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Abstract

Danish seining (or anchor seining) is a fishing technique that is gaining increasing attention because it is considered to be a fuel-efficient fishing method with low environmental impact. However, scientific documentation of the selectivity characteristics of Danish seines is lacking, and the gear generally is grouped with bottom trawls and Scottish seines in fisheries management legislation. In this study, we developed a codend cover to estimate the selectivity of a standard commercial Danish seine codend for four fish species. The data for the dominant species, dab (\textit{Limanda limanda}) and plaice (\textit{Pleuronectes platessa}), was best described by models that combine two or three logistic models, which indicated that more than one selection process was at work. Selectivity of cod (\textit{Gadus morhua}) was best described by a Richard curve and selectivity of red gurnard (\textit{Chelidonichthys lucernus}) by a logistic curve. The estimated selectivity curve of dab indicated, contrary to cod and plaice, low retention of individuals below MLS. Confidence limits for larger length classes of cod and red gurnard were relatively wide. For plaice, the estimated selection factor, which is the length with 50\% retention divided by mesh size, was comparable to literature values from trawl studies. The average value for cod was similar for Danish and Scottish seines, but lower for trawls. The results are discussed in the context of fisheries management with focus on the landing obligation of the new Common Fisheries Policy.
Keywords: Demersal seining, Discard ban, Landing obligation, Multiple selection, Selectivity modelling, Skagerrak

1. Introduction

Although a decline in the number of seiners in Denmark is evident (1990: 252; 2000: 118; 2015: 32; EuroStat, 2016), Danish seining is still an important fishing technique. In recent years, interest in Danish seining has increased because it is viewed as a fuel-efficient fishing method (Thrane, 2004) and because its environmental impacts are said to be less than those of other active demersal fishing gears such as beam trawls or bottom trawls (ICES, 2006, 2010; Suuronen et al., 2012; Eigaard et al., 2015).

The main target species of Danish seiners in Skagerrak and the North Sea are flatfish, primarily plaice (Pleuronectes platessa), which has been within safe biological limits for the last three years (ICES, 2015). Nevertheless, there is a general lack of scientific documentation of the selectivity of Danish seines. The sparse existing data (e.g. ICES, 2010; Suuronen et al., 2012) are often based on assumptions or older studies, where other regulations existed, different gears or vessels were used or where data were not analysed following the standards described in Wileman et al. (1996).

A new Common Fisheries Policy that includes a landing obligation (discard ban) system was introduced in most European Union (EU) waters, including Skagerrak and the North Sea, by 1 January 2016 (EEC, 2011, 2012; Condie et al., 2014b; Condie et al., 2014a; Elíasen, 2014; Uhlmann et al., 2014; Sardà et al., 2015). The specific challenge for the industry, and the major difference from the earlier landing quota system is that the catch of all sizes of listed species is counted against the quota. A minimum conservation reference size (MCRS, generally equal to current minimum landing size, MLS) will be introduced for several commercial species and individuals below this size are prohibited from being sold for direct human consumption. Consequently, information about the selective properties of fishing gears is of great importance for the economy and fisheries management as selectivity parameters like L50 (length at which 50% of the fish are retained) and SR (selection range; L75–L25,) give an indication of which sizes of fish can be expected by the fishery. This information is important to estimate
the probability that the fisheries will adhere the objectives of the landing obligation. Furthermore, if the
expectations of the landing obligation are too high (e.g. due to high bycatches of fish below MCRS), the
data may allow for recommendations to be made on how to adjust the fisheries to the new system.

By EU law, Danish seines belong to the same legislative category of fishing gears as Scottish
seines and bottom trawls. All three gears follow the same technical regulations such as mesh size and
selective devices. Several older studies regarding selectivity of Scottish seines exist (Reeves et al., 1992;
Isaksen and Lokkeborg, 1993) but the overall state of knowledge is low. A recent theoretical study by
Herrmann et al. (2015) estimated the selectivity of Scottish seines on the basis of one of those earlier
studies using suitable statistical methods. Nevertheless, they concluded that further studies have to be
conducted using currently used demersal seines. The understanding of selectivity in bottom trawls is
much greater as the majority of selectivity studies for gears from this legislative category focused on
trawls (e.g. Reeves et al., 1992; Graham et al., 2004; Frandsen et al., 2010b; Madsen et al., 2012).

