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Stochastic Ex-ante LCA Under Multidimensional Uncertainty: Anticipating the Production of Undiscovered Microalgal Compounds in Europe

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KEYWORDS

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ABSTRACT

Due to their biodiversity, microalgae represent a promising source of high-value compounds that bioprospecting is aiming to reveal. Performing an ex-ante Life Cycle Assessment (LCA) to anticipate and potentially minimize the environmental burden associated with the European production of a bioprospected microalgal compound is subject to substantial and multi-factorial uncertainty, as the compound remains undiscovered. Given that any microalgal strain could potentially host the compound of interest, the ex-ante LCA should consider this bioprospecting uncertainty together with the uncertainty on the technology and the production mix.

Using a parameterized cultivation simulation and consequential LCA model, and an extensive stochastic pseudo Monte Carlo approach, we define and propagate techno-operational, bioprospecting, and production mix uncertainties for a microalgal compound being currently bioprospected in Europe. We perform global sensitivity analysis using different sampling strategies to identify the main contributors to the total output variance. Overall, the uncertainty propagation allowed us to define and analyze the probabilistic scope for the potential environmental impacts in the emerging production of high-value microalgal compounds in Europe, based on current knowledge. These findings can support policy-making as well as actors in the microalgal sector towards technological paths with lower environmental impact.

SYNOPSIS

We anticipate the environmental impacts associated with the European production of a currently bioprospected microalgal compound via a stochastic, ex-ante, and consequential LCA based on microalgal cultivation simulations.

TEXT

1. Introduction

The biological diversity of microalgae makes them a promising biological group for biotechnological applications such as the production of organic products of high commercial value. Among these products, while microalgae-based 3rd generation biofuels have so far failed to compete economically with fossil fuels¹, a few high-value microalgal compounds are already commercialized^{2,3}. Recent discoveries from bioprospecting, i.e. the search for compounds and properties within the biodiversity that could be valuable for human activities, range from antimicrobial to antitumoral lipids, proteins, and carbohydrates in microalgal strains^{4–10}. This suggests that the European microalgae sector might develop substantially in the near future. Anticipating the environmental consequences of such a development is crucial as early-stage emerging systems are characterized by a high design freedom¹¹ implying potential detrimental scenarios and because sustainable development processes “*are timely, anticipatory, integrative, flexible and action focused*”¹². Life Cycle Assessment (LCA) constitutes a robust holistic tool to quantitatively anticipate such impacts on a systemic level and offers flexibility and parametrization to project different scenarios.

LCA studies on the environmental impacts of microalgae production address primarily bioenergy-oriented microalgae growth in Open Pond Raceways^{13–18}, while high-value compounds production would likely require stable and contamination-free photobioreactors (PBR). When reviewing eighteen existing LCA studies of microalgae^{19–27}, we found that they consider only seven well-studied strains. This is likely because the microalgal sector has historically focused on bioenergy applications and therefore primary data and assumptions on yields and operating

conditions exist only for a restricted set of strains that are fit for bioenergy and have already been already cultivated industrially (lipid-rich, robust etc.). This data alone cannot support a comprehensive assessment of the future consequences of the microalgae sector's development which could potentially take as many paths as there are microalgal strains and promising compounds to be discovered via bioprospecting. Furthermore, while previous bioengineering studies^{28,29} model the potential for microalgal productivities in different PBRs and locations, previous LCA studies only assess productions localized in one or a few sites. Yet, it is reasonable to assume that an increase in demand for new microalgal compounds would, in the long run, result in the supply of microalgae from several different locations with distinct productivities, thermoregulation needs, and electricity mixes. In fact, 447 microalgae or cyanobacteria farms are already active in 23 European countries³⁰.

Forecasting the environmental consequences of future developments in the European microalgae sector therefore requires extending the scope of the previous assessments to include many possible scenarios. This immense number of possibilities is inherent to bioprospecting, which implies that the desired property of a bioprospected compound is known (for instance an anti-inflammatory compound), but not the organism which will produce it. Besides the case of bioprospecting, it is common³¹ to consider a large number of scenarios in ex-ante LCA as it aims at anticipating the environmental impact of emerging technologies before these are actually implemented at industrial scale³². A major challenge in ex-ante LCA is thus the need to generate Life Cycle Inventories (LCI) for future production systems which currently exist only at a low level of technological and market maturity³³ such as pilot scale applications. In a recent review of common practices in ex-ante LCA³¹, one third of the 18 included studies resorted to process simulation to obtain LCIs at an industrial scale, mainly using technology-specific simulation software. In the case of microalgal

compounds which are still being bioprospected, the upscaling anticipation can be done using parameterized physical and biological models to simulate photobioreactors and strains in different conditions. In fact, it is common practice^{18,21,23–25,27,34} to simulate inventories for LCAs in the microalgal sector and recent studies show advanced models taking various biophysical phenomena into account³⁵. Microalgae LCA models with extensive parameterization can be found associated with some forms of uncertainty and sensitivity analysis via stochastic sampling^{14,34,36}.

These parametrized models, however, cannot cover the scope of possibilities associated with a microalgal compound that is not found yet, and whose production upscaling and development in a European production market are indeterminate. Yet, the consequences of an increase in demand for high value microalgal compounds must be assessed early on to inform both policy makers in the sustainability domain and the microalgal sector. The ex-ante assessment should put these stakeholders in a better position to evaluate the likelihood of microalgal high-value compounds' environmental superiority over alternatives as well as to identify the optimal scenarios and try to aim for them.

To do so, the present work aims at investigating the impacts of an increase in demand for a microalgal compound with a desired property that is currently being bioprospected in Europe (a bioprospected microalgal compound). The uncertainty space associated with this technological development is shaped by several forms of uncertainty that can either be epistemic i.e. due to a lack of knowledge, or aleatory because stemming from inherently random processes³⁷. We build on previous work³⁸ developing a parameterized model to simulate the microalgae cultivation technology from a life cycle perspective. The “techno operational uncertainty” addressed in this previous work is due to the impossibility of accurately predicting the productivity of a new specific strain in a specific photobioreactor and location. To tackle our research question, we here need to

extend the parameterized model and the scope of previous LCAs in the microalgal sector to include other sources of uncertainty, namely the lack of knowledge about the exact nature of the strain and the bioprospected compound, the associated PBR, and the geographic developments of the market. We anticipate that despite these large uncertainties, the stochastic propagations will provide insightful density-based representations of the environmental consequences of an increase in demand for yet undiscovered microalgal compounds. This work also contributes to the developments and discussions around uncertainty, global sensitivity analysis and the associated terminologies in the field of ex-ante LCA.

