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Towards harmonizing natural resources as an area of protection in Life Cycle Impact Assessment

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Abstract

Purpose

In this paper, we summarize the discussion and present the findings of an expert group effort under the umbrella of the United Nations Environment Programme (UNEP)/Society of Environmental Toxicology and Chemistry (SETAC) Life Cycle Initiative proposing natural resources as an Area of Protection (AoP) in Life Cycle Impact Assessment (LCIA).

Methods

As a first step, natural resources have been defined for the LCA context with reference to the overall UNEP/SETAC Life Cycle Impact Assessment (LCIA) framework. Second, existing LCIA methods have been reviewed and discussed. The reviewed methods have been evaluated according to the considered type of natural resources and their underlying principles followed (use-to-availability ratios, backup technology approaches, or thermodynamic accounting methods).

Results and discussion

There is currently no single LCIA method available that addresses impacts for all natural resource categories, nor do existing methods and models addressing different natural resource categories do so in a consistent way across categories. Exceptions are exergy and solar energy-related methods, which cover the widest range of resource categories. However, these methods do not link exergy consumption to changes in availability or provisioning capacity of a specific natural resource (e.g. mineral, water, land etc.). So far, there is no agreement in the scientific community on the most relevant type of future resource indicators (depletion, increased energy use or cost due to resource extraction, etc.). To address this challenge, a framework based on the concept of stock/fund/flow resources is proposed to identify, across natural resource categories, whether depletion/dissipation (of stocks and funds) or competition (for flows) is the main relevant aspect.

Conclusions

An LCIA method - or a set of methods - that consistently address all natural resource categories is needed in order to avoid burden shifting from the impact associated with one resource to the impact associated with another resource. This paper is an important basis for a step forward in the direction of consistently integrating the various natural resources as an Area of Protection into LCA.

1. Introduction

Life Cycle Assessment (LCA) is the compilation of inputs (consumption of resources) and outputs (emissions) and the evaluation of related potential environmental impacts of a product system throughout its life cycle (ISO 2006). Other types of LCA exist, e.g. social LCA, but in this paper, the term LCA refers to environmental LCA. According to the new Life Cycle Impact Assessment (LCIA) framework (Frischknecht and Jolliet 2016), environmental impacts can be expressed on the level of individual impact categories or can be aggregated into so-called damage categories, or Areas of Protection (AoP), including ‘Human Health’, ‘Ecosystem Quality’ (sometimes referred to as ‘Natural Environment’) and ‘Natural Resources’ (see also EC-JRC 2010; Hauschild and Huijbregts 2015). While the former two are well-established and accepted, the role of the latter in LCA is still debated and there is no consensus on how this AoP should be tackled methodologically (see e.g. EC-JRC 2010; Mancini et al. 2013; Dewulf et al. 2015a). However, the natural environment provides natural resources, i.e. the substances/materials and flows that humans can use (e.g. metals, water, or wind), and changes on these provisions can therefore be considered an environmental impact.

Natural resources play a role in two phases of LCA: as elementary flows in the inventory analysis and as an AoP in LCIA. The focus of this paper is on LCIA methods and the AoP ‘Natural Resources’ (see Table S1 for naming in different methods). Natural resource consumption inventory flows (e.g. consumption of minerals, fossil fuels, land, or water) may have an impact on the AoP ‘Natural Resources’, but also on the other AoPs ‘Ecosystem Quality’ and ‘Human Health’. For instance, land use may impact biodiversity (Koellner et al. 2013) and water consumption may cause shortages for irrigation, resulting in human malnutrition (Pfister et al. 2009). This paper does not address such resulting impacts on the AoP ‘Ecosystem Quality’ and ‘Human Health’. Furthermore, emission inventory flows may have an impact on the AoP ‘Natural Resources’, e.g. emissions to water may decrease freshwater quality and thereby its availability at a specific quality level (Boulay et al. 2011; Bayart et al. 2014). However, these qualitative assessments are a combined assessment of pollution effects causing impacts on humans and ecosystems as well as impacts on resource availability that are not commonly established in LCIA methods.

Existing LCIA methods mainly consider the intrinsic values of human health and ecosystem quality, i.e. their “value by virtue of their pure existence”, and the instrumental value of natural resources, i.e. their “utility to humans” (Frischknecht and Jolliet 2016). However, there is little agreement in the scientific community on what exactly is to be protected under the AoP ‘Natural Resources’ and what kind of metric should be used. Within the

UNEP-SETAC Life Cycle Initiative, it has been argued that the damage to natural resources consists of “the reduced availability of the corresponding type of resource to future generations” (Jolliet et al. 2004). Several approaches have been proposed to account for this, e.g. depletion rates (use-to-stock and use-to-availability ratios) or increased efforts for future generations to access resources in lower quality deposits. On the other hand, some authors claim that short- and medium-term (from a few years to a few decades) *availability* of mineral resources is mainly constrained by socio-economic factors and it is therefore debatable whether natural resource *availability* should be addressed in an environmental assessment (Drielsma et al. 2016). However, changes in the environment’s capacity to provide natural resources is clearly an environmental issue, which should be of concern in an AoP ‘Natural Resources’.

Although LCIA methods traditionally focused on abiotic natural resource depletion (minerals/metals and fossil fuels) (Weidema et al. 2007), there is no generally accepted impact assessment method (or model) for these natural resource categories and several methods exist concurrently (van der Voet 2013 in Mancini et al. 2013). Methods for other resource categories such as water and soil exist in parallel. In general, no method addressing impacts on natural resources, neither at midpoint nor at endpoint, can be recommended without restrictions (EC-JRC 2011; Hauschild et al. 2013). This paper reviews existing LCIA methods/models addressing natural resources and discusses their conceptual approaches across different natural resource categories. This is an important basis for further method development and moving towards a more consistent assessment within the AoP ‘Natural Resources’. This paper is an output of a working group within the task force on crosscutting issues mandated by the UNEP-SETAC Life Cycle Initiative as a part of its flagship activities. It is structured as follows: first, natural resources are defined and categorized for the LCA context; second, existing methods that assess impacts on natural resources are briefly reviewed by resource category; and third, existing approaches are analyzed and discussed across resource categories.

2. Definition and categorization of natural resources

Definition of natural resources

From the discussions of the working group, it was concluded that natural resources are of concern in LCA because of their instrumental value to humans. This focus on the instrumental value is consistent with the definition of the new overall LCIA framework of the UNEP-SETAC Life Cycle Initiative (Frischknecht and Jolliet 2016). The working group acknowledges the complexity of defining natural resources and the existence

of different definitions (see e.g. WTO 2010; Fischer-Kowalski and Swilling 2011; Dewulf et al. 2015b). The majority of the group agreed on the following definition of natural resources in LCA, which is compatible with the UNEP-SETAC LCIA framework:

Natural resources are material and non-material assets occurring in nature that are at some point in time deemed useful for humans.

Natural resources include minerals and metals, air components, fossil fuels, renewable energy sources, water, land and water surface, soil, and biotic natural resources such as wild flora and fauna. Natural resources may be distinguished from (primary) raw materials and (primary) energy carriers, which are the result of transformation of natural resources by the primary production sector through operations such as growing, harvesting, mining, and refining (Dewulf et al. 2015b). The World Trade Organization (WTO), for example, does not make this distinction since most resources require some processing before they can be traded or consumed (WTO 2010). However, the WTO also states that “the line of demarcation between natural resources and other goods will always be somewhat arbitrary” (WTO 2010). The WTO distinguishes natural resources from manufactured products (subject to a *substantial* amount of processing) and agricultural goods (cultivated rather than extracted from the natural environment). Also in the LCA context, biotic resources produced by an industrial production process (such as agricultural crops, livestock, fish from aquaculture, or wood from a plantation) are usually not classified as biotic natural resources (Klinglmair et al. 2014). They are produced with natural resource inputs, such as soil and water, and are considered part of the technosphere. Natural biotic resources (and water, surface, and soil) are natural resources and eco-system components (contributing to ecosystem quality) at the same time. Hence, natural biotic resource (or water, surface, or soil) use may have impacts on various AoP, which must be acknowledged by focusing on the issue in question. For instance, fishing would have an impact on the AoP ‘Natural Resources’ when less fish is available as a food source (overfishing), but it could also impact biodiversity (species richness, composition and/or abundance), which would be assessed in the AoP ‘Ecosystem Quality’. Such parallel impacts in various AoPs as a consequence of the same environmental intervention are not new in LCA. For example, a toxic emission may have an impact on aquatic organisms (impacts on AoP ‘Ecosystem Quality’) and also enter the human food chain, e.g. by fish consumption (impacts on AoP ‘Human Health’). The term ‘natural’ indicates that the resource is occurring in nature, untransformed by humans. Anthropogenic deposits such as landfills can also be considered sources for secondary resources or raw materials. However, they are neither addressed as inventory flows nor in LCIA. The resource properties do not necessarily get lost when entering the technosphere, but they may be “occupied or “borrowed” by a user within

the product system. If it can be recycled afterwards, additional extraction of natural resources can be avoided. Natural resources can provide space (e.g. land area), substances and materials, or sources of energy. While some definitions of natural resources only consider these source functions, others also include sink functions (Dewulf et al. 2015b), i.e. the absorption of emissions in soil, water, and air. In existing LCIA methods, emissions to environmental compartments are considered and the corresponding impacts on humans and ecosystems are covered by the AoPs 'Human Health' and Ecosystem Quality'.

Categorization of natural resources

Natural resources are often categorized as stock, fund, or flow resources (see e.g. Udo de Haes et al. 2002; Klinglmair et al. 2014) according to their renewability and exhaustibility (Table 1).

Stock resources are considered to exist as a finite amount and are assumed to be non-renewable (they form and concentrate extremely slowly), and are therefore regarded as exhaustible (i.e. they can be used up). Examples are fossil and mineral resource stocks. Whilst individual chemical elements do not disappear and are not exhaustible, in a strict sense, they can be subject to dissipation such that deposits with some minimum level of concentration (useful to humans) may be finite and therefore exhaustible (Dewulf et al. 2015a). In this sense, the problem with the resource consumption is still a stock resource problem, i.e. a depletion or a dissipation problem.

Fund resources are renewable, i.e. they are continually supplied or re-concentrated once dissipated, but (at least in some cases) also exhaustible if overused (Udo de Haes et al. 2002). The available amount of a fund resource can either be decreased or increased, depending on the ratio of extraction to the renewal rate. Typical examples are fish or wild animals, but the depletion of water bodies such as the Aral Sea can also be considered a fund resource problem.

Flow resources are non-exhaustible and have a limited availability at a certain time (Udo de Haes et al. 2002), which means that they have to be used as, when, and where they occur. They can be considered renewable when they re-occur at the same location. Examples are solar radiation or run-off from rivers.

How to define the boundaries between stocks, funds, and flows, in particular based on regeneration rates, is still an open question.

Special cases are land and water surface areas, which are permanently present and usually constant in the total available amount. They cannot be depleted or dissipated but only occupied and as such are non-exhaustible. This does not fit well into the stock/fund/flow classification and has sometimes been kept a separate category besides abiotic and biotic natural resources that have been categorized into stocks/funds/flows (see e.g. Heijungs

et al. 1997; Lindeijer et al. 2002). Nonetheless, competition for area has been considered to be a flow resource problem because surface area (just quantity, disregarding quality) cannot be depleted and hence is not lost for future generations (Lindeijer et al. 2002). The issue with land quality or soil properties may be considered to be a fund resource problem because soil properties can be deteriorated (or remediated) such that soil loses (or increases) its usefulness for a certain purpose.

<Table 1>

According to the definitions above, only depletion or dissipation of stock and fund resources imply a damage to the resource as such in its available form. Although there is no agreement on how this damage should be assessed, existing methods mainly relate it to potential consequences for future generations (e.g. reduced availability due to depletion or increased efforts for resource extraction). The use of a flow resource may have impacts on its temporary availability and therefore the impact is the consequences of the increased competition for this resource, rather than any lasting impact on the resource itself.

Most existing LCIA methods focus on mineral/metal and fossil fuel natural resources (see Table S1). Water (substance) and land (surface) are generally assessed separately (Klinglmair et al. 2014). Soil can also be assessed as a resource (see e.g. Milà i Canals et al. 2007a; Koellner et al. 2013; Vidal Legaz et al. 2016), and should not be confused with land (surface) use impacts on biodiversity. Table 2 shows a compilation and categorization of natural resources based on Klinglmair et al. (2014), Dewulf et al. (2015a), Goedkoop et al. (2013), and Frischknecht and Büsser Knöpfel (2013). It is specified whether the natural resource consumption potentially causes a stock, fund, or flow resource problem as listed in Table 1. Furthermore, corresponding elementary flows/activities in the *Ecoinvent 2.2* and *3.2* databases (Frischknecht et al. 2007; Ecoinvent 2015) have been added to demonstrate that natural resources in the impact assessment match the resources in the inventory.

<Table 2>

3. Which resources are addressed in current LCIA methods and how?

Most existing methods are restricted to the dissipation or depletion of mineral/metal and fossil fuel natural resources (see Table S1). Exceptions are the differently organized LIME/LIME 2 method (Itsubo et al. 2004; Itsubo and Inaba 2012) and the Stepwise 2006 method (Weidema et al. 2007), which labels other resources as “Human” and “Biotic”. The operational methods covering the widest range of resource categories are thermodynamic accounting methods (CED, CExD, CEENE, SED; see Table 3). The conceptual framework covering the widest range of resource categories is provided by Stewart and Weidema (2005). It focuses on the functionality of resources and relies on two parameters: the ultimate quality limit and the backup technology (Stewart and Weidema 2005). For water and land use, resource specific frameworks were developed within the UNEP-SETAC Life Cycle Initiative (Milà i Canals et al. 2007a; Bayart et al. 2010; Koellner et al. 2013). Since frameworks and methods have been developed for different resource categories, further analysis of existing methods is also structured along five natural resource categories: (1) minerals/metals and fossil fuels (often referred to as abiotic natural resources), (2) water, (3) land and water surface (4) soil, and (5) biotic natural resources. Air components and renewable energy sources (see Table 2) are only covered in exergy and solar energy methods.