Although the netting materials and codend constructions used in Danish seines, Scottish seines,
and bottom trawls are similar, the gears have pronounced differences in construction and in the way they
are operated. Bottom trawls use trawl doors to spread the net (von Brandt, 2005), and the towing speed
is relatively constant throughout the fishing process. Seiners do not use any doors or other spreading
deVICES, and the speed at which the net is dragged is slower than that in trawling, but it continuously
increases during the fishing process. Scottish seiners move forward during the retrieval process, whereas
Danish seiners do not as they are anchored (von Brandt, 2005). With such pronounced differences in
towing speed and net geometry during the fishing process, it is likely that the selection processes differ
among the three types of gears.

Due to the lack of consistent forward motion in Danish seines, it is important to develop a cover
based on the principles of the conventional codend cover (Wileman et al., 1996) to study the selectivity
of this type of gear. Such a device must cope with the different stages of the fishing process and always
keep the cover a sufficient distance away from the codend to avoid a potential masking effect that can
occur when the cover comes in contact with the meshes of the codend (Madsen and Holst, 2002).
The main objective of this study was to estimate the selectivity parameters for species caught with Danish seines using the codend design currently used in the commercial fishery. These selectivity parameters were compared to those of bottom trawls and Scottish seines, and the results should prove useful in terms of technical regulations and management policies. The data will also be used to evaluate the gear in terms of the landing obligation and to estimate the potential consequences for the Danish seine and bottom trawl sector now, and in the future, should other species be added to the landing obligation list.

2. Materials and Methods

2.1. Study site and experimental setup

The experiments were carried out aboard the commercial Danish seiner HG 35 *Vendelbo* (length overall: 15.47 m, engine power: 91 kW) off the coast of Denmark in Skagerrak (ICES area IIIa; Fig. 1) in August and September 2014. The fishing took place in sandy shallow areas close to the coast (~13 m deep, Hauls 1, 2, 3, 6, 7) that are known to be good grounds for flatfish such as plaice and in deeper grounds (~68 m deep, Hauls 4 and 5) that are known to be good for roundfish such as haddock (*Melanogrammus aeglefinus*).

The vessel’s commercial gear was used, which was representative for the Danish seining fleet that operates in Skagerrak and the North Sea. The seine had 380 meshes (nominal mesh size: 120 mm) around the fishing circle, and it consisted of a wing section with a weighted 43.6 m long ground rope, a belly section, and an extension section. The 7 m long non-tapered codend was made of Nymflex 4 mm double twine polyethylene (PE) netting (mesh size: 124.4 ± 3.0 mm, N = 200, measured with an OMEGA gauge (Fonteyne et al., 2007)) with 97 open meshes around the circumference. The codend was constructed with one selvedge that included three meshes, following commercial practice. Although scientific selectivity studies are normally carried out with newly produced codends without additional devices (e.g., round straps, protecting bags, or flappers) that could affect selectivity, the codend in this
The study was equipped with two round straps (Fig. 2; Herrmann et al., 2006). These two round straps were 1.9 m in circumference and mounted 0.5 m ahead of the codline and 2.9 m in circumference and mounted 1.0 m ahead of the codline. Round straps are widely used by commercial vessels to limit a codend’s circumference just in front of the codline to facilitate fast and more controlled emptying of the codend aboard the vessel, which is thought to improve safety for fishermen handling the gear. However, small variations of the specific mounting of these round straps may occur between vessels. Legal regulations regarding round straps are stated in EU regulation 3440/84. The seine warps used in the current trials were ~2860 m long (13 coils), each with a diameter of 21 mm.