2. Methods

The approach consists in applying stochastically generated samples to our previously developed parametrized LCA model³⁸ on which minor changes were made (cf. 2.1) to assess the impacts for an increase in demand and production of 1 kg of bioprospected compound in Europe. The whole model is coded on Python 3.8 with the LCA package Brightway2³⁹ and is available on GitHub⁴⁰. The background database is the consequential version of Ecoinvent 3.6. The impact assessment categories are Global Warming with a 100-year time horizon (GW100), Freshwater Eutrophication (FE), Water Depletion (WD) and Terrestrial Ecotoxicity (TETinf) from ReCiPe Midpoint (H) V1.13.

2.1 Deterministic model and LCA framework

The functional unit is 1 kg of bioprospected compound. Consequential LCA modeling was chosen as, by looking at the future effects of decisions, and including only activities and technologies expected to be able to respond to future changes in demand, the consequential approach is prospective in nature and therefore well-suited to the assessment of emerging technologies.

Moreover, the consequential approach reduces the number of normative assumptions needed, and since it is highly speculative to anticipate normative preferences in the future, this can be argued to be an advantage when performing ex-ante LCA. The foreground product system includes the cultivation of microalgae in an outdoor vertical tubular PBR with the associated energy consumptions for water pumping, mixing, thermoregulation by a heat pump and centrifugation for biomass harvesting. The product system also comprises nutrients, CO₂, water, and glass consumptions. The extraction of the compound is not modeled but cell disruption is accounted for as it likely constitutes the first step of the post-harvest processing, regardless of the biochemical class of the compound (protein, lipid, carbohydrate)^{41,42}. Drying is modeled for the whole biomass so that the co-produced biomass (dependent co-product) is ready to substitute functionally equivalent products already on the market. The cultivation technology and the foreground product system are described in detail in SI I.2 and in our previous work and model³⁸, for which a few modifications were made to the product system. First, thermoregulation of the PBR was assumed to be provided by a reversible heat pump⁴³ with a Coefficient of Performance (COP) of 3, which we chose as an average value over locations and seasons, instead of electric heating and a fluid thermal exchanger. Additionally, considering multiple potential strains with different characteristics made it necessary to model two additional possible substitution routes for the co-produced biomass (cf. SI I.3). This biomass can now enter the animal feed energy and feed protein markets⁴⁴, or be directly incorporated in fish feed after modification of the reference fish feed composition as previously modeled³⁸. As a third possible substitution route, the co-produced biomass can be digested in an anaerobic digester for biogas production and substitution on the biogas marginal market based on a functional unit of 1MJ of heating capacity. The biogas yields depend on the composition of the microalgal strains (cf. SI I.3.2.2).

The life cycle inventory results from a simulation of microalgal cultivation assumed from April to September within Europe. This simulation and the LCI result from the interaction of 27 parameters characterizing the strain and its compound, 22 techno-operational parameters characterizing the cultivation technology, the PBR geometry and setup, 3 geographic parameters defining the location, and 6 physical parameters. The cultivation simulation uses climatic data from the Photovoltaic Geographic Information System (PVGIS)⁴⁵. The simulation, its modules, the parameters, and all the equations are detailed in our previous work³⁸ and in SI I for the additions and modifications.

2.2 Types of uncertainty and assumptions

2.2.1 Bioprospecting uncertainty

The “bioprospecting uncertainty” is caused by the lack of knowledge about the strain and the compound that will be found to feature the desired property (for instance anti-inflammatory) and successfully upscaled after bioprospecting. The bioprospecting uncertainty is epistemic³⁷ in the first place as it stems from the lack of knowledge about which kinds of microalgal strains or compounds classes (lipid, carbohydrate, protein) are more likely to possess the desired property. , We have assumed that all strains and types of compounds have equal probabilities to possess the desired property, as we currently lack arguments and data to hypothesize potential correlations between biological traits and desired properties. Therefore, the bioprospecting process was modeled as a random draw within microalgal biodiversity, whose result is subject to aleatory uncertainty. This is conceptually analogous to a draw in an opaque urn in which the balls would be the strains and their compounds whose proportions in the box depend on our knowledge on biodiversity(cf. 3.3).

To do so, we first use 27 parameters defining a "strain-compound pair", i.e. a specific compound hosted in a specific strain. Some parameters define biological characteristics such as biomass composition, nitrogen source, or photosynthetic efficiency. Other parameters define if the strain-compound pair requires PBR thermoregulation at night or characterize the fate of the co-produced biomass (biogas, fish feed, or animal feed) that we consider strain-specific because depending on cell wall characteristics, digestibility, toxicity etc.^{46–48} Finally, compound-specific parameters define respectively: whether the bioprospeted compound is a lipid, protein, or carbohydrate, i.e. the biochemical class of the compound; and the mass fraction of compound in such class, for example measured as the mass of compound per total mass of proteins in the strain if the compound is a protein. The values of these compound-specific parameters are also characterized by uncertainty which means that the same functional unit of 1 kg of bioprospeted compound can be provided by different reference flows of cultivated microalgal biomass.

We modeled the random draw within the microalgal biodiversity by sampling random values for the parameters for which variation ranges are reported in the literature for the microalgal biodiversity (23 out of 27 parameters, cf. SI II.1). For 17 parameters out of 23, we modeled a uniform distribution within the range in lack of further arguments to use other distributions.

2.2.2 *Techno-operational uncertainty*

This epistemic uncertainty was specifically addressed in our previous work³⁸ and is due to the unpredictable behavior and growth of a strain defined by its biological parameters in a specific vertical tubular PBR and location. More precisely, we used the model from Williams and Laurens (2010)²⁸ which estimates a maximum areal yield ($g_{dw} \cdot m^{-2} \cdot d^{-1}$) depending on the ground horizontal irradiance ($kJ \cdot m^{-2} \cdot d^{-1}$) and the strain-specific theoretical energetic yield ($g_{dw} \cdot kJ^{-1}$). Since

Williams and Laurens (2010)²⁸ observe that real cultivations in PBR would often reach 30% of this maximum yield, in this study the techno-operational uncertainty refers to the lack of knowledge about which PBR geometry and operational set-up will enable the modeled strain to reach this percentage. The techno-operational uncertainty was addressed by simulating random values of geometrical (e.g. tube diameter and distance between tubes) and operational (flow rate, biomass concentration) parameters defining the PBR for a strain and location. The values were sampled within ranges reported in the literature for different vertical tubular PBRs with different strains in various locations (cf. SI I.1). The sampled combinations were all assumed to have equal chances of enabling the strain to reach 30% of its maximum yield.

In addition to the geometrical and operational parameters, the uncertain vertical distance between the water source (river or well) and the PBR, and the wind-dependent convective exchange coefficient ruling the thermal exchange between the PBR and the surrounding air were included in the techno-operational uncertainty.

In total, 7 techno-operational parameters are therefore considered uncertain.