<Table 3>

3.1. Minerals/Metals and Fossil Fuels

A wide range of methods is available for the abiotic natural resource categories minerals/metals and fossil fuels. These methods (and their underlying models and indicators) have been distinguished into four different types in literature (see e.g. Stewart and Weidema 2005; Steen 2006; Rørbech et al. 2014; Swart et al. 2015):

1. methods aggregating natural resource consumption based on mass or energy
2. methods relating natural resource consumption to natural resource stocks or availability

3. methods relating current natural resource consumption to consequences of future extraction of natural resources (e.g. potential increased energy use or costs)
4. methods quantifying consumption of exergy or solar energy

Method types 1 and 4 can be grouped together as “Resource Accounting Methods” (RAM) (Swart et al. 2015). The fact that RAM do not explicitly link used amounts of resources to changes in their availability or provisioning capacity is perceived by many as a drawback. Type 1 methods are not further discussed here. However, the type 1 indicator Cumulative Energy Demand (CED) can serve as a screening indicator for environmental performance (Huijbregts et al. 2010) and is widely applied in practice. Moreover, type 1 indicators, such as Material Input per Service-Unit (MIPS), are widely used to calculate material footprints (Saurat and Ritthoff 2013). Type 4 methods are more comprehensive than CED due to the assessment of the quality of energy and the inclusion of non-energetic resources (Bösch et al. 2007). In this paper, they are referred to as “thermodynamic accounting methods”.

Type 2 methods are based on use-to-availability ratios. However, there are different estimates for resource availability and the terminology differs between different organizations (e.g. the US Geological Service (USGS) and the Committee for Mineral Reserves International Reporting Standards (CRIRSCO)) (Drielsma et al. 2016). Terms such as “reserves” can therefore be misleading (for a comparison of terms, see Table S2). For example, “Ultimately extractable reserve” (Guinée and Heijungs 1995) and “Extractable global resource” (Drielsma et al. 2016) both relate to the amount of crustal content that will ultimately be extractable, which constitutes the resource stock relevant for depletion (Guinée and Heijungs 1995). The often used USGS reserve base on the other hand is not a fixed stock but its size is defined by technical, economic, legal, and other factors and hence can increase or decrease (Drielsma et al. 2016). Accordingly, use-to-availability ratios can increase or decrease over time when using a dynamic size such as the USGS reserve base or reserves for availability. In the case of copper, for example, on a global scale exploration success still outpaces annual production (Northey et al. 2014). Furthermore, these dynamic sizes underestimate the availability of less explored minerals and metals when compared to well-explored minerals and metals since more exploration efforts increase reserve estimates. Therefore, these methods do not account for dissipation or depletion of a fixed stock and are here labeled use-to-availability ratios (see Table 3). On the other hand, both the ADP_{Ultimate Reserves} (Guinée and Heijungs 1995; van Oers et al. 2002) and the updated version of the AADP methods (Schneider et al. 2015) are examples of use-to-stock ratios (see Table 3). It is acknowledged that, on the one hand, the ultimate reserves (estimated by

260 multiplying the average concentrations of chemical elements in the earth's crust by the mass of the crust) will
261 never be fully accessible. On the other hand, although the ultimately extractable reserves is the only relevant
262 parameter in terms of depletion of the useful (to humans) geological stock, its estimation is always bound to
263 large uncertainties because it depends on the future development of extraction technologies (Guinée and
264 Heijungs 1995). Table 4 summarizes the issues related to different deposit estimates used for use-to- availability
265 ratios.

266
267 <Table 4>

268
269 While Guinée and Heijungs (1995) recommend to use crustal content, Schneider et al. (2015) (AADP method)
270 estimate ultimately extractable reserves as a percentage of crustal content. Both papers acknowledge the implicit
271 assumption that the ratio between the two is equal for all resources. If the natural resource is dissipated into
272 concentrations that are below a threshold that allows for recovery, it is lost and the stock decreases.

273 Type 3 methods relate current resource consumption to potential consequences for future extraction of
274 resources. These methods quantify these potential consequences as: a) additional energy requirements (e.g. Eco-
275 Indicator 99, IMPACT 2002+, and TRACI and TRACI 2); b) additional costs (e.g. EPS 2000/2015, ReCiPe,
276 LIME and LIME 2, Surplus Cost Potential (SCP), and Stepwise2006 (based on additional energy
277 requirements)); or c) additional ore material that has to be dealt with (e.g. Ore Requirement Indicator (ORI) and
278 Surplus Ore Potential (SOP/LC-Impact)). The rationale of type 3 methods is based on the conception that in the
279 long run the effort to extract resources will increase due to declining quality of deposits. Cumulative grade-
280 tonnage relationships have been used to show declining ore grades with increasing cumulative metal produced
281 using the example of copper (see e.g. Gerst 2008; Vieira et al. 2012). However, at the global scale the initial ore
282 grades of new porphyry copper mines have not declined over the past 150 years (Crowson 2012) and there is no
283 apparent decline in the grades of different nickel ores (Mudd and Jowitt 2014). At the more regional scale on the
284 other hand, data for Australia, Canada, and the United States shows a gradual decline of ore grades over time
285 (see e.g. Mudd 2009). This decline also reflects the ageing of mines and the rising share of production from
286 lower-grade ores that became technically accessible with time (Crowson 2012). When lower ore grades are
287 mined, more waste is removed to access the minerals, which generally also leads to increases in energy
288 consumption across mining operations unless investments are made in more efficient processes (EEX 2016). In
289 reality, such investments combined with the closure of old mines and the opening of new mines mean that

relationships between ore grade and energy consumption change within a particular sector or jurisdiction over time. While grade-tonnage relationships have been used to evaluate the physical availability of natural resources, cost-tonnage relationships have been used to account for the economic availability (Vieira et al. 2016a). For the period from 2000 to 2013, available data shows increasing costs and declining ore grades with increasing cumulative copper produced although the causal relationship between ore grade decrease and surplus costs is unknown and the authors acknowledge that data over a longer period would be desirable (Vieira et al. 2016a). Furthermore, as the example of copper shows, technological advances and economies of scale may offset the higher costs of mining lower ore grades (Crowson 2012). However, the long-run need to use lower ore grades and access more remote and more difficult to process deposits, even if it may not be driven by depletion of high grade deposits (West 2011), will eventually lead to increasing opportunity costs, i.e. what society has to sacrifice to get another unit of a mineral or metal (Tilton and Lagos 2007).

3.2. Water

In LCIA, impacts from emissions to water have traditionally been captured by impact categories such as (eco)toxicity, acidification, and eutrophication, which are usually connected to the AoP 'Ecosystem Quality' (Boulay et al. 2014). A general framework connecting water use to other AoP, such as the effects of the depletion of water stock and funds on future generations, has been proposed by Bayart et al. (2010). Several methods have been developed that entirely or partially address the different impact pathways outlined in their framework. A review and analysis of methods is presented in Kounina et al. (2013). Some methods quantify water scarcity/stress based on a use-to-availability ratio (similar to Type 2 methods for abiotic natural resources, see 3.1 and Table 4). However, these methods usually assess a pressure on flow water resources accounting for competition amongst different users and they are not connected to the AoP 'Natural Resources'. Pfister et al. (2009) additionally use a future consequences/surplus energy concept, similar to Type 3 methods above (see 3.1).

The framework for water use by Kounina et al. (2013) (see also Figure S1) follows the reasoning discussed previously: only depletion of (water) stock and fund resources imply a damage to the resource as such in its available form (as surface or groundwater). Fossil groundwater (no or extremely slow replenishment) is the only water stock resource. Slowly replenishing groundwater bodies or stagnant surface water bodies, such as the Aral Sea, can be considered fund resources, since the available amount of water can either be decreased or increased, depending on the ratio of the extraction to renewal rate. Of all water resources (shown in Table 2), only salt

water and rainwater are not considered in impact assessments. Whereas sea water can be considered an unlimited resource, brackish/saline water may be a local stock or fund that could be depleted. Rainwater is one of the resources (e.g. together with solar radiation, wind, or soil) that are acquired through land occupation (Ridoutt and Pfister 2010).

Methods addressing freshwater use are compiled in Table 3 and in more detail in Table S4.

3.3. Land and Water Surface

Land and water surface are finite and usually (the Aral Sea is an example of an exception) constant in total available amount. They cannot be consumed but only occupied, and they become available again for other uses after occupation. Therefore, they can be considered flow resources. The use of a flow resource may have (local) impacts on the temporary availability of, and therefore the competition (among humans and the environment) for, this resource. Therefore, these impacts have not been connected to the AoP 'Natural Resources', but instead to the AoP 'Ecosystem Quality' by several already existing methods assessing land use impacts on biodiversity (see Table 3). Furthermore, land (and water) surface use can be summed up as in the Recipe method at the midpoint level (Goedkoop et al. 2013), and they can be assessed with thermodynamic accounting methods quantifying consumption of exergy or solar energy (type 4, see 3.1). Finally, the Ecological Footprint method quantifies the area necessary to sustain consumption and activities, e.g. of a nation, expressed in units of world-average biologically productive area (Borucke et al. 2013).

3.4. Soil

Soil mass (3D-quantity), its properties, and related soil functions are important in addition to land surface (2D-quantity). Soil is defined as the top layer of the earth's crust formed by mineral particles, organic matter, water, air and living organisms (EC 2015). Soil functions include storing, filtering, cycling and transforming nutrients, substances, and water, biomass production, harboring biodiversity, carbon storage, being a source of raw materials, and being a physical environment for humans. The main threats to soil are erosion, loss of soil organic matter (SOM), compaction, salinization, acidification, contamination, sealing, landslides, flooding, desertification, and soil biodiversity loss (EC 2006; EC 2012; Stoessel et al. 2016). The variety of soil properties and functions and the variety of threats posed to them indicate the complexity of a holistic assessment of impacts on soil and so far no standardized method for a universal assessment of soil-quality impacts has been created (Garrigues et al. 2012; Vidal Legaz et al. 2016). Furthermore, this complexity corresponds to little

agreement on the framework level (EC-JRC 2010; Koellner et al. 2013; Alvarenga et al. 2015). The threats to the resource soil can result in a physical loss of soil (e.g. of arable land by erosion) or in a change of properties (e.g. if SOM is lost) (see Figure S2). However, soil mass and properties can also be preserved or even increased/improved, e.g. by good agricultural practice, and hence fulfill the criteria of a fund resource as defined before. As for water resources, the depletion of these soil fund resources implies a damage to the resource as such in its available form.

Soil assessment methods and models are listed in Table 3 and Table S5. Some of these methods/models are not operational while others are limited to specific countries (Garrigues et al. 2012; Stoessel et al. 2016). They only address partial impacts relevant for soil degradation (e.g. erosion only) and they do not distinguish between different soil management practices (e.g. tillage or nutrient management) or production standards (e.g. organic or integrated production) (Stoessel et al. 2016). Many of the models have excessive data requirements and are therefore difficult to apply, and none of the methods is made compatible to commonly used existing LCIA methods (Stoessel et al. 2016). Globally, operational models are addressing the following impacts: erosion (Núñez et al. 2013; Saad et al. 2013; Scherer and Pfister 2015), loss of SOM (Milà i Canals et al. 2007b: agriculture and forestry only; Brandão and Milà i Canals 2013), compaction (Garrigues et al. 2013), desertification (Núñez et al. 2010), and salinization (Payen et al. 2016). Acidification and contamination are captured with the impact categories ‘Terrestrial Acidification’ and ‘Terrestrial Eco-toxicity’ but these are not connected to the AoP ‘Natural Resources’. There are several multi-criteria indicators to assess changes in soil properties (Cowell and Clift 2000; Oberholzer et al. 2006; Beck et al. 2010), whereby the LANCA approach (Beck et al. 2010) has been operationalized and is used in the method of Saad et al. (2013) and recently by LANCA developers themselves (Bos et al. 2016). Furthermore, there are exergy methods accounting for occupation of land and marine surfaces (Alvarenga et al. 2013; Taelman et al. 2014). Núñez et al. (2013) use the surplus energy concept and estimate the solar energy required to generate one gram of soil lost by erosion. Furthermore, Brandão and Milà i Canals (2013) promote the land’s long-term ability to produce biomass (referred to as biotic production potential (BPP), calculated based on SOM) as an endpoint in the AoP ‘Natural Resources’.

3.5. Biotic Natural Resources

Biotic natural resources have not received much attention yet (Finnveden et al. 2009). These resources are living at least until the moment of extraction from the natural environment and include wood, fish, and other terrestrial

and aquatic biomass that can be harvested (Klinglmair et al. 2014). Agricultural crops, livestock, fish from aquaculture, or wood from a plantation are usually not classified as biotic natural resources in LCA (Klinglmair et al. 2014) since they are the output of a technical process and are hence already part of the technosphere. Impacts on habitats of biotic natural resources are assessed in the AoP 'Ecosystem Quality'. Impacts on biotic natural resources that are of concern in the AoP 'Natural Resources' are caused by overharvesting, overfishing, and overhunting. Such overuse of biotic natural resources may also affect the natural regeneration rate of these fund resources, leading to feedback mechanisms that may cause their depletion.

Aggregating methods considering biotic natural resources are Eco-scarcity, IMPACT 2002+, EPS 2000/2015, LIME/LIME 2, and exergy methods. However, in many cases the only biotic natural resource considered is wood as an energy resource. For instance, the IMPACT 2002+ method applies energy use from wood as a stand-alone indicator, because it is not part of the non-renewable energy indicator (Jolliet et al. 2003). In the Eco-scarcity method, "the energy content of energy resources not used for energy production (feedstock energy, such as when hydrocarbons are used as refrigerants or wood is used in a building), is also assessed with a primary energy factor. However, only the consumed proportion should be assessed" (Frischknecht and Büsser Knöpfel 2013). The EPS 2000/2015 method takes a different approach by including the AoP 'Ecosystem Production Capacity', which accounts for the ecosystem capacity to produce crops, wood, fish and meat, and clean water (Steen 1999; Steen 2015). In the LIME/LIME 2 methods, the impacts on forestry, crops, and fishery are linked to the AoP 'Social Assets', and the damages are measured as user costs, in monetary units (Itsubo et al. 2004; Itsubo and Inaba 2012).