The covered codend method (Wileman et al., 1996) was applied to catch individuals escaping from the codend. The actual cover was 21 m long and consisted of two main parts (part C and D, Fig. 2), but two additional pieces of netting (part A and B, Fig. 2) were necessary to attach the cover appropriately to the extension part of the seine. The 11 m long part C covered the codend and was made of 0.9 mm thin knotless Dyneema (ultra-high molecular weight PE) twine netting in square mesh orientation (mesh size: 46.2 ± 3.0 mm) to ensure good water flow through the meshes and a low visibility of the netting in order to not affect the escape behaviour of the fish. Furthermore, this configuration allowed the meshes to stay in a fixed position and thus maintain a sufficient opening and distance between codend and cover in order to minimize the risk of masking the codend (Madsen et al., 2001). This part consisted of four panels and had 620 mesh bars in circumference (155 per panel). The 10 m long aft part D was made of 2 mm knotless PE netting (mesh size: 40.8 ± 0.7 mm) in diamond orientation. It consisted of two panels and the number of meshes per panel decreased from 175 in the front to 145 meshes per panel in the end. Three kites, consisting of two PVC-coated trapezoidal canvas parts (ca. 0.5 m² per trapezoid) as described by Madsen et al. (2001) were attached to the cover to ensure that it remained open during faster hauling speeds (Figs. 2 and 3). One kite was attached to each of the starboard panel, the portside panel and the top panel (Figs. 2 and 3). Because Danish seines are dragged at a slower speed than trawls, especially in the beginning of the fishing process, several modifications were made to the cover design described in Madsen et al. (2001). These were made to ensure that the cover did not mask the codend.
Twenty-four egg-shaped floats (buoyancy: 0.2 kg) were attached along each upper selvedge of the front part, and lead ropes (1 kg/m) were attached to the lower panel (Figs. 2 and 3). Additionally, a 1.9 m long PE bar was fixed transversally across the upper panel at the point where the kites have been attached (Figs. 2 and 3). This ensured the cover to spread horizontally and thus allowed sufficient horizontal space between the codend and cover when the gear was not moving or was moving very slowly. This minimized the risk of masking. Finally, a ca. 10 m long zipper was inserted in the top panel of part C to allow handling the codend catch first in order to prevent escapes of fish from the codend into the cover at the surface (Fig. 2). Adjustment and inspection of the cover were conducted in a flume tank (SINTEF, Hirtshals, Denmark) prior to the experiments, with participation of scientists, fishermen, and the net maker who created the cover. Velocities from 0 to 1.8 kn (0.9 m/s), equivalent to the speed of the seine when the majority of fish enter the codend (unpublished data, Thomas Noack, DTU Aqua Hirtshals, Denmark), were tested. As the length of the cover exceeded the flume tank’s dimensions, the last part of the cover was bundled for the tests. By doing so, it was still possible to judge and adjust the modifications around the codend (lead ropes, floats, kites, PE bar) in an appropriate way.

2.2. Data collection and sampling strategy

For each haul, fishing time, depth at the position where the net was deployed, depth at anchor and the sea state were recorded following the protocol of Wileman et al. (1996). A GPS-logger (Canmore G-PORTER GP-102+) tracked the vessel’s movement over the entire fishing process for each haul.

When the catch came aboard the vessel, the codend was emptied first to avoid any fish escaping from the codend into the cover. In order to do so, the cover was tightened up to a level that allowed for a proper opening of the zipper without risking any fish to swim or fall out. As soon as this level was reached, the codend was pulled out of the cover. With the exception of the first haul in which the whole catch was sorted prior to subsampling, subsamples were taken from the non-sorted catch due to large amounts of fish (as outlined by Gerritsen and McGrath (2007)). After sorting and identifying species,
fish were measured to the nearest cm. Individual weights were estimated using length-weight relationships (Shanks, 1981; Coull et al., 1989; Marčeta, 2013).

During the second haul, two underwater video cameras (GoPro, Inc. HERO 3+) were mounted between the cover and codend (pointing downstream and upstream) to document the performance of the cover and the behaviour of the fish in the gear during the fishing process.

