup>24</sup> are available free of charge via the internet and can be used by everyone for the analysis of FT-IR data.

## Materials and Methods

### *Sample Extraction and Analysis Using the Agilent System*

Sample preparation started with an effluent sample (Ryaverket wastewater treatment plant [WWTP],

Göteborg, Sweden) as collected using filtration with a custom filtration device containing a  $\varnothing 100\ \text{mm}$  stainless steel filter ( $10\ \mu\text{m}$  mesh size).

The material collected on the steel filter mesh was treated to extract MP using a method derived by Löder et al.<sup>13</sup> and modified by Liu et al.<sup>22</sup> The filter was sonicated into filtered Milli-Q water containing 5% (w/v) sodium dodecyl sulfate (SDS) to detach the solids and left stirring (100 rpm) at  $50\ ^\circ\text{C}$  for 48 h. The resulting suspension was then filtered onto a 47 mm steel filter ( $10\ \mu\text{m}$  mesh). The particles trapped were re-suspended and first incubated with proteolytic enzymes (Alcalase, Novozymes, Denmark) in a Tris buffer (pH 8.2, stirring at 100 rpm, and  $50\ ^\circ\text{C}$  for 48 h), filtered onto a 47 mm steel filter, and the solids were then removed from the filter surface. A second enzymatic treatment was performed using cellulolytic enzymes (Cellulase blend and Viscozyme, Sigma-Aldrich) in acetate buffer (pH 4.8, stirring at 100 rpm and  $50\ ^\circ\text{C}$  for 48 h) to eliminate the majority of the organic fraction of the sample matrix. The remaining undigested matter was filtered onto a 47 mm steel filter, and the solids were again removed from the filter surface. The gathered solids were oxidized using Fenton reaction (hydrogen peroxide catalyzed by Fe(II) at  $\sim 20\ ^\circ\text{C}$  for 24 h). After a further filtration on a 47 mm steel filter, the solids were removed from the filter surface and recovered. MP were further separated from the inorganic particles in a zinc chloride solution ( $1.7\ \text{g cm}^{-3}$ ) using a glass separatory funnel. After discharging the settled material, the supernatant was filtered ( $47\ \text{mm}$  steel filter) and the material was recovered following the same procedure described for the previous steps using 50% (v/v) ethanol. The extracted MP were transferred in 10 mL glass headspace vial, the liquid was evaporated at  $55\ ^\circ\text{C}$ , and finally, 5 mL 50% (v/v) ethanol solution were added to obtain a known volume in the vial.

In order to minimize MP contamination deriving from the equipment used for sampling and sample preparation, all lab tools were flushed with filtered ( $1.2\ \mu\text{m}$ ) Milli-Q water three times before use. Tools made of glass or metal or which were coated with PTFE were used instead of plastic whenever possible. Sample containers were covered with aluminum foil to reduce airborne contamination, and steel filters were muffled at  $500\ ^\circ\text{C}$  before usage.

The  $\mu\text{FT-IR}$  analysis was performed using a FPA-based  $\mu\text{FT-IR}$  imaging technique provided by a Cary 620  $\mu\text{FT-IR}$  microscope from Agilent Technologies (USA) coupled with a Cary 670 FT-IR spectroscope. The instrument was equipped with a  $128 \times 128$  FPA/mercury–cadmium–telluride (MCT) imaging detector (FPA-MCT-imaging detector). The analysis was carried out in transmission mode, using a  $15 \times$  Cassegrain (visible IR) objective-condenser system which produces  $5.5\ \mu\text{m}$  pixel resolution on the FPA detector. An aliquot of the sample ( $600\ \mu\text{L}$ ) was deposited onto a  $\varnothing 13\ \text{mm} \times 2\text{-mm}$ -thick zinc selenide (ZnSe) transmission window. A background scan was collected before each

sample scan on a clean window at  $8\text{ cm}^{-1}$  spectral resolution, using 120 co-added scans in the spectral range of  $3750\text{--}850\text{ cm}^{-1}$ . Subsequently, an area of  $14 \times 14$  tiles was scanned on the samples window, using 30 co-added scans, and the same settings as for the background scan. The analyzed area covered the entire active surface of the windows (diameter of 10 mm, area  $78.5\text{ mm}^2$ ), recording the spectra of all deposited particles.

### Sample Extraction and Analysis Using the PerkinElmer System

The sample represents a composite sample taken at 30-min intervals across a 24-h period, directly sampling from the effluent of a WWTP. The auto-sampler filtered this water through a woven stainless steel cylindrical filter cartridge (27.8 cm long, nominal pore size  $10\text{ }\mu\text{m}$ ,  $\sim 500\text{ cm}^2$  filter area; Wolftechnik, Germany). The concentrated sample was transferred from the filter into dispersion for processing in the lab. Processing in order to “clean” the sample in preparation for  $\mu\text{FT-IR}$  analysis consisted of two steps: a Fenton’s reaction to chemically degrade organic material and enzyme digestion to remove cellulosic and proteinaceous material.