2.2.3 Geographic locations and uncertainty of production mix

Similarly to any agricultural crop for which an increase in demand in Europe will be answered by different producers in distinct locations (a mix), an increase in demand for the bioprospected compound will be met by distinct microalgae plants in Europe belonging to a “compound production mix”. To anticipate the development of this compound production mix in Europe, we first assumed that the production would take place in the 10 countries which have the highest potential for microalgal biomass production as identified by Skarka⁴⁹: Spain (ES), Sweden (SE), Italy (IT), Portugal (PT), United Kingdom (UK), France (FR), Greece (EL), Cyprus (CY), Ireland

(IE) and Germany (DE). The author identified these countries as the best combinations of temperature, solar irradiance, and available land after exclusion of urban, mountainous, and protected areas. To account for indeterminacy regarding the plants' locations, a grid was first generated with random locations drawn every 2° of latitude in each country (28 locations in total). Mono-dimensional sampling was performed in each of this location, while multi-dimensional sampling only could add production mix uncertainty by sampling production mixes within the grid (cf. 2.3). To do so, random combinations of locations were selected within the grid in three different scenarios regarding the spread of the mix over Europe. Thus, in these scenarios, 5, 15, or 25 locations out of 28 were assumed to answer to the increase in demand.

In consequential LCA, the identification of the marginal suppliers and their shares in the mix are based on the study of market trends, that are shaped by both political and economic factors^{50,51}. In our case, due to the lack of data on market performance we used areal productivity as a proxy and assumed that the plants in locations with higher areal productivities will have a higher chance to be part of the compound production mix. We then assumed equal production shares within this mix, once its locations have been determined: each plant produces 200 g of compound in a production mix for 1 kg if the mix contains 5 plants in different locations. A country-specific electricity mix was used when simulating each plant.

2.2.4 Independence assumptions

The model was designed so that the sampled parameters are independent. For instance, our knowledge does not indicate if a specific strain has a higher chance to grow on nitrate instead of ammonium if it has a high optimal temperature and a short cell diameter. These three parameters were therefore sampled independently. Similarly, a cell diameter does not indicate if the strain has

a higher chance to grow at the expected yield in a certain tube diameter, all other things being considered. Indeed, we did not model direct dependence between techno-operational parameters and productivity for a strain but instead treated this complexity as a part of the uncertainty (cf. 2.2.2) that could be reduced with more knowledge on the dependencies observed across species and locations. On the contrary, resorting to a parameter defining the content of the bioprospected compound in the biomass which would be independent of the biomass composition would lead to unrealistic scenarios. Thus, the biomass composition was first generated by randomly sampling a lipid content of the ash-free dry biomass and an ash content, from which the rest of the composition is calculated²⁸, and the bioprospected compound constitutes a random fraction of a random biochemical class (lipid, protein, carbohydrate).

2.3 Sampling strategies for uncertainty propagation and sensitivity analysis

As illustrated in Figure 1, we implemented two random sampling strategies to propagate the uncertainty and perform sensitivity analysis.

The first strategy named “mono-dimensional sampling” mixes all types of uncertainties previously described (bioprospecting, techno-operational, cf. 2.2) by applying Sobol sampling to all parameters. With this sampling strategy, we obtained for each location of the grid 114688 random combinations of the 23 biological and 7 uncertain techno-operational parameters. The global sensitivity analysis associated with mono-dimensional sampling allows ranking all parameters based on their influence on the dispersion of the impact scores in one location.

To be able to simulate the development of European production mixes producing the same strain-compound pairs in different locations, the second sampling strategy, namely “multi-dimensional sampling”, differentiates between uncertainty types. We first generated a Sobol

sample to obtain a set of 5376 strain-compound pairs that could potentially display the desired property (bioprospecting uncertainty). The cultivation and production of each strain-compound pair was then simulated for each of the 28 locations of the grid, in 150 random PBR geometries and operational setups generated with Monte Carlo sampling on the 7 uncertain techno-operational parameters (techno-operational uncertainty). As the same strain-compound pairs were simulated in all locations of the grid, we could generate production mixes for all pairs by composing 400 random combinations of locations on the grid (Production mix uncertainty). The probabilities for each location to be part of the marginal mix were weighted with the areal productivities (cf. 2.2.3). Unlike mono-dimensional sampling, multi-dimensional sampling allows assessing the uncertainty associated with the environmental impact caused by an unknown European production mix producing the same strain-compound pair, after aggregating techno-operational uncertainty (cf. Figure 1). This sampling strategy also allows studying how sensitive is the impact of the entire European production mix (composed of the 28 locations of the grid) to the bioprospecting uncertainty, i.e. to the parameters defining the strain-compound pair produced in this mix.

The global sensitivity analyses were performed with the Sobol methods from the python Salib package⁵².