2.3. Data analysis

Selectivity modelling was conducted to estimate species-specific selectivity curves and selectivity parameters (e.g., L50 and SR) using the computer software SELNET (Herrmann et al., 2012). Hauls with < 10 measured individuals were excluded from further analyses following Krag et al. (2014). The modelling approach followed the procedure described by Sistiaga et al. (2010), Eigaard et al. (2011), Herrmann et al. (2012), and Madsen et al. (2012). In addition to the logistic model (Eq. 1), six other models (Eq. 2-7), including the three other classical size selection models “probit” (Eq. 2), “Gompertz” (Eq. 3) and “Richard” (generalised logistic model with additional asymmetry parameter 1/δ, Eq. 4) were tested within this study. For detailed descriptions of those see Wileman et al. (1996). Additionally, three more complex models that combined two or three logistic models were considered as candidates. Those were the double logistic model “LogitS2” (Eq. 5; Lipovetsky, 2010), the dual selection logistic model “Dual_selection” (Eq. 6; Sistiaga et al., 2010) and the triple logistic model “LogitS3” (Eq. 7; Frandsen et al., 2010a). All models accounted for overdispersion due to haul-pooling. The retention probability r of a fish of length l can be expressed by r(l,u) with u describing a vector that contains parameters needed by the model.

\[ r(l,u) = \]

\[ \text{Logit} (l,L50,SR) \]  
\[ \text{Probit} (l,L50,SR) \]  
\[ \text{Gompertz} (l,L50,SR) \]  
\[ \text{Richard} (l,L50,SR, 1/\delta) \]  

\[ \text{LogitS2} = c_1 \times \text{Logit} (l,L50_1,SR_1) + (1.0 - c_1) \times \text{Logit} (l,L50_2,SR_2) \] (5)
Models that combine two logistic models have been used in previous studies on trawls separating the selectivity process in a towing phase and haul-back phase (Herrmann et al., 2013a). They have also been used in studies on trawls with sorting grids (Kvamme and Isaksen, 2004; Sistiaga et al., 2010; Herrmann et al., 2013b) where the individual fish can escape either through the grid or through the codend meshes. For the double logistic model LogitS2 (Eq. 5) and dual selection model Dual_selection (Eq. 6), the selection process is assumed to consist of two processes. The double logistic model (Eq. 5) combines two logistic models, one for the first process and one for the second process. The contact ratio parameter $c_1$ indicates hereby the probability for an individual to have its selectivity determined by the first process, i.e. the chance of each individual to get in contact with the selective area within the first process (Herrmann et al., 2013a). Consequently, the probability to have its selectivity determined by the second process is $1.0 - c_1$. $L_{50_1}$ and $SR_1$ or $L_{50_2}$ and $SR_2$ describe the selectivity of the according “sub-process”. The dual selection model (Eq. 6) is similar to the double logistic model, but it is a sequential function. This means that the proportion of individuals that try to escape in the second process is assumed to consist of those that did not attempt to escape in the first process and additionally those that attempted to, but were retained. The triple logistic model LogitS3 (Eq. 7) follows the same principles as the LogitS2, but includes a third stage of selection, i.e. it is the sum of three logit models in which the weights of the contributions add up to 1.0 (Frandsen et al., 2010a). Additional parameters required by this model to describe selectivity are $L_{50_3}$ and $SR_3$ explaining the selection in the third “sub-process” and $c_2$ indicating the probability of an individual to have its selectivity determined by the second process. Consequently, the chance of an individual to have its selectivity determined by the third process is $1.0 - c_1 - c_2$.

Selecting the final model for each species followed the procedure of inspecting goodness of fit as described by Wileman et al. (1996) and by comparing individual Akaike information criterion (AIC)
values (Akaike, 1974). If the fit statistics indicated a lack of model fit, i.e. p-value close to zero, deviance
>> degrees of freedom or low R²-value (ratio of variance explained by model and observed variance),
residuals were inspected for any structural deviation to determine if structural problems with the model
or overdispersion of the data (McCullagh and Nelder, 1989) were causing such results. Uncertainties
were estimated by calculating Efron 95% confidence intervals (CIs; Efron, 1982) for the final selectivity
curves and selectivity parameters.

Selectivity estimates were compared to values from previous studies of Scottish seines and
trawls in the Northeast Atlantic mixed fishery using the estimated selection factor (SF = L50/mesh size).
All studies used codends similar to the one used in the present study (mesh orientation: diamond meshes;
mesh size: 90–150 mm; twine: 4–6 mm double twine; no grids or release panels), all studies analysed
data following guidelines set by Wileman et al. (1996) and all studies were published in peer-reviewed
scientific journals. To account for differences in mesh size measurements due to the use of different tools
(ICES gauge, EU wedge, OMEGA gauge), values were standardized to EU wedge values (wedge =
0.974*OMEGA + 2.96, derived from Ferro and Xu (1996) and Frandsen et al. (2009); wedge = 1.01
ICES + 2.96 (Ferro and Xu, 1996)).