The Fenton’s reaction used a Fe(II) solution (0.05 M  $\text{FeSO}_4 \cdot 7\text{ H}_2\text{O}$ , Fischer Scientific, USA,  $>98\%$  purity) acidified with 0.2% sulfuric acid ( $\text{H}_2\text{SO}_4$ , AnaTaR, 98.07% purity) and 30% hydrogen peroxide ( $\text{H}_2\text{O}_2$ , Fisher Scientific). The reaction was allowed to exhaust itself overnight before the sample was filtered and re-dispersed for enzymatic digestion. The enzyme digestion steps utilized cellulase in a pH 5, phosphate-buffered saline solution (MP Biomedicals, USA) incubated at  $50\text{ }^\circ\text{C}$  for 48 h and trypsin at  $37\text{ }^\circ\text{C}$  for 30 min (Sigma-Aldrich, Germany). The final concentrated sample was dispersed and stored in 50% ethanol (Sigma-Aldrich, Germany) before depositing onto 25 mm diameter  $5.0\text{ }\mu\text{m}$  pore size silver membrane filters (Sterlitech, USA) for  $\mu\text{FT-IR}$  analysis. All reagents were filtered through a  $1.2\text{ }\mu\text{m}$  glass-fiber filter before use, and all processing took place in the Microflow Biological Safety Cabinet, fitted with HEPA filters to control for airborne microplastic contamination during the processing of samples.

For the  $\mu\text{FT-IR}$  analysis, a PerkinElmer Spotlight 400 FT-IR microspectrometer with MCT detector was used for the analysis of the sample. The silver filter containing the processed sample was mounted on a glass slide and held in place with a clamped stainless steel O-ring. The spectrometer collected spectra in the range between  $4000$  and  $700\text{ cm}^{-1}$  in reflectance mode. A background spectrum was collected for each sample from a blank area of the silver filter at a spectral resolution of  $8\text{ cm}^{-1}$  and pixel size of  $25\text{ }\mu\text{m}$ . A total of 90 scans were taken per pixel with an interferometer speed of  $2.2\text{ cm s}^{-1}$ . An optical image was first collected by tiling single field of view images together to cover an area of approximately

$13\text{ mm} \times 13\text{ mm}$ . A smaller mapping area for the FT-IR spectrum of  $11.6\text{ mm} \times 11.6\text{ mm}$  was selected (92 % of the filter area), due to constraints on the file size that could be generated by the PerkinElmer SpectrumIMAGE software. The FT-IR mapping was performed with the same parameters as that of the background scan, but at four scans per pixel. Atmospheric correction was performed on the resulting .fsm file.

### Sample Extraction and Measurement Using the ThermoFisher Scientific System

The sample was taken by filtering water at the effluent of a WWTP over a stainless sieve ( $\varnothing 20\text{ cm}$ , ThermoFisher Scientific) having a mesh size of  $20\text{ }\mu\text{m}$ . The concentrated sample was exposed successively to SDS (one day, 5%, Serva Electrophoresis GmbH, Germany), potassium hydroxide (five days, 10%, Carl Roth GmbH, Germany), and hydrogen peroxide (two days, 32%, Carl Roth GmbH, Germany). During all steps, the sample was incubated in an oven with a temperature set at  $35\text{ }^\circ\text{C}$ . In between steps, the sample was filtered over a  $\varnothing 47\text{ mm}$  stainless steel filter with a mesh size of  $20\text{ }\mu\text{m}$ . Inorganic particles were removed by performing a density separation using a zinc chloride ( $\text{ZnCl}_2$ , Carl Roth GmbH, Germany) solution with a density of  $1.6\text{ g cm}^{-3}$ . Subsequently, sample residues were filtered on an aluminum oxide filter (Anodisc 25 mm, Whatman, UK) which was then dried at  $35\text{ }^\circ\text{C}$  for several days.

In order to minimize MP contamination, all chemicals were filtered through stainless steel filters with a mesh size of  $10\text{ }\mu\text{m}$ . Additionally, all lab equipment was thoroughly rinsed before usage, and the lab surfaces cleaned with ethanol (30%, Carl Roth GmbH and Co. KG, Germany) and Milli-Q water. Whenever possible, plastic equipment was reduced by tools made of glass or metal, and when finishing sample handling these were immediately covered with aluminum foil to reduce airborne contamination.

For the  $\mu\text{FT-IR}$  analysis, an FT-IR microscope equipped with a single MCT detector (Nicolet iN10, ThermoFisher Scientific, USA) was used. For the measurements, an Anodisc filter was placed on a calcium fluoride crystal (EdmundOptics, Germany). About one third of each filter was mapped in transmission mode, with one scan recorded per pixel, the aperture size set at  $50 \times 50\text{ }\mu\text{m}$ , the spectral resolution as  $16\text{ cm}^{-1}$ , and the spatial resolution at  $20\text{ }\mu\text{m}$ . A background scan using the same settings was conducted on a blank area of the same filter.

### Sample Extraction and Measurement Using the Bruker System

For this comparison study, a data set of a sample investigated in previous studies<sup>25,26</sup> was chosen. The sample was from the effluent of the WWTP Holdorf. Sample location

and further sampling information are available within the previous studies.<sup>25,26</sup> Briefly summarized, the sample was directly taken from the effluent of the WWTP. The sample was processed by the enzymatic digestion protocol of Löder et al.<sup>13</sup> and subsequently concentrated onto an Anodisc filter (25 mm diameter, 0.2  $\mu\text{m}$  pore size, GE Whatman). The sample was placed and covered by a BaF<sub>2</sub> window prior to measurement on a custom-made sample holder as described in detail in Primpke et al.<sup>25</sup>

The  $\mu\text{FT-IR}$  measurement was performed using a Bruker TENSOR II spectrometer, which is connected to a Hyperion 3000  $\mu\text{FT-IR}$ -microscope (Bruker Optics GmbH, Ettlingen, Germany). The spectra were collected using a  $64 \times 64$  FPA MCT detector as described in literature.<sup>19</sup> Prior to measurement, a visual image of the sample surface was recorded. The FT-IR measurements were performed using  $15 \times$  Cassegrain objective, in the spectral range of  $3600\text{--}1250\text{ cm}^{-1}$  with  $4 \times 4$  binning at a spectral resolution of  $8\text{ cm}^{-1}$  and six coadded scans. With this setup, a pixel size of  $11.1 \times 11.1\ \mu\text{m}$  per spectra was achieved. All data were collected using Bruker software OPUS 7.5 (Bruker Optics GmbH, Germany).