Net Primary Production (NPP) has been used as proxy for damage assessment in the AoP 'Ecosystem Quality' (e.g. Pfister et al. 2009; Taelman et al. 2016), but also as a resource. For instance, Alvarenga et al. (2015) suggest the NPP deficit, which is the assessment of the decrease of biomass availability due to land use, as an indicator for damage assessment in the AoP 'Natural Resources'. They suggest the surplus cost approach, using algae cultivation in the ocean, as the backup technology (Alvarenga et al. 2015).

Methods for overfishing were initially developed within the EU LC-impact project, but these are not yet operational on a global scale (Emanuelsson et al. 2014).

4. Discussion

Natural resources have been categorized and grouped in many ways, as many LCIA methods (and underlying models and indicators) have been developed for assessing damages to different natural resources. While there

seems to be agreement in the scientific community that declining environmental provision of natural resources should be assessed, there is not yet an agreement on which indicator describes this best (e.g. use-to-availability approaches, surplus cost/energy/ore). Furthermore, there is not yet a consensus on whether and how the functionality of a resource should be taken into account.

Figure 1 shows the framework suggested for all resource categories. The depletion or dissipation of stock and fund resources implies a declining environmental provision of natural resources. The use of a flow resource does not imply such a damage, but it may deprive others from using the resource, as a result of competition for it. Competition for natural resources (including competition for stock and fund resources) is an issue that has not yet been explicitly addressed in LCA. However, possible consequences of competition, such as crop failures due to lacking irrigation water, may be assessed as impacts. In the case of water, impacts of deprivation have been linked to the AoPs ‘Human Health’ and ‘Ecosystem Quality’ (Pfister et al. 2009) (dashed arrows pathway in **Figure 1**). Another possible consequence of competition is indirect land use change, which is of interest in consequential LCIA (Schmidt et al. 2015). However, so far there is no generally established methodological approach to address competition for flow (or fund and stock) resources in LCIA. Since it is debatable to what degree competition is an environmental problem, it is up to discussion whether and how this should be further developed. The same applies for all other pathways not yet established in LCIA, represented by dotted arrows in **Figure 1**.

Another issue not yet consistently addressed throughout existing LCIA methods are impacts on resources by other impact categories, such as the effects of global warming on soil productivity. This issue is partly addressed in the IMPACT 2002+ method, in which global warming is listed as a separate impact category, because it is assumed to impact so-called “life supporting functions” (Jolliet et al. 2003). Similar examples are the LIME methods, in which impacts on biotic production is considered (Itsubo et al. 2004; Itsubo and Inaba 2012).

<Figure 1>

Apart from thermodynamic accounting methods, currently there is no all-inclusive method available to assess impacts for all natural resource categories altogether, nor are methods, proposed for different natural resource categories, able to consistently assess these impacts across methods.

Type 2 methods: scarcity and dissipation/depletion

Use-to-availability ratios are concepts that are widely used in LCIA methods. They may account for dissipation or depletion of stock and fund resources and for pressure on flow resources (see **Figure 1**). Concerning minerals and metals, it is especially important to discuss the denominator in the ratio (see section 3.1 and Table 4). Methods using a dynamic size such as the USGS reserves for availability do not account for dissipation or depletion of a fixed stock and might therefore be misleading. However, estimating the geological stock relevant for dissipation or depletion (i.e. the amount of crustal content that will ultimately be extractable) is also bound to large uncertainties because it depends on the future development of extraction technologies. The two approaches taken for estimating fixed stocks are (i) setting the full crustal content as the availability of the resource (although it will never be fully accessible), and (ii) setting the ultimately extractable resource amount as a percentage of crustal content. Both approaches implicitly assume that the ratio between the crustal content and the ultimately extractable amount is equal for all minerals and metals.

Withdrawal-to-availability and consumption-to-availability ratios have been used to assess water stress or water scarcity. They usually consider the flow resource surface water. However, where the calculated ratio is larger than one, groundwater bodies (stocks or funds) or large surface water bodies (funds) are being depleted as assessed in the method by Pfister et al. (2009). Another issue concerning water availability (to humans) is whether the demand of ecosystems should be considered, and if so how large this demand is (different methods provide values from 35 to 80%) (Boulay et al. 2015).

A special case of a use-to-availability ratio to assess scarcity is the distance-to-target ratio. The Eco-scarcity method is based on this concept using the “current flow” of an environmental pressure (e.g. an emission) and the “critical flow” representing the political target in a weighting step (Frischknecht and Büsser Knöpfel 2013). Efforts to include carrying capacity or planetary boundaries in LCIA have introduced a (distance-to-target) normalization against carrying capacity-based references calculated with scientifically estimated thresholds for different impact categories (Bjørn and Hauschild 2015).

Finally, it should be noted that physical availability may not be the dominating factor when referring to environmental impacts. For instance, for minerals/metals and fossil fuels, greenhouse gas emissions and the climate effect these emissions produce may be of more environmental concern than the availability of these resources (Mudd and Ward 2008; McGlade and Ekins 2015).

Type 3 methods: declining quality and consequent future efforts

Stewart and Weidema (2005) defined two key variables when modelling impacts on natural resources: ultimate quality limit and backup technology. The ultimate quality limit is the limit differentiating whether a material is

reusable with a lower functionality, or rendered unavailable (Stewart and Weidema 2005). Backup technology refers to both the technology applied to recycle a material and the alternative technology applied when reaching the ultimate quality limit, *i.e.* when the material is lost (Stewart and Weidema 2005). Common examples are the desalination of water and the consumption of shale gas and oil sands. It has been discussed whether future efforts (use of backup technologies) of current resource dissipation should be part of the impact assessment or part of the inventory (Finnveden 2005). However, type 3 methods seem to understand these future efforts as a proxy for quantifying the difficulty to access natural resources in the future and hence for quantifying an impact on natural resource provision.

The concept of long-term increasing efforts to access natural resources, as a result of declining quality, has been investigated for several natural resource categories. It has first been applied to minerals/metals and fossil fuels. The decision about which deposits of different quality (e.g. ore grade concentration) are extracted (or defined as extractable) depends (among other factors) on production costs. This is the reason why some LCIA methods use increasing future extraction costs as an endpoint unit. Furthermore, it is generally true that more energy is needed to exploit lower grade ores with the same technology. This is the reason why some methods use increasing energy demand for future extraction as an endpoint unit. Technological advances and economies of scale have offset higher costs of mining lower ore grades in the past and assumptions of increased costs and energy consumption of future resource extraction are highly uncertain. However, since LCA is indicating potential impacts for comparison on a common scale, these methods might still be used to account for declining resource quality. Type 3 methods differ in assumptions, e.g. concerning discount rates to calculate future costs. Even within the ReCiPe method for instance, different characterization factors calculated with different discount rates are provided. However, the fundamental principle (declining quality leading to increasing efforts for resource extraction) remains the same. A backup technology approach assessing surplus costs or energy has also been proposed for water (Pfister et al. 2009) and for biotic natural resources (net primary production) (Alvarenga et al. 2015). Some future effort methods for mineral resources avoid a translation into additional costs or energy requirements and account for potentially increasing ore requirements per mineral/metal extracted. This potential future burden is not related to a backup technology that might be used but to physical mass that may have to be dealt with. There are no similar methods like this last subtype of future effort methods for other natural resource categories.

Type 4 methods: thermodynamic accounting

Thermodynamic accounting methods or methods quantifying consumption of exergy or solar energy are able to capture the widest range of natural resource categories (see Table 3). As they consider the consumed quantities, they could be helpful in resource efficiency calculations. However, these methods do not link exergy consumption to changes in availability or provisioning capacity of the natural resource (mineral, water, land etc.) that is consumed.

Quality, functionality, recycling, substitutability

The UNEP-SETAC Life Cycle Initiative overall framework acknowledges the instrumental value of natural resources, which also depends on their quality and related functionality. Natural resources and (raw) materials are lost if the required qualities for their functionality are lost (e.g. through dissipation). However, these properties may be restored or even enhanced further through recycling and upcycling efforts. If this is not possible, the material may either be used for other purposes or it is lost. However, even when a material is “lost” to humans, its functionality may be replaced by other materials made from other natural resources.

Stewart and Weidema (2005) suggest a conceptual framework focusing on the functionality of natural resources. Methodologically, this approach implies that the quality and functionality of the input and output flows of a production system need to be recorded in the LCI in order to assess whether a natural resource is lost at its functionality level (Stewart and Weidema 2005). This issue has, for example, been addressed for water where water qualities needed for different uses were categorized (Boulay et al. 2011; Bayart et al. 2014).

The use of secondary/recycled and treated materials can lower the demand for natural resources (**Figure 1**). This use is typically modeled in the inventory phase. However, whether the use of recycled materials or the output of recyclable materials should get the environmental credits depends on the allocation modeling choice (Frischknecht 2010). Existing methods only roughly consider material quality, if at all, assuming “functional equivalence” of the substituted material. By contrast, the exergy efficiency approach explicitly considers both the quality of input and output materials. However, exergy might not be the only relevant quality criteria. For a proper inclusion of such criteria, metrics for quality and functionality would need to be defined and recorded in life cycle inventories.

Another aspect leading to the reduction of resource availability by reducing resource quality is the impact on natural resources caused by emissions, such as the pollution of groundwater bodies.

Research needs

In order to further improve impact assessment in the AoP ‘Natural Resources’, the discussion on whether resources should be a part of environmental LCA should be replaced by debates about 1) how environmental issues (we suggest natural provisioning capacity) can best be assessed and 2) how other aspects (e.g. short-term (market) availability) can be assessed in a complementary way. The integration of different resource categories into an AoP ‘Natural Resources’ involves some major challenges. While the distinction of stocks, funds, and flows is helpful, these categories still have to be better defined based on regeneration rates. Furthermore, a deeper discussion on whether and how impacts from competition for resources should be integrated in LCIA is needed. In addition, if ecosystem-relevant resources (land, soil, water, and biotic natural resources) and others (minerals/metals and fossil fuels) are to be assessed with a common unit within the same AoP, impact modelling has to be adapted.

5. Conclusions

The environment’s capacity to provide natural resources of a useful quality with instrumental value to humans is what should be protected under the AoP ‘Natural Resources’. However, we know neither how technological developments influence future accessibility nor what the needs of future generations are. While it is true that because of the instrumental value the issue of concern is actually the functionality of a natural resource, information on the functionality and substitutability of resources is mostly incomplete, especially with regard to the future consumption of resources. Therefore, for the time being, it makes sense to devote time to the assessment of environmental provisioning capacity of natural resources. Thereby, the concept of stock/fund/flow resources is helpful, across natural resource categories, in identifying whether depletion/dissipation (of stocks and funds) or competition (for flows) is the main relevant issue. The former has been of primary interest for the AoP ‘Natural Resources’ and accordingly the damage has been described as a reduced availability of, or as a more onerous access to, natural resources in the future (see e.g. Udo de Haes et al. 2002; Jolliet et al. 2004; Bayart et al. 2010). Two main types of methods/models have been used to account for this: 1) use-to-stock/availability methods focus mainly on the quantitative availability; 2) future effort methods focus more on resource quality and corresponding efforts to make the resource usable. Both method types have been used for several resource categories, but no set of methods is yet available to consistently capture all natural resource categories, except for exergy and solar energy methods. However, the fact that exergy and solar energy methods do not explicitly link exergy consumption to changes in availability or

provisioning capacity of the natural resource (mineral, water, land etc.) that is consumed may be considered to be a drawback.

An LCIA method - or a set of methods - that consistently addresses all natural resource categories is needed in order to assess the AoP 'Natural Resources' in a comprehensive manner and to avoid burden shifting from impacts on one resource to impacts on another resource. This paper reviewed existing LCIA methods/models addressing natural resources and discussed their conceptual approaches across different natural resource categories, which is an important prerequisite for a step in this direction.

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6. Literature

- Alvarenga RAF, Dewulf J, Van Langenhove H, Huijbregts MAJ (2013) Exergy-based accounting for land as a natural resource in life cycle assessment. *Int J Life Cycle Assess* 18:939–947. doi: 10.1007/s11367-013-0555-7
- Alvarenga RAF, Erb K-H, Haberl H, et al (2015) Global land use impacts on biomass production—a spatial-differentiated resource-related life cycle impact assessment method. *Int J Life Cycle Assess* 440–450. doi: 10.1007/s11367-014-0843-x
- Bare J (2011) TRACI 2.0: The tool for the reduction and assessment of chemical and other environmental impacts 2.0. *Clean Technol Environ Policy* 13:687–696. doi: 10.1007/s10098-010-0338-9
- Bare J, Norris GA, Pennington DW (2003) TRACI - The Tool for the Reduction and Assessment of Chemical and Other Environmental Impacts. *J Ind Ecol* 6:49–78. doi: 10.1162/108819802766269539
- Bayart J-B, Bulle C, Deschênes L, et al (2010) A framework for assessing off-stream freshwater use in LCA. *Int J Life Cycle Assess* 15:439–453. doi: 10.1007/s11367-010-0172-7
- Bayart J-B, Worbe S, Grimaud J, Aoustin E (2014) The Water Impact Index: A simplified single-indicator approach for water footprinting. *Int J Life Cycle Assess* 19:1336–1344. doi: 10.1007/s11367-014-0732-3
- Beck T, Bos U, Wittstock B, et al (2010) LANCA® Land Use Indicator Value Calculation in Life Cycle Assessment. Fraunhofer Verlag, Stuttgart
- Berger M, van der Ent R, Eisner S, et al (2014) Water accounting and vulnerability evaluation (WAVE): Considering atmospheric evaporation recycling and the risk of freshwater depletion in water footprinting. *Environ Sci Technol* 48:4521–4528. doi: 10.1021/es404994t
- Bjørn A, Hauschild MZ (2015) Introducing carrying capacity based normalization in LCA: framework and development of references at midpoint level. *Int J Life cycle Assess* 1005–1018. doi: 10.1007/s11367-015-0899-2
- Borucke M, Moore D, Cranston G, et al (2013) Accounting for demand and supply of the biosphere's regenerative capacity: The National Footprint Accounts' underlying methodology and framework. *Ecol Indic* 24:518–533. doi: 10.1016/j.ecolind.2012.08.005
- Bos U, Horn R, Beck T (2016) LANCA® Characterization Factors for Life Cycle Impact Assessment - Version 2.0.