All analyses other than the modelling approach were performed using R Statistical Software (R
Core Team., 2015).

3. Results

3.1. Haul and catch overview

Seven valid hauls were conducted (Table 1), which took between 121 and 140 min from setting
out the anchor until the gear was retrieved. Each haul covered an area between 2.58 and 3.04 km², and
depths varied between 7 and 82 m. Catches ranged from 65 to 1503 kg in the codend and from 327 to
8415 kg in the cover. Thirty-one different fish species were caught in this study and the majority of the
catch was composed of dab (*Limanda limanda*) and plaice. Other species investigated within this study were cod (*Gadus morhua*) and red gurnard (*Chelidonichthys lucernus*).

The inspection of the cover in the flume tank and the underwater recordings from haul 2 indicated that the cover did not mask the codend at any speed within the tests or at any stage of the fishing process in the observed haul. Fish escaping from the codend were not observed to swim back into the codend, although they could easily do so because of the slow towing speed. The observations indicated that the majority of the catch entered the gear relatively late in the catching process. All fish seemed to be in good condition during the whole fishing process and during the handling of the catch on-board.

**3.2. Selectivity estimations and length distributions**

Selectivity curves and parameters were estimated for dab, cod, plaice and red gurnard (Table 2, Table 3). Low numbers of individuals, in combination with relatively high proportions of small fish, resulted in high levels of uncertainty in the analyses. This prohibited an appropriate estimation of selectivity parameters for the other species. A rather high proportion of small fish was also evident for all species where selectivity analyses were possible as the number of individuals in the codend represented only a small part of the total catch (Fig. 4), indicating high numbers of fish escaping into the cover.

A Richard curve with relatively smooth rise (Fig. 4) described the selectivity of cod best (lowest AIC value). The model fit was acceptable (p-value = 0.81, deviance = DOF (Degree of freedom), R² = 0.93; Table 3). Confidence intervals became relatively wide for a range of length classes where the number of observed individuals was low up to length classes with a retention probability of 1.0. The estimated average L50 of 41.6 cm was higher than the current MLS and had, like the estimated SR (12.6 cm), relatively wide confidence limits (Table 3, Fig. 4).

The selectivity of dab was best described by a triple logistic model (Fig. 4) and the model fit was good (p-value = 0.35, deviance = DOF, R² = 1.00; Table 3). Most observed individuals were found in
length classes below the selective area of the gear, but almost all of them were larger than the current MLS of 25 cm (Fig. 4). The selectivity curve itself was steep with narrow confidence limits. L50 was estimated to be larger than the current MLS of 25 cm (31.2 on average, Table 3) and SR was found to be narrow (0.8 cm, Table 3).

A double logistic model best described the selectivity of plaice. Model fit parameters were good (p-value: 0.84, deviance = DOF, R^2 = 1.00; Table 3). Most individuals belonged to length classes of the lower range of where selectivity took place, but confidence limits of the steep curve were narrow for all length classes. The current MLS of 27 cm fell within the selective area and laid within the confidence limits for the estimated L50 (average = 29.1 cm, Table 3). SR was estimated to be 2.2 cm (Table 3).

The selectivity of red gurnard as the only species without MLS (Table 2) could be best described by a logistic model. Since the low p-value (0.00) indicated a potential lack of model fit (Table 3), the residuals were investigated. As structures were not detected, it was assumed that overdispersion was at fault and the model could be applied with confidence. The curve had a smooth rise, but was – especially for length classes with retention probabilities above 0.5 – characterized by few observations and wide confidence limits. The estimated L50 and SR values were 31.0 cm and 11.5 cm, respectively (Table 3).

**Discussion**

The goal of this study was to investigate codend selectivity characteristics for several species of fish in a commercial Danish seine as it is currently used in the Danish fishery off the coast of Denmark. An important part of the experimental work was the development of a covered codend methodology that functions at varying towing speeds but particularly at low or no speed. Both flume tank observations and underwater observations indicated that the current approach of combining floats, weights, a distance bar, and kites with a cover made of four panels functioned very well. Thus, this methodology could be applicable in other, similar fisheries where towing speeds are low and variable.