### Sample Preparation of Algae Samples

Ecotoxicity tests were carried out with the microalga *Raphidocelis subcapitata* with modifications according to “Water Quality–Freshwater Algal Growth Inhibition Test with Unicellular Green Algae, ISO Standard 8692, 2004”. The test included five replicates of the control sample, in which algae were not exposed to the toxicant, and triplicates of algae exposed to five different concentrations of the tested toxicant. The replicates were combined after the toxicity test, and algae from all treatments were preserved with Lugol’s iodine solution for their infrared analysis.

For the infrared analysis described in Kansiz et al.,<sup>27</sup> 2 mL of the preserved control sample was prepared. The cells were centrifuged, and the residues of the preservative and growth media were washed off to prevent interference with algae spectra. The purified cells were resuspended in 200  $\mu\text{L}$  deionized water, and the entire volume was deposited on a 13 mm diameter, 1-mm-thick CaF<sub>2</sub> transmission window, and dried in a vacuum desiccator.

A Cary 620  $\mu\text{FT-IR}$  microscope from Agilent Technologies (USA) coupled with a Cary 670 FT-IR spectroscope was used for the FPA-based  $\mu\text{FT-IR}$  imaging analysis of the algae cells. The analysis was carried out in transmission mode using the  $15 \times$  Cassegrain objective in high magnification mode to create a mosaic with  $1.1\ \mu\text{m}$  pixel resolution on the FPA. A background scan was collected on a clean window in the range of  $3750\text{--}850\text{ cm}^{-1}$  with an  $8\text{ cm}^{-1}$  spectral resolution applying 256 co-added scans. An area of  $6 \times 6$  tiles was scanned on the sample window following the background scan with the same parameters applying 240 co-added scans per pixel.

### Software and Database

The software siMPle is the combination of the software MPhunter presented in Liu et al.<sup>22</sup> and the automated analysis of Primpke et al.<sup>21</sup> It is written in Delphi using Windows 10 as the operating system, which is available at [www.simple-plastics.eu](http://www.simple-plastics.eu), where reference databases for FT-IR and Raman spectroscopy are also provided (Fig. 1).

In this work, the database for automated analysis<sup>24</sup> and the release version (1.0.0) of siMPle was used to analyze all data sets. Prior to analysis, all spectra were converted from the original file format (Agilent:.dmd and PerkinElmer:.fsm) or JCAMP-DX files (Bruker and Thermo Fisher Scientific) into two siMPle file formats, namely .spe and .wno files, allowing fast data access (see How to Use siMPle.pdf Supplemental Material file for further instructions). These file formats are accompanied by an extra file, the MaschineData.ini, which contains all necessary information for size calculations and data handling. In all cases, the mosaic structure is either kept or newly generated, allowing faster data handling. After loading the data and reference spectra, the spectral fit between the two was calculated by Pearson correlation for the untreated data, the first derivative and the second derivative, resulting in their correlation factors  $r_0$ ,  $r_1$ ,  $r_2$ , respectively. If not further specified, the following settings were used: omit CO<sub>2</sub> peak (upper wavenumber:  $2420\text{ cm}^{-1}$ , lower wavenumber:  $2200\text{ cm}^{-1}$ ), suppress negative correlations, and include second-order derivatives.

To investigate the performance between Bruker OPUS and siMPle, the calculation times were measured using a HP KP719AV computer (Intel Core 2 Duo Processor, 8 GB RAM, AMD Radeon HD 5450 graphic card, extra USB 3.0 Controller card and SANDISK Extreme 64 GB USB-Stick) which is the same as used in previous studies.<sup>21,24–26,28–34</sup> Further calculations were performed on a HP Z440 workstation (Intel Xeon E5-2630 v.3 CPU, 64 GB RAM, NVidia Quadro M2000 graphics card) for all other purposes.

The automated analysis pipeline (AAP)<sup>21</sup> identifies the recorded spectra based on the results of the Pearson correlation factors ( $r$ ) calculated for the respective untreated spectra  $r_0$  and the first derivatives  $r_1$ . Only if maximum values of  $r_0$  and  $r_1$  are assigned to the same polymer entry, then the spectra are counted as identified and the polymer type added to the list of analyzed pixels together with the  $x,y$  coordinates and the summarized hit quality index (HQI, Eq. 1):

$$HQI = (r_1 + r_2) \times 1000 \quad (1)$$

This type of data represents a false color image which was then in silico treated by Image Analysis as described in Primpke et al.<sup>21</sup> by a pixel hole closing mechanism prior to the size determination and particle quantification. For further calculations, the data thresholds described in Lorenz et al.<sup>32</sup> were used. To avoid confusion for the reader, we did



**Figure 1.** The software graphical user interface with a loaded reference (blue) and a sample spectrum (orange) using the Match single spectrum function of siMPle on a spectrum assigned as polyvinylchloride using ESM1.csv.

not apply the second analysis pipeline<sup>1,22,35</sup> available in siMPle for the scope of this study.