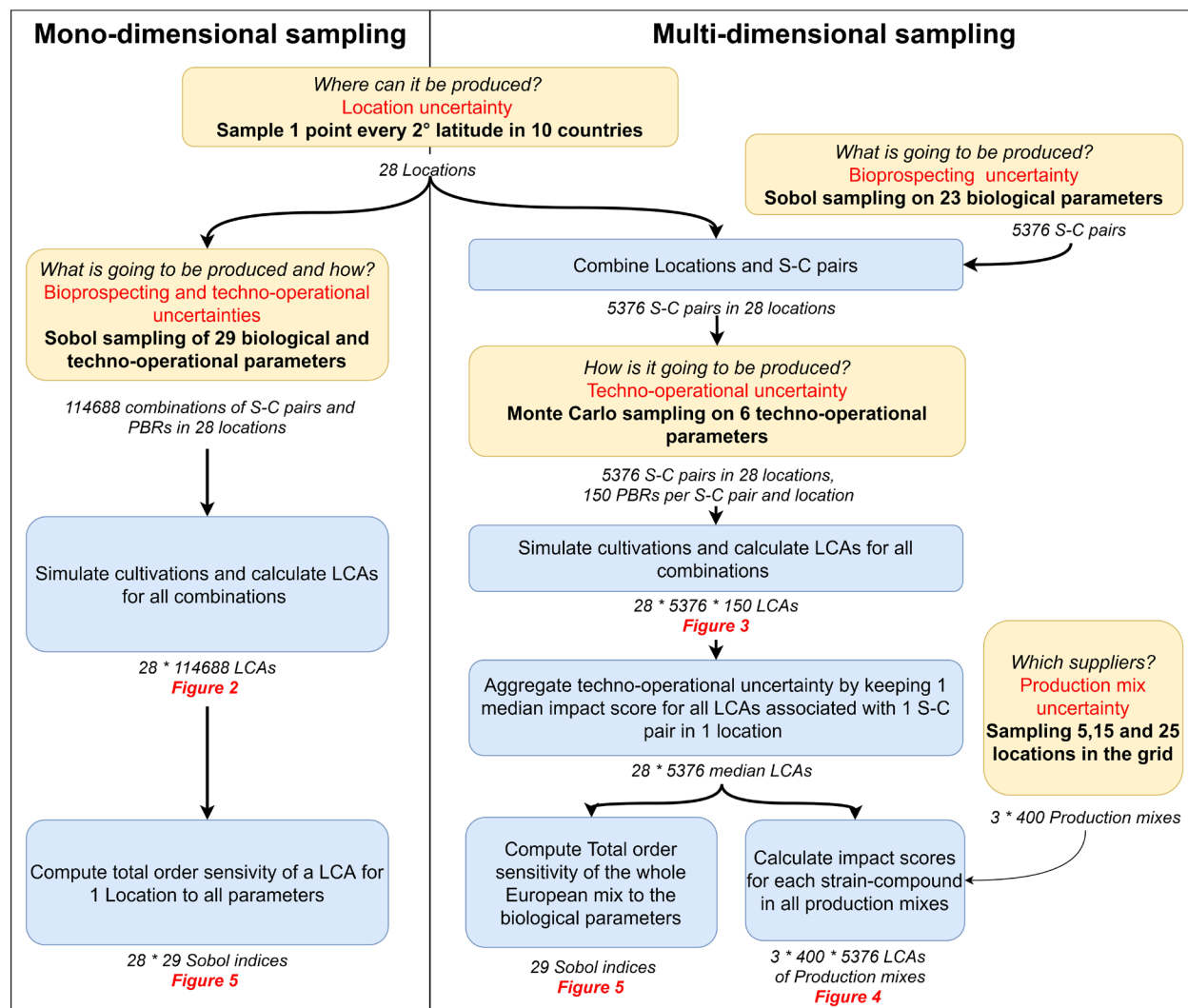


Figure 1: Visualization of the sampling strategies and the associated calculations. Acronyms: S-C= Strain-Compound. Blue and orange boxes respectively indicate a calculation step and a sampling step. We indicated the figures of the article under the results that they display.

3. Results and discussion

3.1 Uncertainty propagation with mono-dimensional sampling

The first notable observation is that the propagation of the uncertainty resulted in a very wide range of impacts scores, with for instance the global warming impact (GW100) per kg of bioprospeted compound ranges from -100 to +89000 kg CO₂-eq / kg compound with mono-dimensional sampling (cf. SI I.3). This very large dispersion is mainly due to the uncertain content of bioprospeted compound in the biomass, which is part of the bioprospecting uncertainty. The bioprospeted compound content is a secondary parameter as it results from the interaction of four uncertain primary parameters (cf. 2.2.4). The propagation of the uncertainty for these four parameters resulted in a target compound content ranging from 0.001 to 0.6 g·g_{driedbiomass}⁻¹ (cf. SI II.2). The highest impact score in the range corresponds to extremely unfavorable conditions: a very low content of bioprospeted compound in the biomass (0.002 g·g_{driedbiomass}⁻¹), coupled with a northern location (Sweden), a high strain-specific thermal range and a large PBR volume involving low volumetric productivity and substantial PBR heating requirements. GW, Freshwater Eutrophication (FE), and Water Depletion (WD) impact scores all increase when the target compound content decreases because of a need to produce more biomass to provide the same functional unit, but a lower target compound content induces a higher Terrestrial Ecotoxicity (TET) impact on average (Figure 2). The observed overall trend in Figure 2 however hides the specific trends for the three equiprobable substitution routes. Indeed, the overall trend is only due to the scenario in which the coproduced biomass substitutes fish feed. Despite constituting one third of all the simulations, the strong positive slope observed for this scenario outweighs the slightly negative slope for the two other scenarios, which results in the trend observed in Figure 2

317 (cf. Figure S13 in SI I. 4.2.4.1). Fish feed substitution almost always implies negative TET scores
318 (positive impact on the environment), which is partly the case for animal feed and never happened
319 for biogas substitution. The environmental superiority of fish feed substitution over biogas
320 production had also been found in a previous study²⁴.