The commercial Danish seine used in this study usually included two rear round straps. Round straps could reduce the mesh opening in a codend and hence the size selectivity by reducing L50, as
demonstrated by Herrmann et al. (2006) in a simulation study of haddock in trawls. For flatfish, where
the morphology of the fish fits a low mesh opening angle, theoretically, the reverse effect could be
expected. Because of this, the comparison among trawls, Scottish seines, and Danish seines could be
influenced by the round straps, as previous studies used codends without any additional devices.
However, effects of other selectivity-influencing factors, such as catch rates, are considered to be
stronger than the effects of round straps (Herrmann et al., 2006).

It was possible to estimate selectivity curves for 4 of the 31 caught fish species. The codend
mesh size was relatively large, which resulted in low retention for most species. Furthermore, catches of
many non-target species were low. For red gurnard, a mismatch between the caught population structure
and the selective area of the mesh size was observed, i.e. most observed fish were between 10 cm and
30 cm, but our model found that full retention was not obtained below 40 cm. For cod, which can grow
bigger, the catches were low, especially for larger length classes. This resulted in wide confidence limits
of L50 and SR for cod as well as for red gurnard. Therefore, the SF values estimated for cod (3.4), which
were on average similar to Scottish seines (3.2), but higher than for trawls (2.4; Table 4), should be used
with caution. Future studies should focus on providing stronger selectivity estimates for cod and other
species that can grow to sizes that are within the selective area of the gear.

Plaice is the most important species in the Danish seine fishery and, as it is also the case for cod,
retention probabilities of small individuals were relatively high. The selectivity curve for plaice indicated
a mismatch between the curve and the current MLS, which means that some plaice below MLS were
retained. The estimated SF value for Danish seines (2.3) was slightly higher than the mean value of
previous trawl studies (2.2), but within their range (2.0 – 2.3; Table 4). This indicates similar amounts
of fish below MLS (MCRS) being caught by both gears, which would be discarded today. Although
discarded plaice may survive (van Beek et al., 1990), they will have to be brought to land within the
landing obligation system and catches will be deducted from the fishermen´s quota. However, earnings
of these smaller fish are likely low as it will be prohibited to sell fish below MCRS for direct human
consumption. The current results would indicate potential consequences of the upcoming landing
obligation system in terms of catches of smaller plaice to be relatively similar for Danish seiners and
trawlers in this area. Uhlmann et al. (2014), however, reported generally lower discard rates for Danish
seiners than for trawlers in the Skagerrak/North Sea and other European waters, indicating that in general
lower amounts of fish below MLS (MCRS) are caught by the Danish seine fishery. Considering the
results of this more general study, the consequences of the change to the landing obligation system are
likely to be more pronounced in the trawl fishery. Expectable expenditures are, for instance, the
separation of the less valuable catch from the catch with fish above MRCS, the storing of the less valuable
part of the catch on board (Sardà et al., 2015) and ultimately the sale of it. As retention probabilities for
fish below MLS (MCRS) are similarly high, cod may also become a problematic species within the
landing obligation system, but indicated by the smaller average SF value, consequences may again be
more pronounced for bottom trawlers. Expenditure in terms of catches of dab and red gurnard are likely
to be low as retention probabilities for dab below MLS (MCRS) are very low and red gurnard will still
be permitted to be thrown back to sea as it is not part of the list of species that are prohibited to be
discarded within the landing obligation.