The siMPle software allows rapid QA/QC for the assigned polymer hits. During image analysis, a designated file is generated named “\_forc.csv,” which contains x,y coordinates of the measured spectra together with the hit quality, the assigned polymer type, and a reference spectrum identifier. Via the button “Load Pipeline Results,” the QA/QC process will be started (Fig. 2).

Using a designated window, each spectrum can be individually assessed and rated from a perfect assignment down to a full misassignment with values ranging from 1 to 0.01. Following the approach of Primpke et al.,<sup>21</sup> the spectra were rated with five different values (1.00, 0.75, 0.5, 0.25, and 0.01) ranging from best match to a full misassignment, respectively. To screen the data of the different instruments and evaluate the successful assignment of polymer identify to particles, a randomly selected number ( $n = 100$ ) of spectra were manually reanalyzed for each instrument data set.

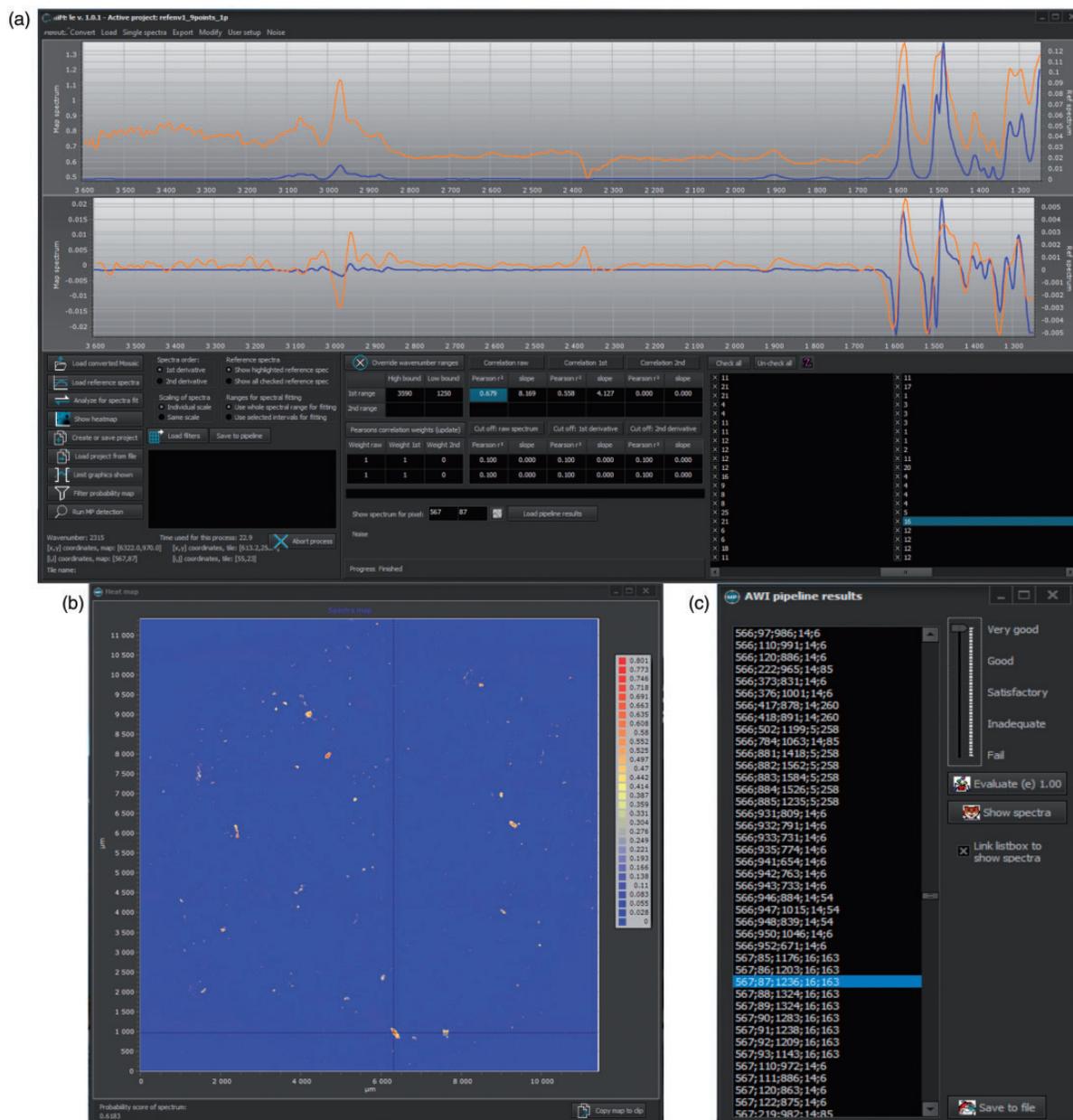
For matching single spectra, the score used within siMPle was first described in Liu et al.<sup>22</sup> In short, prior to the quantification of particles, the score ( $S_d$ ) for the identification of the polymer type was calculated using Eq. 2:<sup>22</sup>

$$S_d = \frac{k_0 r_0^2 + k_1 r_1^2 + k_2 r_2^2}{k_0 + k_1 + k_2} \quad (2)$$

Each result ( $r_0, r_1, r_2$ ) was squared and multiplied by a weight ( $k_0, k_1, k_2$ , respectively) for the respective correlation factor, which can be assigned by the user.