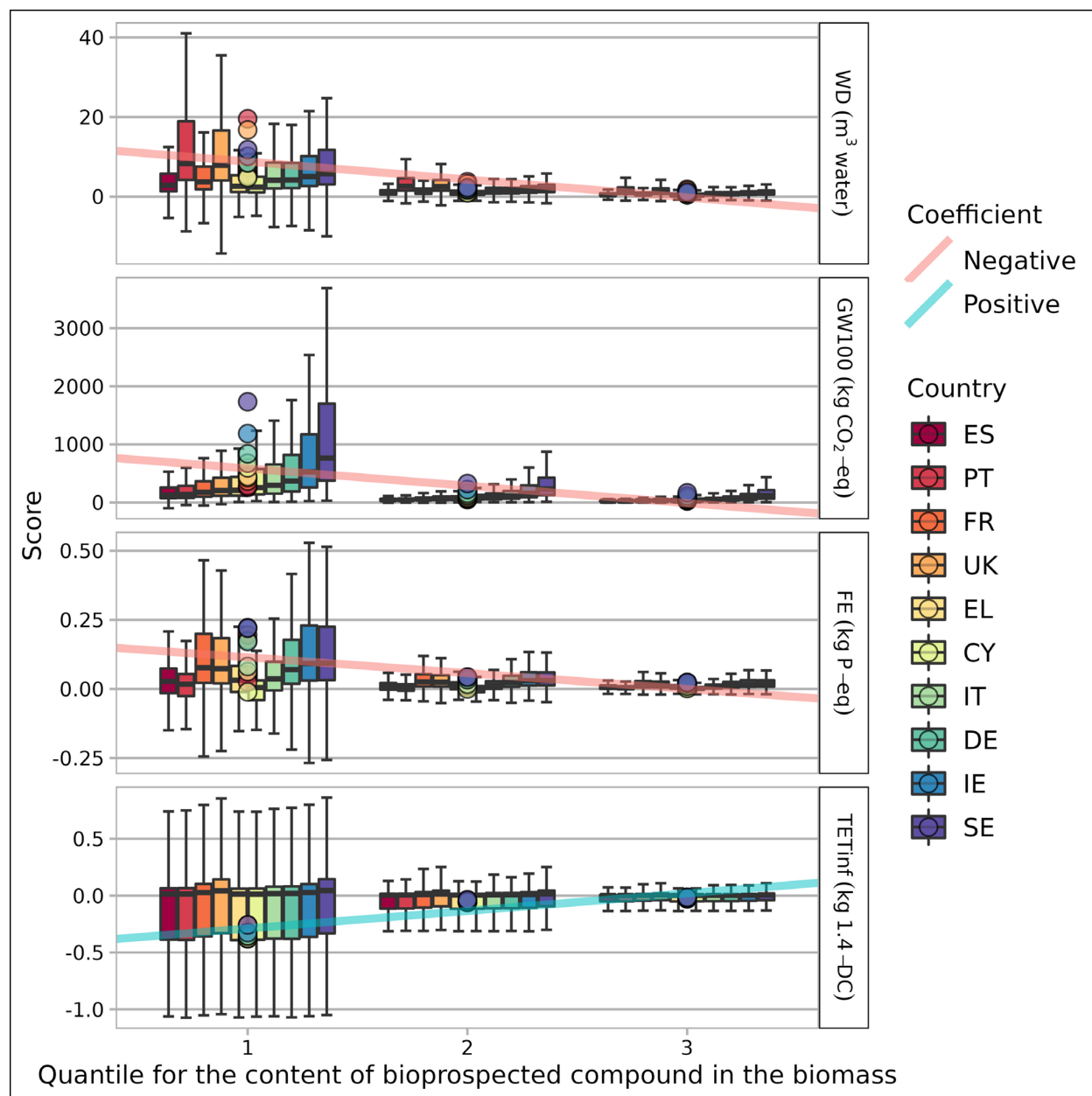


Figure 2: Boxplot of the environmental scores obtained in mono-dimensional sampling. The 211 264 LCA scores resulting from mono-dimensional sampling are divided into 3 quantiles applying to the dispersion of the uncertain bioprospeted compound content in the biomass. Boundaries of the quantiles ($g_{\text{bioprospeted compound}} \cdot g_{\text{driedbiomass}}^{-1}$): 0-0.001, 0.001-0.090, 0.090-0.185, 0.185-0.60. The dots indicate the average scores per country. The regression lines are only displayed to highlight the bioprospeted compound content influence on the scores. The equations of the lines

328 *are* $Score = coeff * Q + intercept$, with Q the quantile numbers 1,2, and 3. More detailed plots can
329 *be found in SI I.4.2.4.1. WD= Water Depletion, GW100=Global Warming with a 100-year time*
330 *horizon, FE=Freshwater Eutrophication, TETinf=Terrestrial Ecotoxicity.*

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3.2 Uncertainty propagation with multi-dimensional sampling

3.2.1 Bioprospecting and Techno-operational uncertainties

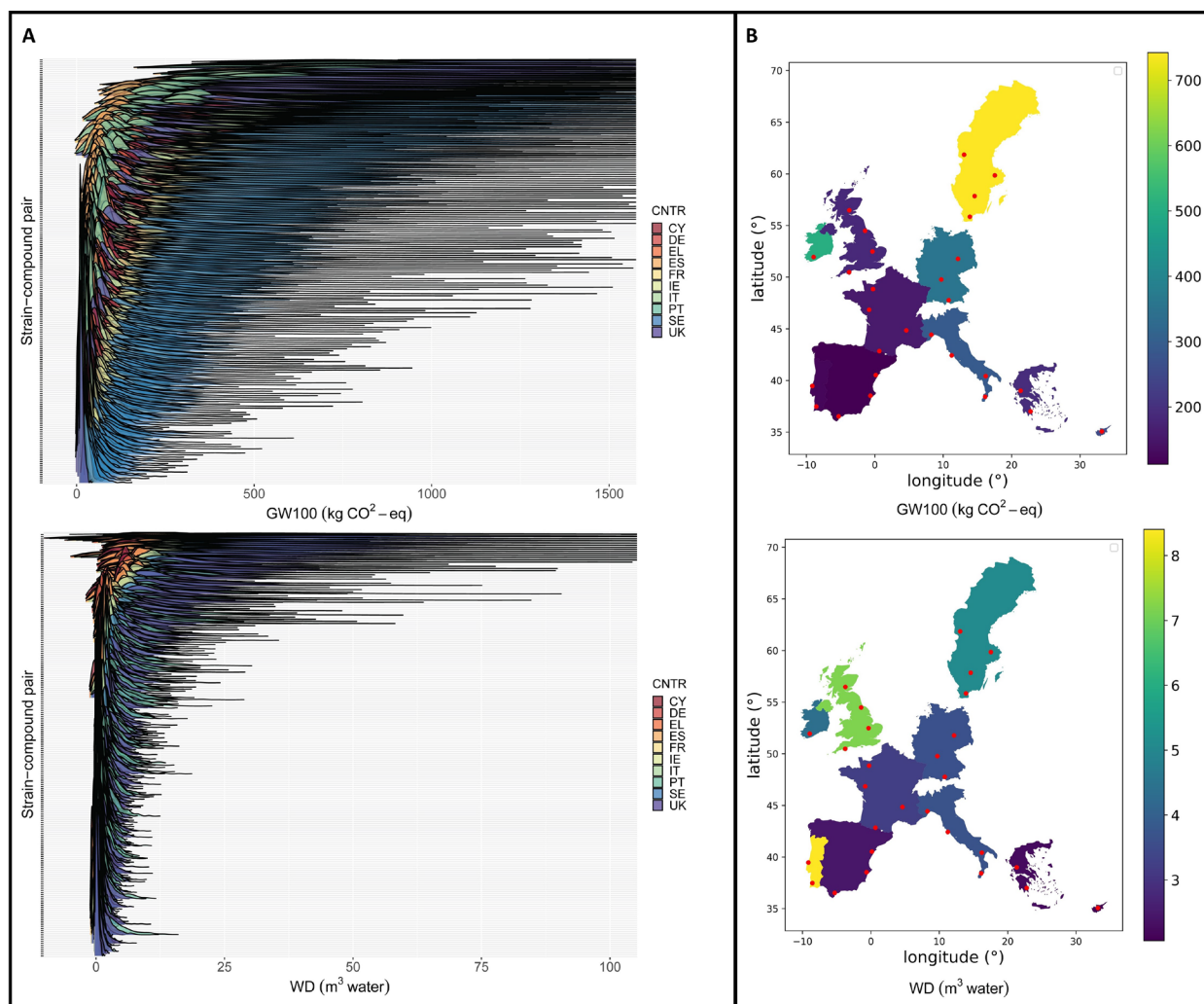


Figure 3: (A) Distinction of uncertainties with strain-compound-specific ridgelines and (B) Mapping of average impacts scores per country. (A) Each horizontal line corresponds to one strain-compound pair for which the LCA scores in different locations and stochastically generated PBRs constitute the density curves. The plots only show 200 of the 5376 randomly generated strain-compound pairs in multi-dimensional sampling. The strain-compound pairs were divided into 200 quantiles regarding the dispersion of the GW impact scores and one pair was chosen per quantile and displayed on the figures. (B) The color gradients indicate the mean impact score per

country, considering all simulations performed in each country. The red dots indicate the locations of the randomly generated grid. The figures for TET and FEP are shown in SI I.4.1.

In Figure 3, for each horizontal line representing a strain-compound pair, the density curves of the impact (horizontal axis in Figure 3) result from the propagation of the techno-operational uncertainty in the different countries. Despite a visible shift of the impact density curves across strain-compound pairs (vertical axis in Figure 3) and different countries (different colors in Figure 3), the significant overlap between strain-compound pairs indicates a large influence of techno-operational uncertainty.