The selectivity of the two species with the strongest data, dab and plaice, was best described by
models indicative of a multiple selection process. Similar models have so far been used when considering
the selectivity process in trawling to consist of two or more processes, e.g. when separating the process
into towing phase and haul-back phase (Herrmann et al., 2013a) or when using selective devices in
addition to the codend (Kvamme and Isaksen, 2004; Sistiaga et al., 2010; Herrmann et al., 2013b).
Various factors (e.g., mesh opening or tension in the codend meshes) may, however, affect selectivity
characteristics during the fishing process of Danish seining in a similar way and could result in multiple
selection processes. For example, increasing hauling speed over time may result in a change of the
selectivity characteristics of the codend, as the increasing speed may involve more traction on the gear
and on the meshes. The video recordings, however, indicated that most fish entered the seine late during
the capture process, thus the number of escapees in the period of slow speeds should be low. Herrmann
et al. (2015) suggested that taking the catch from a Scottish seine aboard in several batches leaves fish
in the codend and extension, where they may be subjected to tightening and relaxing meshes due to wave
movement. This could cause a constant switch from stiff to slack meshes, which in turn could change
selectivity characteristics at the surface and between the underwater and surface parts of the fishing
process. However, catches in the current study were small enough to lift on board at once in most cases.
Slack meshes may also occur when the seine ropes are retrieved and the seine needs to be stopped in
order to be detached from the ropes and attached to the net drum for final retrieval. In contrast to a
trawler, a Danish seiner is anchored at this time, and this stop leads to a complete standstill of the gear.
Slack meshes in combination with lively fish that are in the seine for only a short period compared to
fish in a trawl may explain the observed multiple selection in the Danish seine fishery. Therefore, more
complex models that include dual or multiple models should be considered when describing selectivity
of a Danish seine. Such approaches may result in different selectivity curves or different selectivity
parameter estimates compared to those generated by the more traditional logistic models (Herrmann et
al., 2016).

The selectivity estimates generated in this study provide some initial information about several
fish species for which selectivity data have not been collected previously for Danish seines (all species)
or any other fishing gears (dab, red gurnard). This information is important for assessing the ecosystem
effects of fishing gears, for reference when issuing certificates for sustainable fisheries, and for
evaluating the EU landing obligation system which requires the entire catch of listed species to be
counted against a quota. To gain more knowledge about species that were observed in too few amounts
within this study, more experiments need to be conducted, whereby it may be necessary to use non-
commercial codends with smaller mesh sizes to retain more individuals in the codend.

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part of the Skånfisk Project with financial support from the Danish Ministry of Food, Agriculture, and Fisheries.

References


Figure captions

Figure 1. Area and vessel tracks for the seven hauls conducted for the codend selectivity trials aboard the HG 35 Vendelbo in 2014.

Figure 2. Schematic drawing of the codend cover and its attachment to the seine including information about modifications to account for different stages of the fishing process (kites, floats, lead ropes, PE bar). Information about netting and number of panels/selvedges in the specific parts is also included (cross sections in top of drawing). A and B are necessary parts to attach the cover to the seine net. C represents part of the cover around the area of the codend where the main selection is expected to take place and D serves for storing the fish in the cover. 1: Kites. 2: 1.9 m PE bar (transversal). 3: Floats. 4: Zipper. 5: 3 m long lead rope. 6: 1.7 m long lead rope (transversal). 7: 2.1 m long lead rope.

Figure 3. Preliminary assessment of the cover with wrapped rear part in flume tank. For clarification of parts of cover and items attached to it see Fig. 2 and section 2.1.

Figure 4. Selectivity curves for fish including 95% confidence intervals (grey shaded areas), length-specific retention rates (white diamonds), current species-specific MLS if available (vertical stippled line), and length distributions (stippled line: total; solid line: codend). Numbers in parentheses indicate number of hauls used for analysis (i.e. those that had > 10 measured individuals).
Figure 1:
Figure 2:
Figure 3:
Figure 4:
Table 1. Overview of hauls conducted for the codend selectivity trials aboard the HG 35 *Vendelbo* in 2014, including information about time, haul conditions, and total catches. Duration describes time from setting anchor until gear was retrieved aboard the vessel. Depth is for the position where the anchor was set and where the seine was deployed. Sea states as described by Wileman et al. (1996).