Fraunhofer Verlag, Stuttgart

- Bösch ME, Hellweg S, Huijbregts MAJ, Frischknecht R (2007) Applying cumulative exergy demand (CExD) indicators to the ecoinvent database. *Int J Life Cycle Assess* 12:181–190. doi: 10.1007/s11367-006-0282-4
- Boulay A-M, Bare J, De Camillis C, et al (2015) Consensus building on the development of a stress-based indicator for LCA-based impact assessment of water consumption: outcome of the expert workshops. *Int J Life Cycle Assess* 577–583. doi: 10.1007/s11367-015-0869-8
- Boulay A-M, Bulle C, Bayart J-B, et al (2011) Regional characterization of freshwater use in LCA: Modeling direct impacts on human health. *Environ Sci Technol* 45:8948–8957. doi: 10.1021/es1030883
- Boulay A-M, Motoshita M, Pfister S, et al (2014) Analysis of water use impact assessment methods (Part A): Evaluation of modeling choices based 15 on a quantitative comparison of scarcity and human health indicators. *Int J Life Cycle Assess* 139–160. doi: 10.1007/s11367-014-0814-2
- Brandão M, Milà i Canals L (2013) Global characterisation factors to assess land use impacts on biotic production. *Int J Life Cycle Assess* 18:1243–1252. doi: 10.1007/s11367-012-0381-3
- Cowell SJ, Clift R (2000) A methodology for assessing soil quantity and quality in life cycle assessment. *J Clean Prod* 8:321–331. doi: 10.1016/S0959-6526(00)00023-8
- Crowson P (2012) Some observations on copper yields and ore grades. *Resour Policy* 37:59–72. doi: 10.1016/j.resourpol.2011.12.004
- Dewulf J, Benini L, Mancini L, et al (2015a) Rethinking the Area of Protection “Natural Resources” in Life Cycle Assessment. *Environ Sci Technol* 5310–5317. doi: 10.1021/acs.est.5b00734
- Dewulf J, Boesch ME, De Meester B, et al (2007) Cumulative Exergy Extraction from the natural environment (CEENE): a comprehensive Life Cycle Impact Assessment method for resource accounting. *Environ Sci Technol* 41:8477–8483.
- Dewulf J, Mancini L, Blengini GA, et al (2015b) Toward an Overall Analytical Framework for the Integrated Sustainability Assessment of the Production and Supply of Raw Materials and Primary Energy Carriers. *J Ind Ecol* 0:n/a-n/a. doi: 10.1111/jiec.12289
- Drielsma JA, Russell-Vaccari AJ, Drnek T, et al (2016) Mineral resources in life cycle impact assessment—defining the path forward. *Int J Life Cycle Assess* 21:85–105. doi: 10.1007/s11367-015-0991-7

EC (2015) Soil. http://ec.europa.eu/environment/soil/index_en.htm. Accessed 18 Feb 2016

EC (2006) Establishing a framework for the protection of soil and amending Directive 2004/35/EC. European Commission, Brussels

EC (2012) The implementation of the Soil Thematic Strategy and ongoing activities. European Commission, Brussels

EC-JRC (2010) International Reference Life Cycle Data System (ILCD) Handbook: Framework and Requirements for Life Cycle Impact Assessment Models and Indicators. European Commission, Joint Research Centre, Institute for Environment and Sustainability, Ispra

EC-JRC (2011) International Reference Life Cycle Data System (ILCD) Handbook: Recommendations for Life Cycle Impact Assessment in the European context. European Commission, Joint Research Centre, Institute for Environment and Sustainability, Ispra

Ecoinvent (2015) Activity Overview for ecoinvent 3.2, Undefined. <http://www.ecoinvent.org/support/documents-and-files/documents-and-files.html>. Accessed 1 Feb 2015

EEX (2016) Mining. <http://eex.gov.au/industry-sectors/mining/>. Accessed 17 Mar 2016

Emanuelsson A, Ziegler F, Pihl L, et al (2014) Accounting for overfishing in life cycle assessment: New impact categories for biotic resource use. *Int J Life Cycle Assess* 19:1156–1168. doi: 10.1007/s11367-013-0684-z

Finnveden G (2005) The Resource Debate Needs to Continue. *Int J Life Cycle Assess* 10:372. doi: 10.1065/lca2005.09.002

Finnveden G, Hauschild MZ, Ekvall T, et al (2009) Recent developments in Life Cycle Assessment. *J Environ Manage* 91:1–21.

Fischer-Kowalski M, Swilling M (2011) Decoupling Natural Resource Use and Environmental Impacts from Economic Growth. United Nations Environment Programme (UNEP)

Frischknecht R (2010) LCI modelling approaches applied on recycling of materials in view of environmental sustainability, risk perception and eco-efficiency. *Int J Life Cycle Assess* 15:666–671. doi: 10.1007/s11367-010-0201-6

Frischknecht R, Büsler Knöpfel S (2013) Swiss Eco-Factors 2013 according to the Ecological Scarcity Method. Methodological fundamentals and their application in Switzerland. Federal Office for the Environment

FOEN, Bern

Frischknecht R, Jolliet O (2016) Global Guidance for Life Cycle Impact Assessment Indicators: Volume 1. UNEP/SETAC Life Cycle Initiative, Paris

Frischknecht R, Jungbluth N, Althaus H, et al (2007) Overview and Methodology. ecoinvent Report No. 1. Swiss Centre for Life Cycle Inventories, Dübendorf

Garrigues E, Corson MS, Angers DA, et al (2013) Development of a soil compaction indicator in life cycle assessment. *Int J Life Cycle Assess* 18:1316–1324. doi: DOI 10.1007/s11367-013-0586-0

Garrigues E, Corson MS, Angers D a., et al (2012) Soil quality in Life Cycle Assessment: Towards development of an indicator. *Ecol Indic* 18:434–442. doi: 10.1016/j.ecolind.2011.12.014

Gerst MD (2008) Revisiting the cumulative grade-tonnage relationship for major copper ore types. *Econ Geol* 103:615–628. doi: 10.2113/gsecongeo.103.3.615

Goedkoop M, Heijungs R, de Schryver A, et al (2013) ReCiPe 2008. A LCIA method which comprises harmonised category indicators at the midpoint and the endpoint level. Characterisation. Ministerie van VROM, Den Haag

Goedkoop M, Spriensma R (2001) The Eco-indicator 99 A damage oriented method for Life Cycle Impact Assessment - Methodology Report. Ministerie van VROM, Den Haag

Guinée JB, Heijungs R (1995) A proposal for the definition of resource equivalency factors for use in product life-cycle assessment. *Environ Toxicol Chem* 14:917–925. doi: 10.1002/etc.5620140525

Hauschild MZ, Goedkoop M, Guinée J, et al (2013) Identifying best existing practice for characterization modeling in life cycle impact assessment. *Int J Life Cycle Assess* 18:683–697. doi: 10.1007/s11367-012-0489-5

Hauschild MZ, Huijbregts MAJ (eds) (2015) Life Cycle Impact assessment. Springer, Dordrecht

Heijungs R, Guinée J, Huppes G (1997) Impact categories for natural resources and land use. Centre of Environmental Science (CML), Leiden

Hoekstra AY, Mekonnen MM, Chapagain AK, et al (2012) Global monthly water scarcity: Blue water footprints versus blue water availability. *PLoS One*. doi: 10.1371/journal.pone.0032688

Huijbregts MAJ, Hellweg S, Frischknecht R, et al (2010) Cumulative energy demand as predictor for the environmental burden of commodity production. *Environ Sci Technol* 44:2189–2196. doi:

10.1021/es902870s

ISO (2006) ISO 14040: Environmental management - life cycle assessment - principles and framework.
International Organisation for Standardisation, Geneva

Itsubo N, Inaba A (2012) Lime2. JLCA NewsL Life-Cycle Assess Soc Japan 16.

Itsubo N, Sakagami M, Washida T, et al (2004) Weighting across safeguard subjects for LCIA through the
application of conjoint analysis. *Int J Life Cycle Assess* 9:196–205. doi: 10.1007/BF02994194

Jolliet O, Margni M, Charles R, et al (2003) IMPACT 2002 + : A New Life Cycle Impact Assessment Methodology.
Int J Life Cycle Assess 8:324–330. doi: 10.1007/BF02978505

Jolliet O, Müller-Wenk R, Bare J, et al (2004) The LCIA Midpoint-damage Framework of the UNEP/SETAC Life
Cycle Initiative. *Int J Life Cycle Assess* 9:394–404.

Klinglmaier M, Sala S, Brandão M (2014) Assessing resource depletion in LCA: A review of methods and
methodological issues. *Int J Life Cycle Assess* 19:580–592. doi: 10.1007/s11367-013-0650-9

Koellner T, de Baan L, Beck T, et al (2013) UNEP-SETAC guideline on global land use impact assessment on
biodiversity and ecosystem services in LCA. *Int J Life Cycle Assess* 18:1185–1187. doi: 10.1007/s11367-
013-0580-6

Kounina A, Margni M, Bayart J-B, et al (2013) Review of methods addressing freshwater use in life cycle
inventory and impact assessment. *Int J Life Cycle Assess* 18:707–721. doi: 10.1007/s11367-012-0519-3

Langlois J, Fréon P, Delgenes J-P, et al (2014) New methods for impact assessment of biotic-resource depletion
in life cycle assessment of fisheries: theory and application. *J Clean Prod* 73:63–71. doi:
10.1016/j.jclepro.2014.01.087

Lindeijer E, Müller-Wenk R, Steen B (2002) Impact Assessment of Resources and Land Use. In: Udo de Haes HA,
Finnveden G, Goedkoop M, et al. (eds) *Life-Cycle Impact Assessment: Striving towards Best Practice*.
Society of Environmental Toxicology and Chemistry (SETAC), Pensacola FL, USA, pp 11–64

Mancini L, Camillis C De, Pennington D (2013) Security of supply and scarcity of raw materials.

McGlade C, Ekins P (2015) The geographical distribution of fossil fuels unused when limiting global warming to
2 °C. *Nature* 517:187–190. doi: 10.1038/nature14016

Milà i Canals L, Bauer C, Depestele J, et al (2007a) Key elements in a framework for land use impact
assessment within LCA. *Int J Life Cycle Assess* 12:5–15. doi: 10.1065/lca2006.05.250

- Milà i Canals L, Chenoweth J, Chapagain A, et al (2009) Assessing freshwater use impacts in LCA: Part I - Inventory modelling and characterisation factors for the main impact pathways. *Int J Life Cycle Assess* 14:28–42. doi: 10.1007/s11367-008-0030-z
- Milà i Canals L, Romanyà J, Cowell SJ (2007b) Method for assessing impacts on life support functions (LSF) related to the use of “fertile land” in Life Cycle Assessment (LCA). *J Clean Prod* 15:1426–1440. doi: 10.1016/j.jclepro.2006.05.005
- Mudd GM (2009) Historical trends in base metal mining: backcasting to understand the sustainability of mining. In: Canadian Metallurgical Society (ed) 48th Annual Conference of Metallurgists proceedings. Sudbury, Ontario, Canada,
- Mudd GM, Jowitt SM (2014) A Detailed Assessment of Global Nickel Resource Trends and Endowments. *Econ Geol* 109:1813–1841. doi: 10.2113/econgeo.109.7.1813
- Mudd GM, Ward JD (2008) Will Sustainability Constraints Cause “Peak Minerals”? In: 3rd International Conference on Sustainability Engineering & Science: Blueprints for Sustainable Infrastructure. Auckland, New Zealand,
- Northey S, Mohr S, Mudd GM, et al (2014) Modelling future copper ore grade decline based on a detailed assessment of copper resources and mining. *Resour Conserv Recycl* 83:190–201. doi: 10.1016/j.resconrec.2013.10.005
- Núñez M, Antón A, Muñoz P, Rieradevall J (2013) Inclusion of soil erosion impacts in life cycle assessment on a global scale: application to energy crops in Spain. *Int J Life Cycle Assess* 18:755–767. doi: 10.1007/s11367-012-0525-5
- Núñez M, Civit B, Muñoz P, et al (2010) Assessing potential desertification environmental impact in life cycle assessment. *Int J Life Cycle Assess* 15:67–78. doi: 10.1007/s11367-009-0126-0
- Oberholzer H-R, Weisskopf P, Gaillard G, et al (2006) Methode zur Beurteilung der Wirkungen landwirtschaftlicher Bewirtschaftung auf die Bodenqualität in Ökobilanzen. Agroscope FAL Reckenholz
- Payen S, Basset-Mens C, Núñez M, et al (2016) Salinisation impacts in life cycle assessment: a review of challenges and options towards their consistent integration. *Int J Life Cycle Assess* 577–594. doi: 10.1007/s11367-016-1040-x
- Pfister S, Koehler A, Hellweg S (2009) Assessing the Environmental Impact of Freshwater Consumption in Life

Cycle Assessment. *Environ Sci Technol* 43:4098–4104. doi: 10.1021/es802423e

Potting J, Hauschild MZ (2005) Spatial Differentiation in Life Cycle Impact Assessment - The EDIP 2003 methodology. Danish Environmental Protection Agency

Ridoutt BG, Pfister S (2010) A revised approach to water footprinting to make transparent the impacts of consumption and production on global freshwater scarcity. *Glob Environ Chang Policy Dimens* 20:113–120. doi: DOI 10.1016/j.gloenvcha.2009.08.003

Rørbech JT, Vadenbo C, Hellweg S, Astrup TF (2014) Impact assessment of abiotic resources in LCA : Quantitative comparison of selected characterization models. *Environ Sci Technol* 48:11072–11081.