It must be noted for GW, FE, and WD, that the shifts of the curves along the horizontal axis are not mere linear transposition of the density curves but are associated with a higher standard deviation of the results for strain-compound pairs with higher average impact scores (cf. SI I.4.3). In other words, the dispersion of the impact scores due to techno-operational uncertainty varies with the modeled strain-compound pair. This heteroskedastic statistical behavior stems from two mechanisms particularly visible for GW. First, the same techno-operational input uncertainty is assumed for the cultivation simulations of all strain-compound pairs, which tends to result in a constant coefficient of variation (*Standard deviation/Mean*) and therefore a linear increase of the standard deviation across the mean impact scores of different strain-compound pairs (cf. Figure S16, SI I.4.3). Second, the coefficient of variation is tendentially higher for strains with higher optimum temperatures of culture (cf. Figure S15, S16, S20, SI I.4.3), which are also associated with high impact scores on average. This phenomenon was observed in our previous work³⁸ and stems from the fact that techno-operational uncertainty propagates more when thermoregulation can become a hotspot due to the combination of strain's thermal requirements and location.

3.2.2 Variability across countries and production mix uncertainty

Despite techno-operational and bioprospecting uncertainties which cause substantial overlap between impact density curves (Figure 3.a.), the ten countries can be ranked according to the average impact score for all simulations (Figure 3.b. and also observed in Figure 2 with mono-dimensional sampling). Latitude is an important determinant of the environmental impacts (cf. Figure S5, SI I.4.2.1.1) as it affects the horizontal irradiance and therefore influences biomass productivity. Latitude also determinates the outside temperature and therefore the energy required to thermoregulate the culture, in particular heating requirements, which was highlighted as an environmental hotspot in other studies^{38,53,54}. Thus, the impact scores for all impact categories and countries tendentially increase with the strain-specific optimal temperatures T_{opt} (part of the bioprospecting uncertainty), and the regression slopes associated with $\log(\text{Impact score}) = f(T_{opt})$ are higher for northern countries (cf. SI I.4.2.1.2). Additionally, the impact scores per FU are strongly influenced by the impact profiles of the national electricity mixes (cf. SI I.4.2.2). To make an example, despite of its southern location, production in Portugal is associated with the highest WD impact score among all countries due to the high WD impact of the Portuguese electricity mix (Figure 2,3). However, it is important to note that current marginal mixes used in consequential modeling only have a limited period of validity in the near/medium term future.

Differences between the impact scores of distinct countries do not necessarily imply additional uncertainty for the impact scores associated with the production of 1 kg of bioprospected compound produced by a European production mix. Figure 4 shows how the dispersion of the impact scores for the strain-compound pairs varies across the generated production mixes. The differences observed between the distributions, which illustrate the production mix uncertainty,

shrink when more locations are considered per mix. Indeed, the more locations there are in the randomly generated mixes, the closer the latter get to a full European mix composed of the 28 locations of the grid.

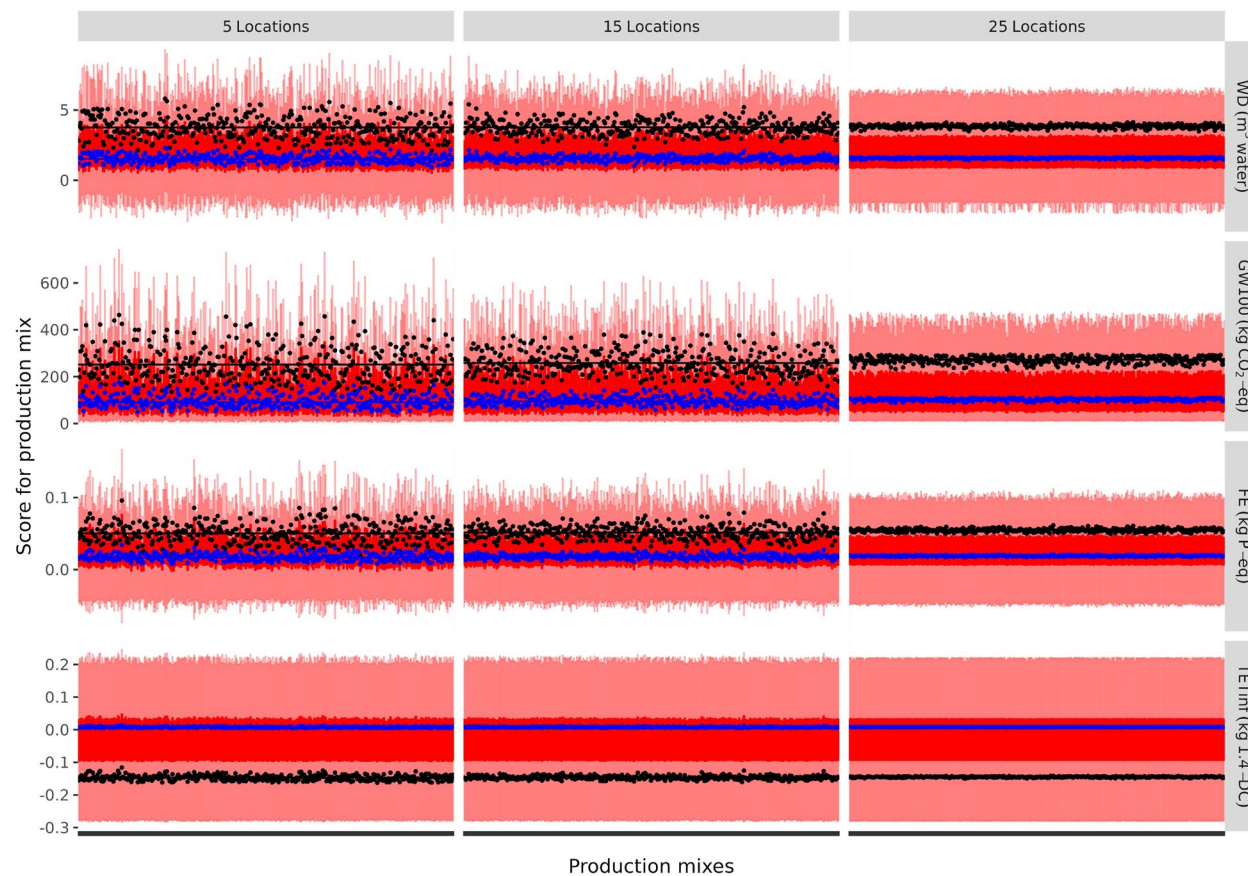


Figure 4: Production mix and bioprospecting uncertainties in aligned boxplots. The red area consists of a succession of narrow boxplots along the horizontal axis corresponding to 400 randomly generated production mixes. The range of each boxplot on the vertical axis corresponds to the dispersion of impact score per strain-compound pair produced in the production mix. The impact score for a strain-compound i in the production mix p is calculated as follows: $SMimp_{p,i} = \frac{\sum_{L=0}^{N_p} Median([imp_{i,L,0}, ..., imp_{i,L,a}, imp_{i,L,150}])}{N_p}$, with N_p the number of locations in the production mix p , L the identifier of a location being part of p and $imp_{i,L,a}$ the impact score calculated for the

production of strain-compound pair i , in location L and in PBR a . The black and blue dots respectively indicate the means and medians of each boxplot. The horizontal black and blues lines respectively represent the mean across boxplots and the means of the medians across boxplots.