## Results and Discussion

The software allows two types of data analyses: first, the matching of single spectra and second, the analysis of large filter areas. To interpret a single spectrum, the reference database must first be loaded and then a single spectrum in a defined file format (Paragraph S1; ESM1.csv, Supplemental Material) must be loaded (Fig. 1). In this example, the spectrum was assigned to polyvinylchloride (PVC) with a  $S_d$  of 0.7246 using the default options of siMPle ( $k_0 = 0$ ,  $k_1 = 1$ ,  $k_2 = 1$ ). By assigning weights, the user can decide which correlation method should be represented by  $S_d$ , e.g., for comparison with studies using Bruker OPUS<sup>28,30–33</sup> where only the first derivative was used for single spectrum analysis. For the chosen example, values of  $k_0 = k_2 = 0$ ,  $k_1 = 1$  (first derivative only) were used. In this case, the  $S_d$  decreased to 0.6689. If the user decides to include all correlation results ( $k_0 = k_1 = k_2 = 1$ ), the  $S_d$  was further decreased to 0.6435. Therefore, it is mandatory to state the weights  $k_0$ ,  $k_1$ ,  $k_2$  within the material and methods section for comparison of studies if siMPle is used. In general, this analysis is independent from the instrumental source of



**Figure 2.** Quality assurance/quality control for siMPLE for results of the image analysis of sample RefEnvl<sup>24</sup> via the AWI pipeline allowing a direct comparison (a) between sample spectrum (orange) and database spectrum (blue) by clicking on the determined pixel (c). The process allows checking other spectra by clicking on another database entry as well as the individual Pearson correlation factors. The heatmap (b) allows the user to locate the assigned spectra on the map.

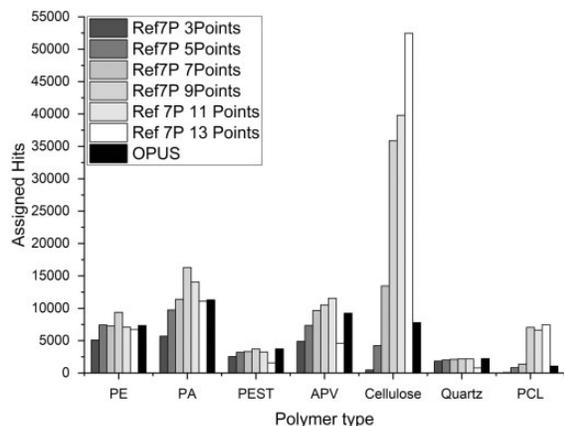
the data. All types of data in the described file format (see Paragraph S1) can be processed independently from manufacturer and measurement method. An example of this using preprocessed Raman spectra is shown in Fig. S1. Future data processing steps for single spectra or lists of spectra will be introduced in future releases. All releases will be accompanied via a change log and a living manual on the website.

During this process, QA/QC is easily available, because the spectrum with the highest hit is indicated at the end of

the analysis together with a hit list for all other database entries (Fig. 1). Together with this, the hit result can be exported for further use.

### Analysis of Chemical Imaging Files

For larger data sets, the siMPLE software allows a time-efficient loading of data using a harmonized file system for data storage which is independent from the original file format. The file formats introduced were optimized for



**Figure 3.** Assigned polymer types for sample Ref7P<sup>24</sup> using different smoothing factors by siMPle (3 to 13) versus OPUS with an unknown fixed value for polyethylene (PE), polyamide (PA), polyester (PEST), acrylates/polyurethanes/varnish (APV), cellulose, quartz and polycaprolactone (PCL).

the fast use within the software. The commonly shared file format by the International Union of Pure and Applied Chemistry (IUPAC) J-CAMPdx file format was limited to the import of data, due to long loading times of such text type based files. Currently, siMPle is able to convert native data from Agilent and PerkinElmer systems using the file import function while for Bruker and ThermoFisher Scientific systems extra steps are necessary (see How to Use siMPle.pdf, Supplemental Material).

To validate the performance of siMPle, it was tested against existing reference data sets from literature. These reference sets consist of materials of known origin (Ref7P) or from environmental samples (RefEnv1 and RefEnv2). These three samples were analyzed using siMPle and the results were compared to the automated analysis via Bruker OPUS. Starting with Ref7P, we started to analyze the performance on an artificial sample only containing MP (polyethylene (PE), polyester (PEST), polyamide (PA) and polyurethane (PU)), cellulose, quartz, and diatomaceous earth. Out of these seven particle types, only diatomaceous earth could not be detected within the wavenumber range of Anodisc. In comparison to OPUS, siMPle identified almost twice the number of spectra on the specific polymer types (Fig. 3). Especially, cellulose (plant fiber in the database) and polycaprolactone (PCL, not originally introduced as a material) were affected by factors larger than four (Fig. 3, Ref7P 9Points and OPUS).

This difference was rather striking and the main difference between both kinds of software was found in the data handling for the calculation of the first derivative. In the default settings, siMPle adds a nine data points smoothing to reduce the noise. For Bruker OPUS, it is not documented if smoothing is applied. To test for a better comparison of the results, a range of this value from 3 to 13 data points was investigated for siMPle (Fig. 3).