Rugani B, Huijbregts MAJ, Mutel CL, et al (2011) Solar Energy Demand (SED) of Commodity Life Cycles. *Environ Sci Technol* 45:5426–5433. doi: 10.1021/es103537f

Saad R, Koellner T, Margni M (2013) Land use impacts on freshwater regulation, erosion regulation, and water purification: A spatial approach for a global scale level. *Int J Life Cycle Assess* 18:1253–1264. doi: 10.1007/s11367-013-0577-1

Saurat M, Ritthoff M (2013) Calculating MIPS 2.0. *Resources* 2:581–607. doi: 10.3390/resources2040581

Scherer L, Pfister S (2015) Modelling spatially explicit impacts from phosphorus emissions in agriculture. *Int J Life Cycle Assess* 20:785–795. doi: 10.1007/s11367-015-0880-0

Schmidt JH, Weidema BP, Brandão M (2015) A Framework for Modelling Indirect Land Use Changes in Life Cycle Assessment. *J Clean Prod* 99:230–238. doi: 10.1016/j.jclepro.2015.03.013

Schneider L, Berger M, Finkbeiner M (2015) Abiotic resource depletion in LCA - background and update of the anthropogenic stock extended abiotic depletion potential (AADP) model. *Int J Life Cycle Assess* 20:709–721. doi: 10.1007/s11367-015-0864-0

Schneider L, Berger M, Finkbeiner M (2011) The anthropogenic stock extended abiotic depletion potential (AADP) as a new parameterisation to model the depletion of abiotic resources. *Int J Life Cycle Assess* 16:929–936. doi: 10.1007/s11367-011-0313-7

Steen B (1999) A systematic approach to environmental priority strategies in product development. CPM - Centre for Environmental Assessment of Products and Material Systems

Steen B (2015) The EPS 2015d impact assessment method – an overview.

Steen BA (2006) Abiotic Resource Depletion. Different perceptions of the problem with mineral deposits. *Int J*

Life Cycle Assess 11:49–54. doi: 10.1065/lca2006.04.011

Stewart M, Weidema B (2005) A consistent framework for assessing the impacts from resource use: A focus on resource functionality. *Int J Life Cycle Assess* 10:240–247. doi: 10.1065/lca2004.10.184

Stoessel F, Bachmann D, Hellweg S (2016) Assessing the environmental impacts of agricultural production on soil in a global Life Cycle Impact Assessment method: A framework. In: *LCA Food 2016 - Proceedings*.

Swart P, Alvarenga RAF, Dewulf J (2015) Abiotic Resource Use. In: *Life Cycle Impact Assessment*. Springer, Dordrecht, pp 247–271

Swart P, Dewulf J (2013) Quantifying the impacts of primary metal resource use in life cycle assessment based on recent mining data. *Resour Conserv Recycl* 73:180–187. doi: 10.1016/j.resconrec.2013.02.007

Taelman SE, De Meester S, Schaubroeck T, et al (2014) Accounting for the occupation of the marine environment as a natural resource in life cycle assessment: An exergy based approach. *Resour Conserv Recycl* 91:1–10. doi: 10.1016/j.resconrec.2014.07.009

Taelman SE, Schaubroeck T, De Meester S, et al (2016) Accounting for land use in life cycle assessment: The value of NPP as a proxy indicator to assess land use impacts on ecosystems. *Sci Total Environ* 550:143–156. doi: 10.1016/j.scitotenv.2016.01.055

Tilton JE, Lagos G (2007) Assessing the long-run availability of copper. *Resour Policy* 32:19–23. doi: 10.1016/j.resourpol.2007.04.001

Udo de Haes HA, Finnveden G, Goedkoop M, et al (eds) (2002) *Life-Cycle Impact Assessment: Striving towards Best Practice*. Society of Environmental Toxicology and Chemistry (SETAC), Pensacola FL, USA

van Oers L, de Koning A, Guinée JB, Huppes G (2002) Abiotic resource depletion in LCA. Road and Hydraulic Engineering Institute of the Dutch Ministry of Transport

Vidal Legaz B, Maia De Souza D, Teixeira RFM, et al (2016) Soil quality, properties, and functions in Life Cycle Assessment: an evaluation of models. *J Clean Prod* 1–14. doi: 10.1016/j.jclepro.2016.05.077

Vieira MDM, Goedkoop MJ, Storm P, Huijbregts MAJ (2012) Ore grade decrease as life cycle impact indicator for metal scarcity: The case of copper. *Environ Sci Technol* 46:12772–12778. doi: 10.1021/es302721t

Vieira MDM, Ponsioen TC, Goedkoop MJ, Huijbregts MAJ (2016a) Surplus Cost Potential as a Life Cycle Impact Indicator for Metal Extraction. *Resources* 5:2. doi: 10.3390/resources5010002

Vieira MDM, Ponsioen TC, Goedkoop MJ, Huijbregts MAJ (2016b) Surplus Ore Potential as a Scarcity Indicator

- for Resource Extraction. J Ind Ecol. doi: 10.1111/jiec.12444
- Weidema B, Finnveden G, Stewart M (2005) Impacts from Resource Use A common position paper. Int J Life Cycle Assess 9:382. doi: <http://dx.doi.org/10.1065/lca2005.11.003>
- Weidema BP, Hauschild MZ, Jolliet O (2007) Preparing characterisation methods for endpoint impact assessment. Available from lca-net.com/files/Stepwise2006v1.5.3.zip
- West J (2011) Decreasing Metal Ore Grades: Are They Really Being Driven by the Depletion of High-Grade Deposits? J Ind Ecol 15:165–168. doi: 10.1111/j.1530-9290.2011.00334.x
- WTO (2010) World Trade Report 2010: Trade in natural resources. World Trade Organization (WTO)

7. Tables

Table 1 Classification of potential resource problems according to renewability and exhaustibility of resources

	Renewability	Exhaustibility
Potential stock problem	non-renewable	exhaustible
Potential fund problem	renewable	exhaustible
Potential flow problem	re-occurring or permanently present	non-exhaustible

802 **Table 2** Compilation and categorization of natural resources based on Klinglmair et al. (2014), Dewulf et al. (2015a), Goedkoop et al. (2013), and Frischknecht and Büsser
803 Knöpfel (2013), including the corresponding elementary flows in the *Ecoinvent 2.2* and *3.2* databases (Frischknecht et al. 2007; Ecoinvent 2015)

Resource Categories		Resource(s)	Stock/Fund/Flow resource problem	Resources in inventory according to <i>Ecoinvent 2.2 and 3.2</i>
Minerals and metals	Aggregates	Rock	Stock	e.g. Granite, Shale...
		Gravel	Stock/Fund	Gravel, in ground
		Sand	Stock/Fund	Sand, unspecified, in ground
		Clay	Stock/Fund	Clay, bentonite, in ground; Clay, unspecified, in ground
Radioactive elements	Elements	Minerals	Stock	e.g. Anhydrite, Dolomite...
		Metals	Stock	e.g. Copper, Gold...
		Elements in water	Stock	Bromine, Iodine, Magnesium
		Elements in air	Stock	Krypton, Xenon
Air components		Uranium (and others)	Stock	Uranium, in ground
Fossil fuels		Air components	Stock	Carbon dioxide, Nitrogen, Oxygen, Argon-40
	Coal	Peat	Stock	Peat, in ground
		Brown coal	Stock	Coal, brown, in ground
		Black coal	Stock	Coal, hard, unspecified, in ground
		Petroleum	Stock	Oil, crude, in ground
(Abiotic) renewable energy sources	Oil & gas ^a	Natural gas	Stock	Gas, natural, in ground
		Solar power	Flow	Energy, solar, converted
		Wind power	Flow	Energy, kinetic (in wind), converted
		Hydropower	Potential	Energy, potential (in hydropower reservoir), converted
Water	Salt water	Potential	Fund/Flow	
		Wave power	Flow	
		Tidal power	Flow	
		Geothermal power	Flow/Fund	Energy, geothermal, converted
	Freshwater	Sea water	Flow/Fund	Water, salt, ocean
		Brackish/saline water	Stock/Fund	Water, salt, sole
		Surface water	Flow/Fund	Water: river, lake, cooling, turbine use, unspecified
		Groundwater	Fund/Flow	Water, well, in ground
Land and water surface	Soil ^b	Fossil groundwater	Stock	Water, well, in ground
		Water in air	Flow/Fund	Water, in air
		Land surface	Flow (competition for area)	Land occupation/transformation (various categories)
		Water surface	Flow (competition for area)	Land occupation/transformation: inland waterbody, lake, river, wetland, unspecified
Biotic natural resources	Flora: terrestrial	Sea(bed) surface	Flow (competition for area)	Land occupation/transformation: seabed
		Soil	Fund	
		Wild plants/wood	Fund	Energy, gross calorific value, in biomass, Wood: hard, primary forest, soft, unspecified
		Wild aquatic flora	Fund	Energy, gross calorific value, in biomass
	Fauna: terrestrial	Game	Fund	Energy, gross calorific value, in biomass
		Wild fish, seafood...	Fund	Energy, gross calorific value, in biomass

^a Including unconventional oil and gas such as shale gas

^b A special case is the consideration of volumes needed to dispose waste in the Ecological Scarcity method

Table 3 Natural resource coverage by method; based on Klinglmair et al. (2014), Rørbech et al. (2014), Hauschild and Huijbregts (2015), and literature indicated

Method/Model	History/Comment	Literature	Minerals & metals	Radioactive elements ^c	Air components	Fossil fuels ^d	(Abiotic) renewable energy sources ^e	Water	Land & water surface	Soil	Biotic natural resources
USE-TO-STOCK/USE-TO-AVAILABILITY											
Metals/Minerals and Fossil Fuels											
CML-IA: ADP ^{Ultimate Reserve} ADP ^{Reserve Base} /ILCD ADP ^{(Economic) Reserve}	Use-to-stock Use-to-availability 2002, Use-to-availability	(Guinée and Heijungs 1995) (van Oers et al. 2002) (van Oers et al. 2002)	48	Yes	-	4	-	-	-	-	-
AADP AADP ^{Update}	Use-to-availability Use-to-stock ^f	(Schneider et al. 2011) (Schneider et al. 2015)	10 35	- -	- -	- -	- -	- -	- -	- -	- -
EDIP 97/2003	Use-to-availability	(Potting and Hauschild 2005)	29	Yes	-	4	Partial ^g	-	-	-	Wood: energy
Eco-scarcity (2013) (Switzerland)	1990, 1997, 2006	(Frischknecht and Büsser Knöpfel 2013)	Yes	Yes	-	4	5	Yes	Yes	-	Wood: energy
Water											
Boulay et al.		(Boulay et al. 2011)	-	-	-	-	-	Yes	-	-	-
Milà i Canals et al.	CML approach (ADP)	(Milà i Canals et al. 2009)	-	-	-	-	-	Yes	-	-	-
WDI		(Berger et al. 2014)	-	-	-	-	-	Yes	-	-	-
WFN Water Scarcity		(Hoekstra et al. 2012)	-	-	-	-	-	Yes	-	-	-
WII		(Bayart et al. 2014)	-	-	-	-	-	Yes	-	-	-
WSI/Pfister et al.		(Pfister et al. 2009)	-	-	-	-	-	Yes	-	-	-
Biotic Natural Resources											
Emanuelsson et al.	OF & OB (see Table S6)	(Emanuelsson et al. 2014)	-	-	-	-	-	-	-	-	Fish
Langlois et al.		(Langlois et al. 2014)	-	-	-	-	-	-	-	-	Fish
FUTURE CONSEQUENCES											
Eco-Indicator 99	1995	(Goedkoop and Spriensma 2001)	12	-	-	4	-	-	(Yes AoP EQ)	-	-
EPS 2000/2015	1996	(Steen 1999) (Steen 2015)	67	Yes	-	3 ^h	-	Yes ⁱ	-	-	(Crops), wood, fish & meat ^j
IMPACT 2002+		(Joliet et al. 2003)	13	Yes	-	5	-	-	(Yes	-	Wood: energy

^c Uranium^d Peat, Brown coal, Black coal, Petroleum, Natural gas, Sulfur^e Solar, Wind, Water, Geothermal^f The resource stocks ultimately available for human use in the long-term are estimated on the basis of the resources in the upper continental crust^g Factors only provided for wood and freshwater at a global level (Hauschild and Huijbregts 2015)^h Only one coal categoryⁱ In AoP 'Ecosystem Production Capacity'^j In AoP 'Ecosystem Production Capacity'

<i>Method/Model</i>	<i>History/Comment</i>	<i>Literature</i>	Minerals & metals	Radioactive elements ^c	Air components	Fossil fuels ^d	(Abiotic) renewable energy sources ^e	Water	Land & water surface	Soil	Biotic natural resources
LC-Impact	see SOP		51 ^k	Yes	-	-	-	(Yes, AoP HH & EQ)	AoP EQ (Yes AoP EQ)	-	-
LIME		(Itsubo et al. 2004)	Yes	-	-	Yes	-	-	(Yes, changes in NPP, AoP EQ)	-	Forest resources consumption
LIME 2		(Itsubo and Inaba 2012)									
ORI		(Swart and Dewulf 2013)	9	-	-	-	-	-	-	-	-
Pfister et al.		(Pfister et al. 2009)	-	-	-	-	-	Yes	-	-	-
ReCiPe (2008)	Based on CML-IA (midpoint only) + EI99 (endpoint only) ^l	(Goedkoop et al. 2013)	19	Yes	-	6	-	Yes (Midpoint)	(Yes AoP EQ)	-	-
SCP		(Vieira et al. 2016a)	12 ^m	Yes	-	-	-	-	-	-	-
SOP/LC- Impact		(Vieira et al. 2016b)	58 ⁿ	Yes	-	-	-	-	-	-	-
Stepwise 2006	Based on EDIP 2003 and IMPACT 2002+	(Weidema et al. 2007)	Yes	Yes	-	Yes	-	-	(Yes AoP EQ)	-	-
TRACI	Fossil fuel assessment based on Eco-Indicator 99	(Bare et al. 2003)	-	-	-	Yes	-	-	(US only AoP EQ)	-	-
TRACI 2		(Bare 2011)									
LOSS OF USEFUL PROPERTY											
Thermodynamic Accounting											
CEENE		(Dewulf et al. 2007) (Taelman et al. 2014)	53	Yes	Yes	4	Yes	Yes	Yes (incl. sea surface)	-	Wood
CExD		(Bösch et al. 2007)	57	Yes	Yes	6	5	Yes	-	-	Wood
Exergy NPP		(Alvarenga et al. 2013)	-	-	-	-	-	-	Exergy/NPP	-	Exergy/NPP
SED		(Rugani et al. 2011)	Yes	Yes	Yes	Yes	Yes	Yes	Yes	-	Yes
Soil											
BPP	Based on SOM	(Brandão and Milà i Canals 2013)	-	-	-	-	-	-	-	SOM	-
Compaction		(Garrigues et al. 2013)	-	-	-	-	-	-	-	Pore volume loss	-
Desertification		(Núñez et al. 2010)	-	-	-	-	-	-	-	Desertification (includes erosion)	-
Erosion		(Núñez et al. 2013)	-	-	-	-	-	-	-	Erosion	-
Erosion and P-loss		(Scherer and Pfister 2015)	-	-	-	-	-	-	-	Erosion & P-loss	-
ERP	Using the LANCA tool (Beck et al. 2010)	(Saad et al. 2013)	-	-	-	-	-	-	-	Erosion	-
LANCA		(Beck et al. 2010; Bos et al. 2016)	-	-	-	-	-	-	-	Several indicators	-
Salinization		(Payen et al. 2016)	-	-	-	-	-	-	-	Salinization	-
SOM		(Milà i Canals et al. 2007b)	-	-	-	-	-	-	-	SOM	-