3.3 Sensitivity Analysis

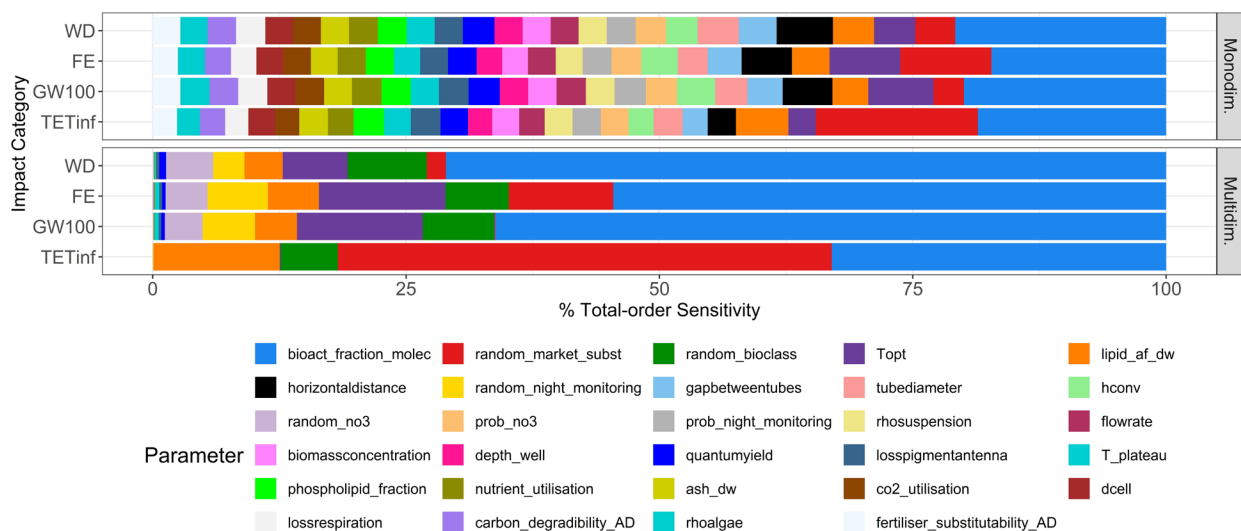


Figure 5: Shares of the Sobol total-order sensitivity for the uncertain parameters in multi-dimensional and mono-dimensional samplings. The parameters are detailed in SI II.1. In mono-dimensional sampling the share of the total-order sensitivity for a parameter x is calculated as

$$Share_x = \frac{\frac{\sum_{L=0}^{28} Indice_{x,L}}{28}}{\sum_{i=0}^p \left(\frac{\sum_{L=0}^{28} Indice_{i,L}}{28} \right)},$$

with $Indice_{x,L}$ the total-order sensitivity indice calculated in location L for parameter x , and p the number of uncertain parameters in mono-dimensional sampling. In multi-dimensional sampling, $Share_x = \frac{Index2_x}{\sum_{i=0}^p Index2_i}$ with $Index2_x$ the total-order sensitivity index

of the biological parameter x associated with the impact for a strain-compound in Europe (cf. Figure 1). The error bars and dispersion of the different indexes are shown in SI I.5.3.

Mono-dimensional and multi-dimensional samplings allow for a multifaceted understanding of the model's sensitivity to the different parameters. Regarding the impact scores associated with one strain-compound pair in one location, Figure 5 shows for mono-dimensional sampling that the uncertainty on the fraction of bioprospected compound in the biochemical class (lipid, carbohydrate, protein) dominates in the output uncertainty for all impact categories. This was expected as this parameter eventually affects the overall content of bioprospected compound in the biomass and therefore substantially influences the reference flows in the product system. The uncertainty on the parameter assigning the substitution route also dominates the output uncertainty for TET and FE, which is to be put in relation to the significantly different LCA profiles of the three substitution routes (cf. 3.1 and Figure S13 in SI I.4.2.4.1, SI I.4.2.3). Eventually, the techno-operational uncertainty on the geometry of the PBR and the resulting culture volume (tube diameter, horizontal distance between stacks, gap between tubes) accounts for 15-20% of the output variance for all impact categories. As also described in our previous work³⁸ this is mainly due to the influence of the PBR volume on the thermoregulation requirements, together with the strain's thermal requirements (T_{opt} , $T_{plateau}$) and location. The 22 other parameters explain around 60% of the output uncertainty, which makes it difficult to decide on which parameters could be fixed to a unique value without losing information on the output uncertainty.

While mono-dimensional-sampling can spot uncertainty hotspots in the details of the model at the techno-operational level, only multi-dimensional-sampling can investigate the sensitivity of the model to a strain-compound pair at the European production level.

The latter as well shows that the fraction of bioprospected compound in the biochemical class (*bioact_fraction_molec*) and the parameter assigning the substitution route (*random_market_subst*) dominate the output uncertainty (75 %). The similar ranking of the common parameters between mono and multi-dimensional sampling shows that the same strain-compound-specific parameters strongly influence the impact scores for a production both in a unique location and when the same strain-compound pair is produced all over Europe.

Interestingly and in accordance with former observations (cf. 3.1), almost 100% of the output uncertainty for TETinf comes from the substitution mechanisms, determined by the biomass composition and substitution route. Thus, one could theoretically provide an educated estimate of the future TETinf impact of the production in Europe as soon as the biomass composition and the substitution market for a newly found strain-compound are known. Nevertheless, this estimate should be done keeping in mind the sensitivity of the model at the techno-operational level revealed by the mono-dimensional sampling.

3.3 Reflecting on Variability and Uncertainty

In this work, we have consistently used the word “uncertainty” to qualify the need to resort to a stochastically sampled set of values instead of using a static set of values for the parameters of our LCA model. The distinction between variability and uncertainty is key within the LCA community and more generally in modeling disciplines which cannot settle for a mere deterministic assessment to support decision-making^{55–57}. While variability is intrinsic to real-world phenomena and processes, uncertainty is often defined as being due to a lack of knowledge about the model and its parameters^{56–58}. This semantic overlapping between variability and uncertainty depending on the formulation of the research question is well described by Frey⁵⁵. Our case confirms and illustrates how uncertainty and variability merge in some cases, depending on the research

question. The geographic variability becomes production mix uncertainty in our ex-ante LCA, as it stems from an irreducible lack of knowledge a priori about the future development of the mix. Similarly, while there is a vast diversity of microalgal strains and compounds, this biological variability translates into uncertainty associated with our research question about the impacts of an increase in demand for *a* microalgal compound that is currently being bioprospected in Europe. The uncertainty is here similar to the one applying to the result of a random draw (aleatory uncertainty) within a diverse population expressing variability (the biodiversity). Finally, the techno-operational conditions should also be understood as part of the uncertainty rather than just variability, as generating random PBR geometries and setups does not aim at representing alternative routes for a same compound production⁵⁷, but instead represent equiprobable scenarios for one strain to reach a specific productivity according to our limited knowledge.