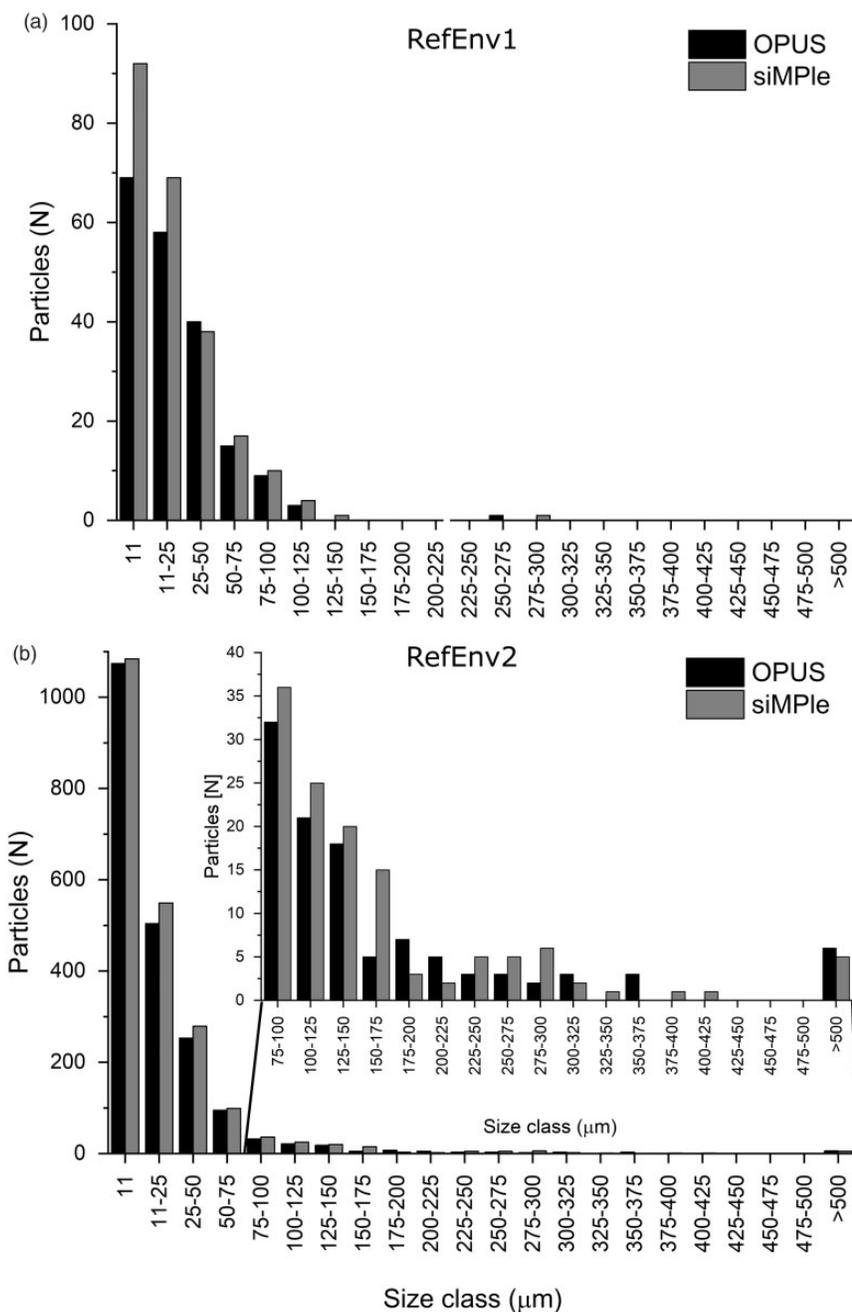
By screening the number of data points for smoothing, it was found that an optimal hit was reached with nine data points for most polymer type assignment (Fig. 3). Only cellulose kept an increase in assigned polymer hits while PCL reached a constant level. The data determined by OPUS could not be assigned to a smoothing factor applied by siMPle. Still, for this particular sample, the high number of assignments to PCL started using this number of data points for smoothing. To avoid any misassignment issues, a manual reanalysis on the assigned spectra to PCL was performed.

Through quality assurance, it was found that the spectra assigned to PCL were caused by a misinterpretation of the measured PEST spectrum. This spectrum has a high similarity with PCL in transmission, which was not visible by using Bruker OPUS during cluster performance analysis.<sup>24</sup> In Fig. S2, one of the assigned spectra from Ref7P is plotted against the assigned reference spectrum and the spectra of the original material. The original database states this material as a pure PEST, but via an extended material research, it was found that the material was meanwhile relabeled to copolyester by the manufacturer. Due to these differences, the material could not be assigned to the original PEST cluster, as no pure PEST spectrum was yielded. This issue will be addressed in the future by a database update including more materials and using siMPle for cluster performance analysis. All samples in the following were analyzed using the default nine data points smoothing.

The data sets RefEnv1 and RefEnv2 were also analyzed using siMPle and OPUS. The siMPle analysis required only 2 h for RefEnv1 and 3 h for RefEnv2, which is 12 times faster than the analysis with OPUS using spectral correlation only for raw and first derivative data. With a look at the polymer composition (see Table S1 for details), it was found that siMPle was more sensitive and identified higher numbers of polymers and also larger sized polymer particles (Fig. 4) in comparison to OPUS.

Both analyses found a strong trend toward smaller MP sizes. Especially striking was the higher identification rate for cellulose (plant fiber) during the analysis with siMPle (see RefEnv2, Table S1). Sample RefEnv2 also showed the largest differences in the size distribution and showed a better particle assignment compared to the data derived via OPUS (Fig. S3).