^k Currently being expanded

^l The midpoint and endpoint of mineral resources are new in ReCiPe

^m Currently being expanded

ⁿ Currently being expanded

<i>Method/Model</i>	<i>History/Comment</i>	<i>Literature</i>	Minerals & metals	Radioactive elements ^c	Air components	Fossil fuels ^d	(Abiotic) renewable energy sources ^e	Water	Land & water surface	Soil	Biotic natural resources
Biotic Natural Resources											
Emanuelsson et al.	LPY (see Table S6)	(Emanuelsson et al. 2014)	-	-	-	-	-	-	-	-	Fish
HANPP		(Alvarenga et al. 2015)	-	-	-	-	-	-	-	-	NPP

Abbreviations: (A)ADP: (Anthropogenic stock extended) Abiotic Depletion Potential, BPP: Biotic Production Potential, CEENE: Cumulative Exergy Extraction from the Natural Environment, CExD: Cumulative Exergy Demand, EQ: Ecosystem Quality, ERP: Erosion Resistance Potential, HANPP: Human Appropriation of Net Primary Production, NPP: Net Primary Production, ORI: Ore Requirement Indicator, P: Phosphorous, SCP: Surplus Cost Potential, SED: Solar Energy Demand, SOM: Soil Organic Matter, SOP: Surplus Ore Potential, URR: Ultimate recoverable resource, WDI: Water Depletion Index, WFN: Water Footprint Network, WII: Water Impact Index

Table 4 Metal/mineral deposits used for use-to-availability ratios according to terminology used by the CML-IA method (Guinée and Heijungs 1995), by the US Geological Service (USGS), and by the Committee for Mineral Reserves International Reporting Standards (CRIRSCO) as reported in Drielsma et al. (2016)

Metal/mineral deposits	Advantages	Disadvantages
<ul style="list-style-type: none"> - (Economic) reserves (CML/USGS)/ Mineral reserves (CRIRSCO) - Reserve base (CML/USGS)/ Mineral Resources (CRIRSCO) - Resources (USGS) 	<ul style="list-style-type: none"> - Based on identified deposits 	<ul style="list-style-type: none"> - Dynamic sizes, no stable indicators - Underestimates extractable metals and minerals (especially if less explored)
Ultimately extractable reserves (CML)/ Extractable Global Resource (Drielsma)	<ul style="list-style-type: none"> - Relevant for depletion of useful (to humans) geological stock - (Theoretically) fixed stock 	<ul style="list-style-type: none"> - Depends on future technological developments, highly uncertain estimations
Ultimate Reserves (CML)/ Crustal content (Drielsma)	<ul style="list-style-type: none"> - Fixed stock - Data available 	<ul style="list-style-type: none"> - Not relevant for depletion of useful (to humans) geological stock because part of it is not accessible

8. Figure Captions

Figure 1 Impact pathways from use of different natural resource types to areas of protection; “competition for resource” means that there is not enough provided to match the demand of all users (including the environment); “within renewability rate” means that the fund resource is used in way that it is not depleted in the long term and that there is no competition; the dashed arrow shows the pathway of how indirect effects of competition have been assessed; the dotted arrows show pathways not yet established in LCIA methods (it is up to discussion whether and how they should be established)

9. Figures

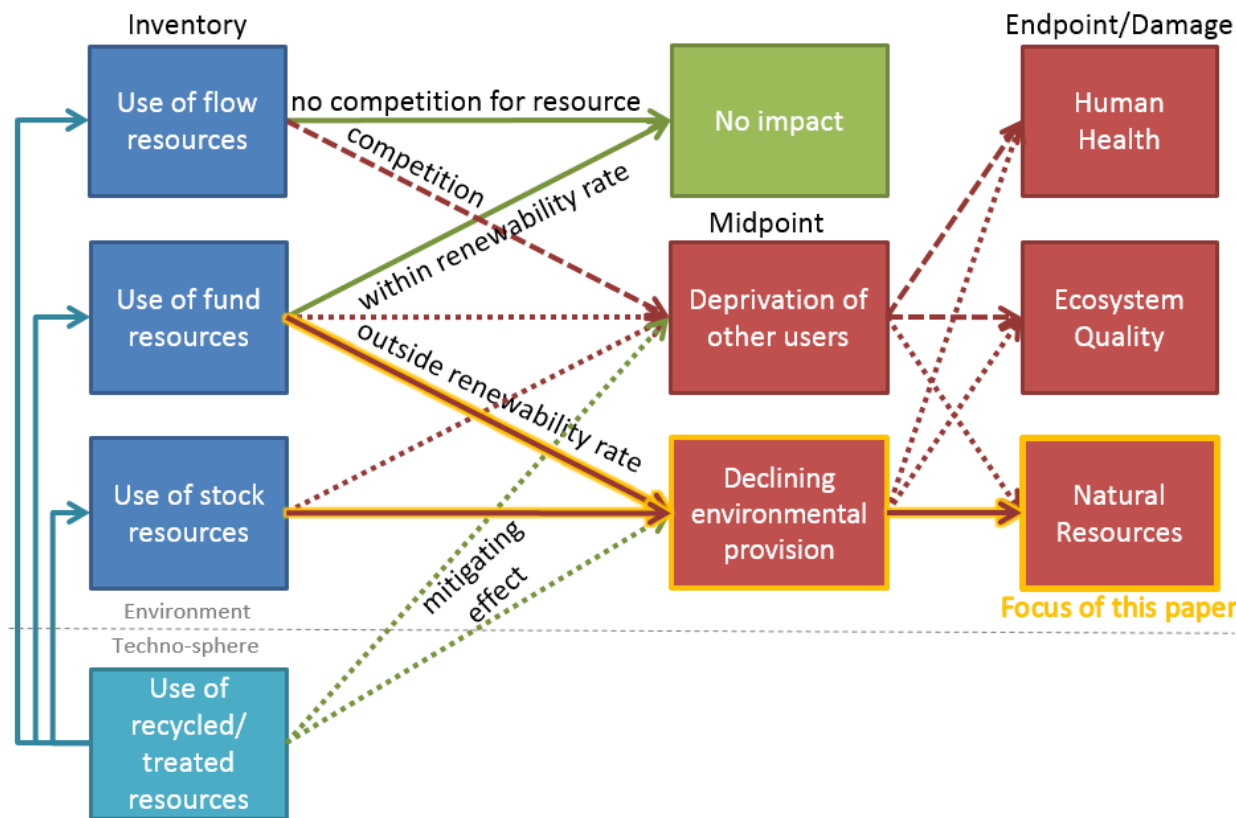


Figure 1 (created with Microsoft Power Point)

Towards harmonizing natural resources as an area of protection in Life Cycle Impact Assessment

SUPPLEMENTARY MATERIAL

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Abbreviations

AoP	Area of Protection
LCA	Life Cycle Assessment
LCIA	Life Cycle Impact Assessment

Areas of Protection in current LCIA methods

Most LCIA methods currently associate the final impact pathway damages to the three AoPs 'Human Health', 'Ecosystem Quality', and 'Natural Resources' ([Table S1](#)). Exceptions are the EPS 2000, the LIME, and the Stepwise method:

The EPS 2000 method has an additional AoP called "Ecosystem Production Capacity", assessing the effect of decreased yields of crop, fish and meat, wood, and freshwater associated with the production capacity of ecosystems (Steen 1999). This includes biotic resources produced by an industrial production process, which – as discussed before – in the LCA context are usually not considered to be natural resources.

The IMPACT 2002+ method models impacts of climate change on life supporting functions as a separate AoP (Joliet et al. 2003). This has been kept for the IMPACT World+ method and water impacts is another optional reporting category (Impact World+ 2016).

The LIME/LIME 2 method is organized in four AoP grouped in two categories: 'Human Life/Human Society' includes the AoPs 'Human Health' and 'Social Assets'; 'Ecosystem' includes the AoPs 'Biodiversity' and 'Primary Production' (of ecosystems) (Itsubo et al. 2004; Itsubo and Inaba 2012). Social assets cover the impact on agricultural products, forests, marine products, mineral resources, and fossil fuels measured in monetary units (Itsubo and Inaba 2012). It could be considered the equivalent to the AoP "Natural Resources" in other methods although – as in the AoP "Ecosystem Production Capacity" of the EPS 2000 method – biotic resources produced by an industrial production process are included.

The Stepwise method does not only consider natural resources but also man-made biotic resources and human resources: "human resources are the available labor force with its different productive abilities, biotic resources are the natural or manipulated biota with its inherent or artificially enhanced abilities to grow and propagate, and abiotic resources are the natural or manufactured raw materials or catalysts for human or biotic production" (Weidema et al. 2007). Accordingly, impacts on resource productivity include the impacts on human health, which indirectly have an impact on human productivity, and the impacts on agricultural production from global warming and photochemical ozone (Weidema et al. 2007).

Table S1: Areas of Protection (AoP) in various LCIA methods; despite different names, the resource-AoPs are usually restricted to mineral and fossil fuels

Method	AoPs				Natural Resources	Connected Impact Categories
	Humans	Ecosystem	Natural Resources	Other	Endpoint unit	
Eco-Indicator 99 (Goedkoop and Spriensma 2001)	Human Health	Ecosystem Quality	Mineral and Fossil Resources	-	MJ surplus energy	- Resource concentration (Minerals & Fossils)
EPS 2000 (Steen 1999)	Human Health	Biodiversity	Abiotic Resources	Ecosystem Production Capacity		
ILCD (EC-JRC 2012)					Cost increase in \$ (from ReCiPe 2008)	Resource depletion – mineral, fossils and renewables
IMPACT 2002+ (Joliet et al. 2003)	Human Health	Ecosystem Quality	Resources	Climate change (Impact on life support functions)	MJ surplus energy	- Non-renewable energy - Mineral extraction
LC-Impact (http://www.lc-impact.eu/about-lc-impact)	Human Health	Ecosystem Quality	Resources		kg _{ore} /kg _x	- Mineral resources depletion
LIME2 (Itsuno and Inaba 2012)	<i>Human Society:</i> Human Health	<i>Ecosystem:</i> Biodiversity	<i>Human Life:</i> Social Assets	<i>Ecosystem:</i> Primary Production	Social cost: Yen	- Farm products - Land - Marine products - Energy - User cost (of mineral resource consumption)
ReCiPe (2008) (Goedkoop et al. 2013)	Human Health	Ecosystems	Resources	-	Cost increase in \$	- Resource concentration (Minerals & Fossils)
Stepwise (2006) (Weidema et al. 2007)	Human Well-Being	Ecosystems	<i>Resource productivity:</i> - Natural resources	<i>Resource productivity:</i> - Human resources - Biotic resources	€ ₂₀₀₃	- Non-renewable energy - Mineral extraction
TRACI 2					MJ surplus energy (from Eco-Indicator 99)	- Resource concentration (Fossils)

Table S2: Definitions of different resource deposits

	CML-IA (Guinée and Heijungs 1995)	USGS	Definition	CRIRSCO/ (Drielsma et al. 2016)	Definition
Stable estimates	Ultimate reserves		"Ultimate reserves are estimated by multiplying the average concentrations of chemical elements in the earth's crust by the mass of the crust. [...] ultimate reserves cannot be extracted completely, as some locations will be inaccessible [...]" (Guinée and Heijungs 1995)	Crustal content	"Crustal content represents the total amount of an element in a given layer of the Earth's crust." (Drielsma et al. 2016)
	Ultimately extractable reserves		"Those reserves that can ultimately be technically extracted may be termed the ultimately extractable reserves. Between the ultimate reserve and the ultimately extractable reserve there is likely to be a substantial difference. In terms of depletion, the ultimately extractable reserve is the only relevant reserve parameter. However, data on this type of reserve are unavailable and will never be exactly known because of their dependence on future technological developments." (Guinée and Heijungs 1995)	Extractable global resource	"Extractable global resource is the amount of crustal content that will ultimately prove extractable by humans." (Drielsma et al. 2016)
Dynamic estimates		Resources	"A concentration of identified naturally occurring mineral in or on the earth's crust in such form and amount that economic extraction is currently or potentially feasible. However resource estimates are based on current knowledge and are continually revised in the context of technological changes and shifts in prices and costs." (Schneider et al. 2015)		
	Reserve base	Reserve base	"The reserve base is that part of an identified resource that meets specified minimum physical and chemical criteria related to current mining practice. It encompasses that fraction of the resources with a reasonable potential for becoming economically available within planning horizons beyond those that assume proven technology and current economics." (Guinée and Heijungs 1995)	Mineral resource	"Mineral resource is a concentration or occurrence of solid material of economic interest in or on the Earth's crust in such form, grade or quality, and quantity that there are reasonable prospects for eventual economic extraction." (Drielsma et al. 2016)
	Economic reserves	Reserves	"The economic reserve is that part of the reserve base that can be economically extracted at the time of determination." (Guinée and Heijungs 1995)	Mineral reserve	"Mineral reserve is the economically mineable part of a measured and/or indicated mineral resource." (Drielsma et al. 2016)