To go further, the multi-dimensional strategy uses two independent loops as in the two-dimensional Monte Carlo simulation proposed by Michiels and Geeraerd⁵⁶ to distinguish variability and uncertainty. The difference is that we use this approach to distinguish bioprospecting uncertainty from techno-operational and production mix uncertainty. By doing so, we neglect biological variability by assuming that, once found and cultivated in a European mix, a strain does not express any phenotypical or genomic plasticity and features the same biological parameters' values in all locations. This is an oversimplification as the same strain producing the same compound in a European mix could for instance accumulate more or less lipids for different locations due to different light regimes⁵⁹. In this sense, mono-dimensional sampling, while making impossible to distinguish between the types of uncertainty, also tackles a less fixist⁶⁰ and therefore more realistic concept of “microalgal strain” by simulating a continuum of biological parameters in a continuum of different PBRs and locations.

3.4 Limits

Our forecast partly relies on our choices regarding the modeling of microalgal biodiversity. Adapting the generic model from Williams and Laurens²⁸ to represent a strain could be argued to be simplistic and not cover the immense diversity of microalgae. For instance, sinking rates can in reality vary from the Stokes' law estimates we used to estimate centrifugation energy consumption⁶¹, depending on the microalgal taxa and cell shapes⁶². Adapting the centrifugation technology may be necessary for some strains⁶³. Furthermore, using ranges and distributions for the biological parameters based on results obtained within the known biodiversity to simulate undiscovered strains could be a good example of survivor bias⁶⁴. In fact, we believe this bias benefits the representation of uncertainty by taking into account the demonstrated difficulty to cultivate many strains which can be observed in their environment^{65,66}. The discovered and upscaled strain will therefore likely be relatively similar to the ones we already know. Finally, and as previously mentioned (cf. 2.2.2), the model would benefit from further refinement of the interactions between techno-operational, biological and geographic variables to limit the weight of unlikely combinations in the uncertainty representation.

3.5 Use of the results for decision-making

The high dispersion of the results associated with a very *large LCA space*⁶⁷ and the complex overlapping of the uncertainties must lead us to question the usability of the estimates for decision making. Ideally, the results should be used for planning and providing insightful indications on whether this technology will likely be beneficial and compete with alternatives. We can summarize the results by using the median impact score per kg of bioprospected compound across production mixes and strain-compound pairs: 1.5 m³ for WD; 96 kg CO₂-eq for GW; 0.017 kg P-eq for FE and 0.007 kg 1.4-DC-eq for TET (cf. Figure 4 and Table S3 in SI I.3 for complete statistical

description). These values, however, are obtained by keeping one median score per strain-compound pair and location, thus aggregating techno-operational uncertainty (cf. Figure 1). A first comparison of magnitudes can be made with other bioactive compounds such as drugs from industrial chemistry whose impacts can range from 30 to 3000 kg CO₂-eq per kg of drug⁶⁸. Overall, if a solution based on a bioprospected microalgal compound was to be compared with an alternative technology for decision-making prior to technology development, the whole distribution of the results should be considered and different statistical measures could be used⁶⁹ (cf. SI I.3).

It must be highlighted that the results presented in this article can be understood as a *null model* by analogy with its use in ecology⁷⁰. Thus, the patterns of the model's output densities are obtained for a set of standards assumptions associated with our current level of knowledge. Additionally, the understanding of the uncertainty propagation combined with the sensitivity analysis allow anticipating the shape of the densities when other assumptions are made or more knowledge is gained. A key assumption supporting the *null model* is that bioengineers will find the combination of photobioreactor geometry and operational setup associated with 30% of the strain-specific energetic yields (cf. 2.2.1), as observed for cultivated strains by Williams and Laurens²⁸. Running the model with more pessimistic or optimistic assumptions regarding the capacity of bioengineers to optimize photobioreactors for specific strains would shift the impact density curves. Another assumption is that there is no restriction on the possible biochemical class of the target compound (protein, lipid, carbohydrate) and content in the biomass. Finally, we do not account for market mechanisms that could trim the density curves by making the worst cases economically non-viable, for instance due to very high energy consumption per functional unit. This assumption can be

qualified as realistic as the context of high-value compounds does not exclude cases with high production costs provided that the market prices of the compounds follow.

4 Outlook

Through a heavy stochastic simulation of microalgae cultivations across strains, technological settings, and locations, this work demonstrates the use of computational resources to investigate the uncertainty associated with the future environmental impacts of a technology at very early stage. The stochastic approach, coupled with an explicit classification and separation of the uncertainties, allowed isolating the most important uncertain parameters but also to understand how techno-operational, bioprospecting and production mix uncertainties interact with each other. It is key to note that, by propagating uncertainty regarding the LCA of one bioprospected microalgal compound, our approach eventually drew the LCA profile of a whole biological group (microalgae) and its sector (productions of high-value microalgal compounds), in a whole market (Europe). An even more accurate LCA portrait of the microalgal high-value compounds sector would benefit from including background uncertainty, but also the extraction procedures which highly depend on the compound and strain. Overall, the approach can be generalized to technologies at a conceptual level of development for which the modelers know enough about the ruling biological and physical phenomena to determine the key model variables, draw dependencies, and eventually parameterize a model in which uncertainties are singled out. The value of the approach is enhanced when simultaneously applied to competing alternatives so that probability estimates resulting from the same method and same understanding of the uncertainties can be compared.

Finally, an additional step towards educated decision-making process and planning would be to use the model to go beyond the presented *null scenario* and propagate uncertainties using

prospective databases for market information and future marginal suppliers, but also climate projections that could substantially influence the forecasts.

ASSOCIATED CONTENT

“Supporting Information I” (docx): Product system, description of additional calculations to the model, additional figures, elementary composition of microalgal molecules’ biochemical classes.

“Supporting Information II” (xlsx):

- **Sheet "SI II.1"**: Table of parameters.
- **Sheet "SI II.2"**: Statistical description of all model's outputs in mono-dimensional sampling.
- **Sheet "SI II.3"**: Correspondence between foreground activities and background activities.

The code of the model allowing reproduction of the figures is available at https://github.com/PJGilmw/Bioprospected_LCA⁴⁰.

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