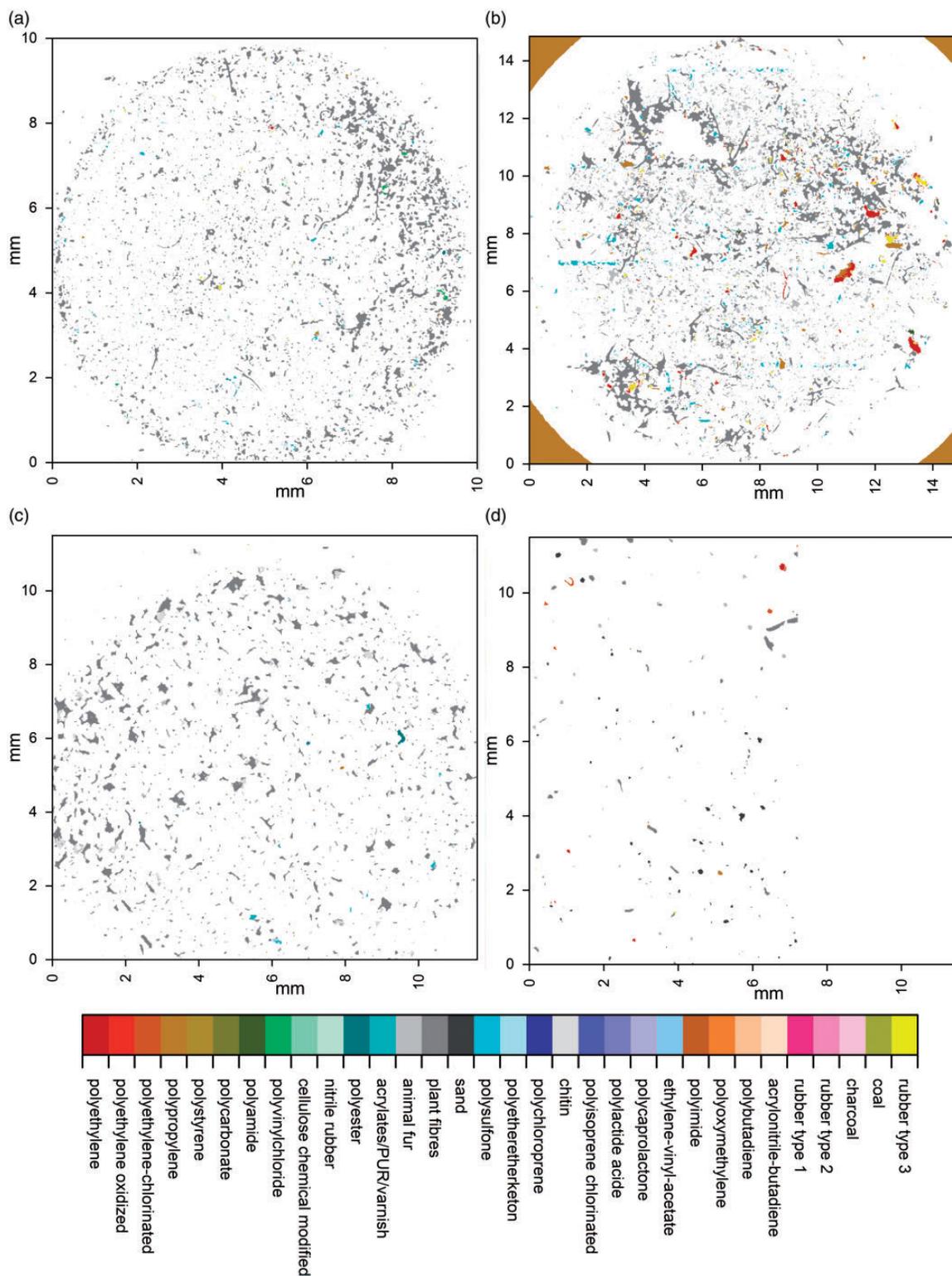
Here, it was found that larger particles were identified more accurately with siMPle in comparison to OPUS, which missed areas of larger particles yielding in a separation into two particles. Furthermore, the analysis using siMPle improved closing holes, which is important for morphological analysis of the particles, see for example, the large PP particle on the rightmost edge of the filter. In summary, these results show that data determination with siMPle is better suited for the analysis of imaging data due to transparent data handling and easy data validation.



**Figure 4.** MP size classes derived from the automated image analysis<sup>21</sup> for the reference data sets (a) RefEnv1<sup>24</sup> and (b) RefEnv2<sup>24</sup> analyzed using Bruker OPUS and siMPle.

**Table I.** Calculation times of the different data sets measured on systems of four different manufacturers using siMPle.

Data set	Size .spe GB	Spectra N	Pixel size $\mu\text{m}$	Calculation time s	Calculation performance spectra s	Filter area measured mm
Agilent	9.01	3 211 264	5.5	29 979	107	10 × 10
Bruker	4.14	1 806 336	11.05	16 464	110	14.9 × 14.9
PerkinElmer	0.66	215 296	25	2877	75	11.6 × 11.6
ThermoFisher Scientific	0.25	221 184	20	1129	195	11.5 × 7.2



**Figure 5.** Overview images of the measured filters using the automated analysis pipeline for the (a) Agilent, (b) Bruker, (c) PerkinElmer, and (d) Thermo Fisher Scientific samples. Sample (d) was measured in a rectangular shape and the area of on the right side was left blank to avoid irritations.

To test the ability for a harmonized analysis of MP, the performance of siMPle was assessed for the FT-IR imaging data from instruments from the four mentioned manufacturers. In this case, the computation of a full spectral analysis based on Pearson correlation for the untreated data, the first derivative and the second derivative were performed.<sup>22</sup> The calculation time was determined on the same computer systems applied for the automated analysis via the OPUS software of the Bruker data set. This allows a comparison of calculation times with existing studies using OPUS.<sup>26,28</sup> The determined calculation times are summarized in Table 1 together with further information on the data sets, such as pixel size on the filter area, the size of the analyzed area and the number of spectra recorded.