Abbreviations: CML: Centre of Environmental Science, Leiden University; IA: Impact Assessment; CRIRSCO: Committee for Mineral Reserves International Reporting Standards; USGS: United States Geological Service

Table S3: Basic concepts of abiotic resources in LCIA methods; based on Rørbech et al. (2014) and literature indicated in Table 3

Method	Model concept	Unit
Type 2a: Use-to-stock		
CML-IA: ADP _{Ultimate Reserve}	Use-to-stock ratios based on present production and ultimate reserves in the upper crust of the earth	kg Sb-eq
AADP _{Update}	Use-to-stock ratios based on present extraction and the extractable geologic stock + the anthropogenic stock: the extractable geologic stock is estimated on the basis of the resources in the upper continental crust	kg Sb-eq
Type 2b: Use-to-availability		
CML-IA: ADP _{Reserve Base/ILCD}	Use-to-stock ratios based on present production and the reserve base (resources that have a reasonable potential for becoming economically and technically available)	kg Sb-eq
CML-IA: ADP _{(Economic) Reserve}	Use-to-stock ratios based on present production and part of the reserve base which could be economically extracted or produced at the time of determination	kg Sb-eq
AADP	Use-to-availability ratios based on present extraction and the USGS "Resources" (see Table S2) + the anthropogenic stock	kg Sb-eq
LIME & LIME 2	Midpoint: Fossil fuels and Mineral resources	MJ/kg Sb-eq
EDIP 97 & 2003	Economic availability based on economic reserves per person	Person Reserves
Eco-scarcity	Based on ratio of current flow (pollutant load/resource extraction) and critical flow, which is determined by political targets	-
Type 3: Future consequences		
Eco-Indicator 99	Surplus energy representing assumed additional energy requirements for extraction and processing of low grade deposits in the future	MJ
IMPACT 2002+	Surplus energy representing assumed additional energy requirements in future, similar to EI99, with the addition of including the extractable energy content of resources used destructively (fossils and uranium) to the surplus energy of these	MJ
TRACI & TRACI 2 Stepwise	Based on Eco-indicator 99 Difference between the current energy requirement for extraction and an assumed future energy requirement for extraction from lower grade ores	MJ EUR ₂₀₀₃
EPS 2000	Future extraction costs related to mining of average earth crust composition with existing technologies, but using renewable energy sources only	ELU
ReCiPe	Marginal increase in future extraction costs relative to current extraction costs (assumed that future ore concentrations decline when cumulative production increases)	\$
LIME & LIME 2 ORI	Endpoint: User cost (AoP Human Life/Society) Ore requirement indicator: present (i.e. current observed) annual change in ore requirements per kg of metal content	Yen kg/year
SOP/LC-Impact	Surplus Ore Potential: assumed extra amount of ore produced in the future per unit of mineral extracted based on cumulative grade-tonnage relationships	kg _{ore} /kg _x
SCP	Surplus Cost Potential: assumed surplus cost in the future per unit of mineral extracted based on cumulative grade-tonnage relationships	USD ₂₀₁₃ /kg _x
Type 4: Advanced accounting/Thermodynamic losses		
CExD	Cumulative Exergy Demand: total removal of exergy from nature embedded in processed material (including slags and tailing), as exergy difference between the material and a defined reference state in the natural environment, to society	MJ _{ex} -eq
CEENE	Cumulative Exergy Extraction from the Natural Environment: total extraction of exergy from nature embedded in target resources, as the exergy difference between a resource and a defined reference state in the natural environment, to society	MJ _{ex} -eq
SED	Solar Energy Demand: total direct and indirect solar energy requirements needed to provide a product or service	MJ _{se} -eq

Abbreviations: ADP: Abiotic Depletion Potential, AADP: Anthropogenic stock extended Abiotic Depletion Potential, CEENE: Cumulative Exergy Extraction from the Natural Environment, CExD: Cumulative Exergy Demand, ELU: environmental load unit, ORI: Ore Requirement Indicator, SED: Solar Energy Demand, SOP: Surplus Ore Potential

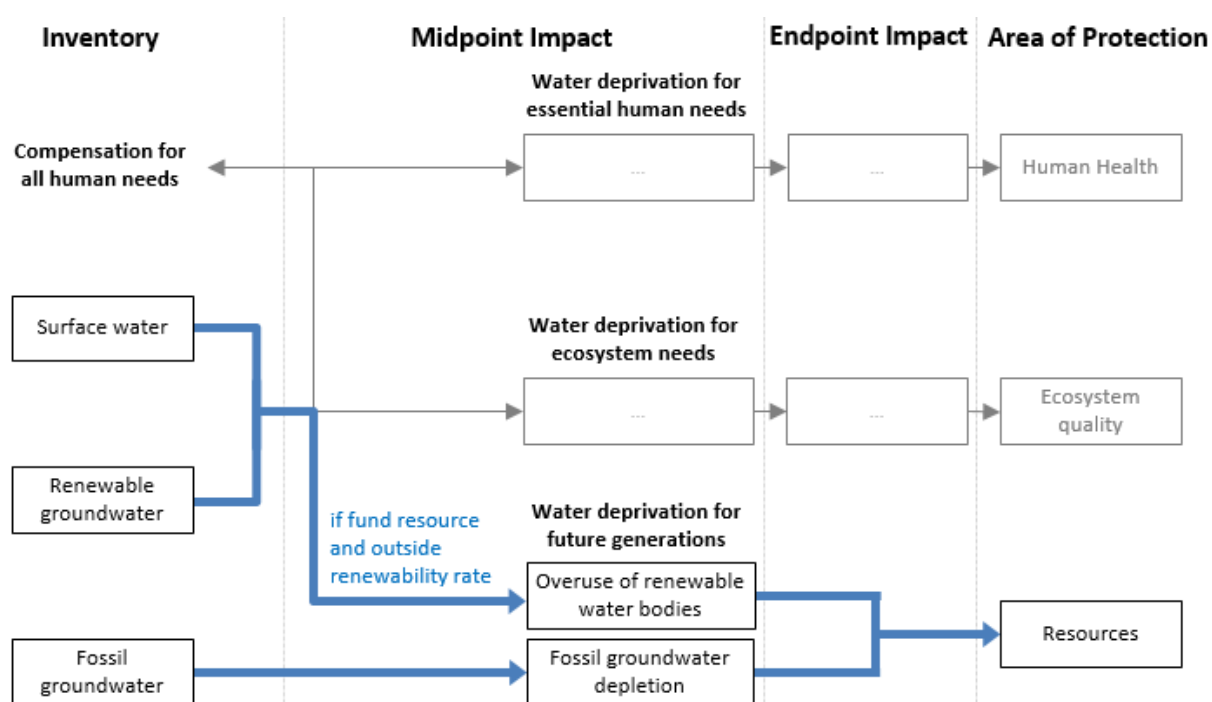


Figure S1: Cause-effect chains from the (water) inventory to the AoP according to Kounina et al. (2013)

Table S4: Methods and underlying models addressing freshwater use at midpoint and at endpoint for the AoP 'Natural Resources'; based on Kounina et al. (2013), water methods available in SimaPro 8, and literature indicated

Method	Model concept	Unit
Thermodynamic accounting/Exergy		
(Bösch et al. 2007)	Potential exergy is applied on potential energy in water used to run a hydroelectric plant The chemical exergy value for water from Szargut (2005) was attributed to freshwater. Seawater is a reference species and does not feature exergy.	MJ/m ³ MJ/MJ
Midpoint		
<i>Eco-scarcity</i> (Frischknecht and Büsser Knöpfel 2013)	Based on withdrawals to availability	unitless: m ³ (unavailable to other users)/ m ³ (used)
<i>WSI</i> (Pfister et al. 2009)	Based on withdrawals to availability (WaterGap)	unitless: m ³ (unavailable to other users)/ m ³ (used)
<i>WDI</i> (Berger et al. 2014)	Based on withdrawals to availability (WaterGap)	unitless: m ³ (unavailable to other users)/ m ³ (used)
<i>WFN Water Scarcity</i> (Hoekstra et al. 2012)	Fraction between consumed (referred to as blue water footprint) and available water (all runoff water, of which 80% is subtracted to account for environmental water needs)	unitless: m ³ (unavailable to other users)/ m ³ (used)
(Boulay et al. 2011)	Stress (α) based on CU/Q90 (WaterGap) for different sources and qualities (categories i): $\sum_i (\alpha_i \cdot V_{i,in}) - \sum_i (\alpha_i \cdot V_{i,out})$ With and without ("simplified") inclusion of quality	H ₂ O equivalent
<i>WII</i> (Bayart et al. 2014)	Quality index (not available at global level , focused on targets that should be met to ensure good ecological status of natural water bodies) combined with WSI $\sum_i (W_i \cdot Q_{W_i} \cdot WSI_i) - \sum_j (R_j \cdot Q_{R_j} \cdot WSI_j)$	m ³
(Milà i Canals et al. 2009)	Specifically for resources, ADP (CML) no regionalized characterization Surplus energy proposed with desalination as backup technology	Sb-equivalents
Regulation and Quality		
<i>FWRP & WPP</i> (Saad et al. 2013)	Based on the LANCA tool (see Table S5) FreshWater Regulation Potential (FWRP): groundwater recharge potential Water Purification Potential (WPP): physiochemical filtration potential and mechanical filtration potential Further impact in (Saad et al. 2013): Erosion Regulation Potential	mm year ⁻¹ cmol _c kg _{soil} ⁻¹ cm day ⁻¹
Endpoint resources		
<i>Water Depletion</i> (Pfister et al. 2009)	Fraction of water depleted per water amount used (WaterGap)	unitless: m ³ (depleted)/m ³ (used)
(Pfister et al. 2009)	Surplus energy with desalination as ultimate backup technology	MJ

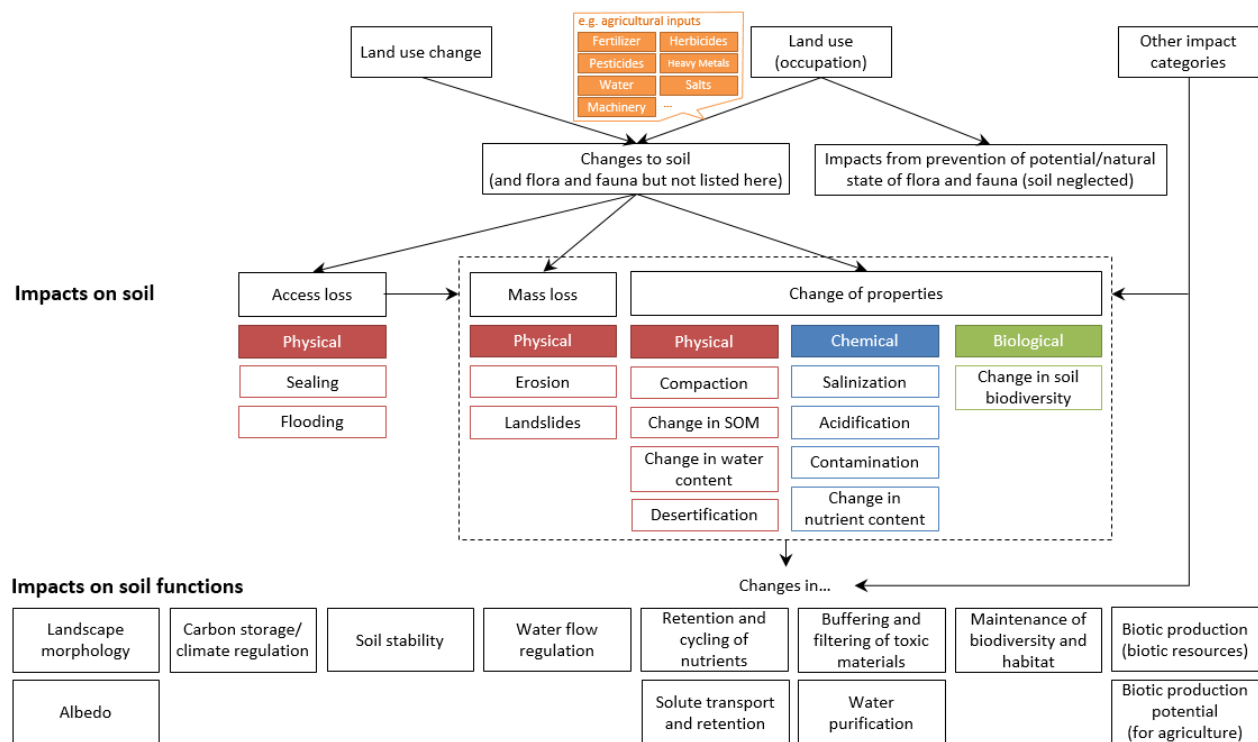


Figure S2: Impacts on soil and soil functions; own graphic based on (Cowell and Clift 2000), (Garrigues et al. 2012), (Koellner et al. 2013), (Núñez et al. 2013), and (Stoessel et al. 2016)

Table S5: Methods and underlying models addressing impacts from land use on soil as a natural resource; based on Garrigues et al. (2012), (Stoessel et al. 2016), and literature indicated