As mentioned previously, the calculation time on the Bruker data sets was reduced considerably when applying siMPle (5 h, in comparison to 48 h using OPUS) which also included the second derivative (it was omitted during OPUS analysis). When the calculation time was normalized to the number of analyzed spectra, the spectra from Thermo Fisher Scientific were correlated twice as fast compared to the other data sets. The reason is the spectral resolution of  $16\text{ cm}^{-1}$  instead of  $8\text{ cm}^{-1}$ . Still, one has to keep in mind that the Bruker system and the Thermo Fisher Scientific system need an additional transformation step within their respective software which increases the overall handling and calculation times independent from siMPle. In Fig. 5, the false color images of the analyzed samples are shown.

Qualitatively, it can be observed from the images that the samples are similar in nature, containing a large proportion of natural materials (Fig. 6, gray colors), among which a number of artificial polymers are successfully identified, irrespective of the manufacturer of the instrument or the various sampling and extraction methods employed prior to analysis (see ESM2.xlsx, Supplemental Material, for

details). Due to the varying nature of the WWTP sampled from, and the variety of sampling and extraction methods utilized between samples, commentary on any differences in enumeration of MP between the samples is beyond the scope of this study. However, application of the software on these real-world example data sets demonstrates promising consistency in the proportion of particles identified which are of synthetic origin (3–25%) and of the major polymer types which are identified across the samples (see ESM2.xlsx). Compared to existing commercial software solutions, the harmonized analysis via siMPle saves working time and computational costs. Current computers can run several instances of the software unattended, allowing the data analysis of up to 16 samples per day compared to OPUS (two days) for 1.8 million spectra per file. Still, it is possible also to use low-cost office computers which normally can calculate up to three samples per day containing 1.8 million spectra. As a minimum requirement, a processor speed of 3 GHz is advised with 8 GB of RAM. To assess the performance of siMPle on these data sets, a QA/QC analysis on the overall result was performed (Fig. 6).

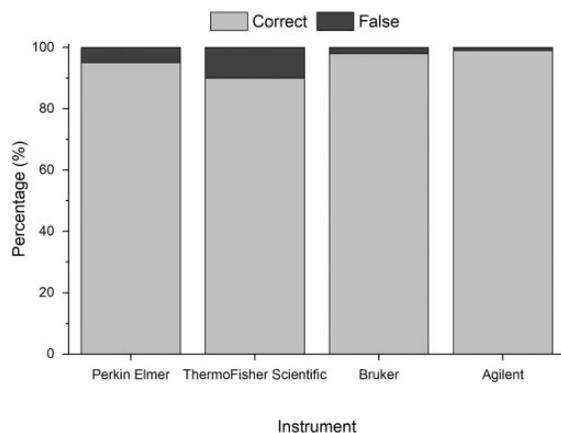
In all cases, correct assignment rates  $>90\%$  were reached, for three systems (PerkinElmer, Bruker, and Agilent), and these were even  $>95\%$  (Fig. 6). Those correct assignment rates were exceedingly high, independently from instrument and sample preparation, proving the high potential of siMPle as a harmonized tool for MP analysis. Still, it is suggested and recommended that each study conducts an own QA/QC analysis for each sample series for each polymer type identified as demonstrated, e.g., in Lorenz et al.<sup>32</sup> Further questions, such as a comparison between existing analytical pipelines, their harmonization, and a full QA/QC analysis will be addressed in a later detailed study.

To conclude, it is noteworthy that the siMPle software is not limited to MP analysis, and it also allows the analysis of other types of data like the spectral comparison of nano-FT-IR data<sup>36</sup> or the analysis of single algae species (Fig. S4).

Using siMPle, single cells can be selected or specific characteristics can be highlighted (Fig. S4). Here, the data show a strong Halo effect (Fig. S4b) mainly caused by interference between the sample and the surface of the  $\text{CaF}_2$  window, which is not visible using a heatmap (Fig. S4a). In the future, heatmaps based on the integration of specific regions will be introduced to allow even more control over the data. Further, additional functions are currently planned to be introduced, and new possibilities can be determined by contacting the authors to explore its application in a broader scope for future research.

## Conclusion

With siMPle, we present a freeware data analysis tool for the harmonized and systematic analysis of spectroscopic data, with application, for example, in the identification of



**Figure 6.** Assignment rates of correct and misidentified spectra for the different instruments based on manual reanalysis similar to Primpke et al.<sup>21</sup>

MP in the environment. It allows data determination and interpretation in a transparent and reproducible manner. In addition, it provides a simpler quality QA/QC compared to existing commercial software tools like Bruker OPUS and shows an increased identification rate. Furthermore, it allows the analysis independently from the instrument manufacturer for a single spectrum but also for large fields generated by imaging techniques. In particular, the field of FT-IR imaging benefits greatly, as the data calculation time is reduced from several days to 5 h using this software tool. Compared to other techniques, all spectra are correlated via three different data treatments with the database yielding high-quality results for all investigated instrument systems. This new tool improves the application of FT-IR imaging in monitoring studies for MP, as it is accessible for most types of spectrometers, free of charge and reduces the human bias during manual data analysis.

### Declaration of Conflicting Interests

The author(s) declared no potential conflicts of interest with respect to the research, authorship, and/or publication of this article.

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### Supplemental Material

All supplemental material mentioned in the text is available in the online version of the journal.

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