Method	Model concept	Unit
Land stress		
<i>Land Stress Index</i> (Pfister et al. 2011)	The local NPP of the natural reference vegetation is divided by a global maximal NPP as indicator for land quality No impact on soil as a resource assessed	unitless
Erosion and/or soil fertility		
<i>Achten et al. (2009)</i>	Midpoint: soil fertility (cation exchange capacity (CEC) and base saturation (BS) of the topsoil) and soil structure (SOM of the topsoil and soil compaction) Endpoint: Ecosystem Functional Quality and Ecosystem Structural Quality Practical application seems complicated	
<i>Erosion</i> (Núñez et al. 2013)	Land occupation only; surplus energy concept: combines the inventory flow soil loss (based on USLE) with the local available soil reserves and the solar energy factor of soil (around 24 MJ of solar energy is required to generate a gram of soil lost by erosion)	$\text{MJ}_{\text{se}} \text{m}^2 \text{year}^{-1}$
<i>ERP</i> (Saad et al. 2013)	Based on LANCA (see below) Erosion Regulation Potential (ERP): measured in tons of soil eroded per hectare per year, representing the capacity of a land surface to resist water erosion Further impacts in (Saad et al. 2013): Freshwater regulation potential Water purification potential	$\text{t ha}^{-1} \text{year}^{-1}$
<i>P-loss due to erosion</i> (Scherer and Pfister 2015)	Method developed for phosphorus loss; includes soil loss due to erosion based on USLE and loss of P (fertility) due to this erosion	$\text{g P m}^{-2} \text{year}^{-1}$
Loss of Soil Organic Carbon/Matter (SOC/SOM)		
<i>SOM</i> (Milà i Canals et al. 2007b)	SOM as sole indicator of soil quality Only for application in agriculture and forestry; impacts not directly involving SOM, such as acidification and salinization, are excluded. SOM does not influence erosion (probably the most important impact globally), via structural aggregate stability, enough to represent it fully	Impact: Mg C year
<i>SOC/BPP</i> (Brandão and Milà i Canals 2013)	SOC (constant conversion from/to SOM) as indicator of soil quality and its long-term ability to produce biomass (Biotic Production Potential BPP)	$\text{CF}_{\text{trans}}: \text{kg C year m}^{-2}$ $\text{CF}_{\text{occ}}: \text{kg C year m}^{-2} \text{year}^{-1}$
Compaction		
(Garrigues et al. 2013)	Loss of pore volume due to use of agricultural machines	$\text{m}^3 \text{ha}^{-1} \text{crop}^{-1}$
(Stoessel et al. 2016)	Yield loss due to compaction (caused by agricultural machines)	% yield loss
Salinization		
(Feitz and Lundie 2002)	Soil salinization from irrigation practices Based on the relationship between the sodium adsorption ratio (SAR) and the electrolyte concentration (EC) Model has to be adapted to particular sites	

Leske and Buckley (2004)	Total salinity potential for different compartments (atmosphere, surface water, natural surfaces, and agricultural surfaces) Relevant for South African conditions	
Desertification		
<i>Desertification</i> (Núñez et al. 2010)	Considers aridity, water erosion, aquifer overexploitation, and fire risk	dimensionless
Multi-criteria indicators		
(Cowell and Clift 2000)	Soil quantity and quality, many factors discussed; three impact categories: erosion, change in organic matter, compaction not operational	
LANCA (Beck et al. 2010) (Bos et al. 2016)	Based on methods of Baitz (2002) and the framework of (Milà i Canals et al. 2007a) Indicator values are calculated outside the LCA software site-specifically and then included into the LCA software in form of indicator value flows for practical reasons If specific data is not available the tool provides data on country-level	
SALCA-SQ Oberholzer et al. (2006)	Nine indicators High data requirement Calibrated for Swiss farms only	
Thermodynamic accounting/Exergy		
(Wagendorp et al. 2006)	Solar Exergy Dissipation (SED) not operational	
(Alvarenga et al. 2013)	Biomass (NPP) and area occupation	$CF_{\text{natural}}: MJ_{\text{ex}} MJ^{-1}$ $CF_{\text{human-made}}: MJ_{\text{ex}} m^{-2} year^{-1}$
(Taelman et al. 2014)	Occupation of the marine environment	$MJ_{\text{ex}} m^{-2} year^{-1}$

Abbreviations: BPP: Biotic Production Potential; NPP: Net Primary Production; occ: Occupation; SOC: Soil Organic Carbon; SOM: Soil Organic Matter; trans: transformation; USLE: Universal Soil Loss Equation

Table S6: Methods and underlying models addressing biotic natural resources

Method	Model concept	Unit
Fish		
(Emanuelsson et al. 2014)	<u>Lost Potential Yield (LPY)</u> Units of lost yield per current yield from biomass stock x during year y averaged over a time period T	unitless (yield/yield)
	<u>Overfishing through fishing mortality (OF)</u> Ratio of actual yield (fishing mortality) to maximum sustainable yield minus 1	unitless (kg/kg)
	<u>Overfishedness of Biomass (OB)</u> Ratio of optimal biomass size (size of the spawning stock at maximum sustainable yield) to actual biomass size minus 1.	unitless (size/size)
(Langlois et al. 2014)	<u>Species level</u> Ratio of fished mass to maximum sustainable yield (per year) resulting in a maximum potential regeneration time in years (extended formula for overexploitation)	years
	<u>Ecosystem level</u> Ratio of extracted organic matter (C) to Net Primary Productivity in tonnes of organic C (per year) resulting in the time in years required to regenerate the amount of biomass removed from the sea at the ecosystem level	years
HANPP		
(Alvarenga et al. 2015)	<u>Midpoint</u> $CF = HANPP_{LUC} = \text{Natural potential NPP} - \text{Actual NPP}$	$\text{kg dry matter m}^{-2} \text{ year}^{-1}$
	<u>Endpoint</u> Backup technology of producing seaweed at 14.50 \$ per kg dry matter: $CF = HANPP_{LUC} \times 14.50$	$\text{\$ m}^{-2} \text{ year}^{-1}$

Literature

- Alvarenga RAF, Dewulf J, Van Langenhove H, Huijbregts MAJ (2013) Exergy-based accounting for land as a natural resource in life cycle assessment. *Int J Life Cycle Assess* 18:939–947. doi: 10.1007/s11367-013-0555-7
- Alvarenga RAF, Erb K-H, Haberl H, et al (2015) Global land use impacts on biomass production—a spatial-differentiated resource-related life cycle impact assessment method. *Int J Life Cycle Assess* 440–450. doi: 10.1007/s11367-014-0843-x
- Bayart J-B, Worbe S, Grimaud J, Aoustin E (2014) The Water Impact Index: A simplified single-indicator approach for water footprinting. *Int J Life Cycle Assess* 19:1336–1344. doi: 10.1007/s11367-014-0732-3
- Beck T, Bos U, Wittstock B, et al (2010) Land Use Indicator Value Calculation in Life Cycle Assessment – Method Report.
- Berger M, van der Ent R, Eisner S, et al (2014) Water accounting and vulnerability evaluation (WAVE): Considering atmospheric evaporation recycling and the risk of freshwater depletion in water footprinting. *Environ Sci Technol* 48:4521–4528. doi: 10.1021/es404994t
- Bos U, Horn R, Beck T (2016) LANCA® Characterization Factors for Life Cycle Impact Assessment.
- Bösch ME, Hellweg S, Huijbregts MAJ, Frischknecht R (2007) Applying cumulative exergy demand (CExD) indicators to the ecoinvent database. *Int J Life Cycle Assess* 12:181–190. doi: 10.1007/s11367-006-0282-4
- Boulay A-M, Bulle C, Bayart J-B, et al (2011) Regional characterization of freshwater use in LCA: Modeling direct impacts on human health. *Environ Sci Technol* 45:8948–8957. doi: 10.1021/es1030883
- Brandão M, Milà i Canals L (2013) Global characterisation factors to assess land use impacts on biotic production. *Int J Life Cycle Assess* 18:1243–1252. doi: 10.1007/s11367-012-0381-3
- Cowell SJ, Clift R (2000) A methodology for assessing soil quantity and quality in life cycle assessment. *J Clean Prod* 8:321–331. doi: 10.1016/S0959-6526(00)00023-8
- Drielsma JA, Russell-Vaccari AJ, Drnek T, et al (2016) Mineral resources in life cycle impact assessment—defining the path forward. *Int J Life Cycle Assess* 21:85–105. doi: 10.1007/s11367-015-0991-7
- EC-JRC (2012) Characterisation factors of the ILCD Recommended Life Cycle Impact Assessment methods: database and supporting information.
- Emanuelsson A, Ziegler F, Pihl L, et al (2014) Accounting for overfishing in life cycle assessment: New impact categories for biotic resource use. *Int J Life Cycle Assess* 19:1156–1168. doi: 10.1007/s11367-013-0684-z
- Feitz AJ, Lundie S (2002) Soil Salinisation: A Local Life Cycle Assessment Impact Category. *Int J LCA* 7:244–249.
- Frischknecht R, Büsser Knöpfel S (2013) Swiss Eco-Factors 2013 according to the Ecological Scarcity Method. Methodological fundamentals and their application in Switzerland. 254.
- Garrigues E, Corson MS, Angers D a, et al (2013) Development of a soil compaction indicator in life cycle assessment. *Int J Life Cycle Assess* 18:1316–1324. doi: DOI 10.1007/s11367-013-0586-0
- Garrigues E, Corson MS, Angers D a., et al (2012) Soil quality in Life Cycle Assessment: Towards development of an indicator. *Ecol Indic* 18:434–442. doi: 10.1016/j.ecolind.2011.12.014
- Goedkoop M, Heijungs R, de Schryver A, et al (2013) ReCiPe 2008. A LCIA method which comprises harmonised category indicators at the midpoint and the endpoint level. Characterisation. Den Haag
- Goedkoop M, Spriensma R (2001) The Eco-indicator 99 A damage oriented method for Life Cycle Impact Assessment - Methodology Report.
- Guinée JB, Heijungs R (1995) A proposal for the definition of resource equivalency factors for use in product life-cycle assessment. *Environ Toxicol Chem* 14:917–925. doi: 10.1002/etc.5620140525
- Hoekstra AY, Mekonnen MM, Chapagain AK, et al (2012) Global monthly water scarcity: Blue water footprints versus blue water availability. *PLoS One*. doi: 10.1371/journal.pone.0032688
- Impact World+ (2016) Methodology and Models. <http://www.impactworldplus.org/en/methodology.php>; accessed 15.02.2016. Accessed 15 Feb 2016
- Itsubo N, Inaba A (2012) Lime2. JLCA Newsl Life-Cycle Assess Soc Japan 16.
- Itsubo N, Sakagami M, Washida T, et al (2004) Weighting across safeguard subjects for LCIA through the application of conjoint analysis. *Int J Life Cycle Assess* 9:196–205. doi: 10.1007/BF02994194
- Jolliet O, Margni M, Charles R, et al (2003) IMPACT 2002 + : A New Life Cycle Impact Assessment Methodology. *Int J Life Cycle Assess* 8:324–330. doi: 10.1007/BF02978505
- Koellner T, de Baan L, Beck T, et al (2013) UNEP-SETAC guideline on global land use impact assessment on biodiversity and ecosystem services in LCA. *Int J Life Cycle Assess* 18:1185–1187. doi: 10.1007/s11367-013-0580-6

- Kounina A, Margni M, Bayart J-B, et al (2013) Review of methods addressing freshwater use in life cycle inventory and impact assessment. *Int J Life Cycle Assess* 18:707–721. doi: 10.1007/s11367-012-0519-3
- Langlois J, Fréon P, Delgenes J-P, et al (2014) New methods for impact assessment of biotic-resource depletion in life cycle assessment of fisheries: theory and application. *J Clean Prod* 73:63–71. doi: 10.1016/j.jclepro.2014.01.087
- Milà i Canals L, Bauer C, Depestele J, et al (2007a) Key elements in a framework for land use impact assessment within LCA. *Int J Life Cycle Assess* 12:5–15. doi: 10.1065/lca2006.05.250
- Milà i Canals L, Chenoweth J, Chapagain A, et al (2009) Assessing freshwater use impacts in LCA: Part I - Inventory modelling and characterisation factors for the main impact pathways. *Int J Life Cycle Assess* 14:28–42. doi: 10.1007/s11367-008-0030-z
- Milà i Canals L, Romanyà J, Cowell SJ (2007b) Method for assessing impacts on life support functions (LSF) related to the use of “fertile land” in Life Cycle Assessment (LCA). *J Clean Prod* 15:1426–1440. doi: 10.1016/j.jclepro.2006.05.005
- Núñez M, Antón A, Muñoz P, Rieradevall J (2013) Inclusion of soil erosion impacts in life cycle assessment on a global scale: application to energy crops in Spain. *Int J Life Cycle Assess* 18:755–767. doi: 10.1007/s11367-012-0525-5
- Núñez M, Civit B, Muñoz P, et al (2010) Assessing potential desertification environmental impact in life cycle assessment. *Int J Life Cycle Assess* 15:67–78. doi: 10.1007/s11367-009-0126-0
- Pfister S, Bayer P, Koehler A, Hellweg S (2011) Environmental impacts of water use in global crop production: hotspots and trade-offs with land use. *Environ Sci Technol* 45:5761–5768. doi: 10.1021/es1041755
- Pfister S, Koehler A, Hellweg S (2009) Assessing the Environmental Impact of Freshwater Consumption in Life Cycle Assessment. *Environ Sci Technol* 43:4098–4104. doi: 10.1021/es802423e
- Rørbech JT, Vadenbo C, Hellweg S, Astrup TF (2014) Impact assessment of abiotic resources in LCA : Quantitative comparison of selected characterization models. *Environ Sci Technol* 48:11072–11081.
- Saad R, Koellner T, Margni M (2013) Land use impacts on freshwater regulation, erosion regulation, and water purification: A spatial approach for a global scale level. *Int J Life Cycle Assess* 18:1253–1264. doi: 10.1007/s11367-013-0577-1
- Scherer L, Pfister S (2015) Modelling spatially explicit impacts from phosphorus emissions in agriculture. *Int J Life Cycle Assess* 20:785–795. doi: 10.1007/s11367-015-0880-0
- Schneider L, Berger M, Finkbeiner M (2015) Abiotic resource depletion in LCA - background and update of the anthropogenic stock extended abiotic depletion potential (AADP) model. *Int J Life Cycle Assess* 20:709–721. doi: 10.1007/s11367-015-0864-0
- Steen B (1999) A systematic approach to environmental priority strategies in product development.
- Stoessel F, Bachmann D, Meier MS, et al (2016) Assessing the environmental impacts of soil compaction in LCA. (in preparation).
- Taelman SE, De Meester S, Schaubroeck T, et al (2014) Accounting for the occupation of the marine environment as a natural resource in life cycle assessment: An exergy based approach. *Resour Conserv Recycl* 91:1–10. doi: 10.1016/j.resconrec.2014.07.009
- Wagendorp T, Gulink H, Coppin P, Muys B (2006) Land use impact evaluation in life cycle assessment based on ecosystem thermodynamics. *Energy* 31:112–125. doi: 10.1016/j.energy.2005.01.002
- Weidema BP, Hauschild MZ, Joliet O (2007) Preparing characterisation methods for endpoint impact assessment. available from lca-net.com/files/Stepwise2006v1.5.3.zip