<table>
<thead>
<tr>
<th>Haul</th>
<th>Date</th>
<th>Duration (min)</th>
<th>Covered area (km²)</th>
<th>Depth (m)</th>
<th>Sea state</th>
<th>Total Catch (kg)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td></td>
<td></td>
<td></td>
<td>Anchor</td>
<td>Seine</td>
<td>Codend Cover</td>
</tr>
<tr>
<td>1</td>
<td>29.08.2014</td>
<td>136</td>
<td>2.69</td>
<td>25.6</td>
<td>18.3</td>
<td>1503 8415</td>
</tr>
<tr>
<td>2</td>
<td>01.09.2014</td>
<td>136</td>
<td>2.85</td>
<td>12.8</td>
<td>9.1</td>
<td>198 1328</td>
</tr>
<tr>
<td>3</td>
<td>01.09.2014</td>
<td>137</td>
<td>3.04</td>
<td>12.8</td>
<td>12.8</td>
<td>207 1275</td>
</tr>
<tr>
<td>4</td>
<td>02.09.2014</td>
<td>122</td>
<td>2.83</td>
<td>65.8</td>
<td>82.3</td>
<td>512 1174</td>
</tr>
<tr>
<td>5</td>
<td>02.09.2014</td>
<td>121</td>
<td>2.58</td>
<td>65.8</td>
<td>56.7</td>
<td>470 1068</td>
</tr>
<tr>
<td>6</td>
<td>03.09.2014</td>
<td>140</td>
<td>2.93</td>
<td>7.3</td>
<td>11.0</td>
<td>65 327</td>
</tr>
<tr>
<td>7</td>
<td>03.09.2014</td>
<td>135</td>
<td>2.82</td>
<td>7.3</td>
<td>12.8</td>
<td>69 1023</td>
</tr>
</tbody>
</table>

Table 2. Analysed catch data including information about length range, number of measured individuals, and sampling ratio. Current MLS (minimum landing size; if available) is given in parentheses. * indicates species that will have a minimum reference conservation size in the future. NA indicates that there is no MLS present for this species.

<table>
<thead>
<tr>
<th>Species</th>
<th>Length range (cm)</th>
<th>Codend</th>
<th>Cover</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td></td>
<td>No. measured</td>
<td>Sampling ratio</td>
</tr>
<tr>
<td>Cod (30 cm)*</td>
<td>10 - 78</td>
<td>620</td>
<td>1</td>
</tr>
<tr>
<td>Dab (25 cm)</td>
<td>9 - 36</td>
<td>1053</td>
<td>1</td>
</tr>
<tr>
<td>Plaice (27 cm)*</td>
<td>9 - 51</td>
<td>2937</td>
<td>0.353</td>
</tr>
<tr>
<td>Red gurnard (NA)</td>
<td>9 - 41</td>
<td>427</td>
<td>1</td>
</tr>
</tbody>
</table>

Table 3. Summary of model parameters selectivity parameters (L50 as length with 50% retention, SR as selection range) with 95% confidence limits, name of model used, and values describing goodness of fit (DOF = degree of freedom). See sections 2.3 and 3.2 for explanations of selectivity parameters and model fit values.

<table>
<thead>
<tr>
<th>Parameters</th>
<th>Cod</th>
<th>Dab</th>
<th>Plaice</th>
<th>Red gurnard</th>
</tr>
</thead>
<tbody>
<tr>
<td>L50</td>
<td>41.6 (27.2 - 46.4)</td>
<td>31.2 (29.6 - 31.6)</td>
<td>29.1 (28.7 - 30.1)</td>
<td>31.0 (28.6 - 38.7)</td>
</tr>
<tr>
<td>SR</td>
<td>12.6 (4.8 - 16.0)</td>
<td>0.8 (0.1 - 2.7)</td>
<td>2.2 (1.7 - 3.6)</td>
<td>11.5 (7.9 - 26.6)</td>
</tr>
<tr>
<td>1/δ</td>
<td>0.5 (0.1 - 1.3)</td>
<td>-</td>
<td>-</td>
<td>-</td>
</tr>
<tr>
<td>L50₁</td>
<td>-</td>
<td>31.3 (30.4 - 148.6)</td>
<td>29.4 (29.1 - 30.5)</td>
<td>-</td>
</tr>
<tr>
<td>SR₁</td>
<td>-</td>
<td>0.5 (0.1 - 59.5)</td>
<td>1.4 (1.0 - 10.4)</td>
<td>-</td>
</tr>
<tr>
<td>L50₂</td>
<td>-</td>
<td>29.8 (16.1 - 31.3)</td>
<td>25.5 (20.0 - 29.7)</td>
<td>-</td>
</tr>
<tr>
<td>SR₂</td>
<td>-</td>
<td>2.2 (0.1 - 20.3)</td>
<td>6.5 (1.6 - 11.0)</td>
<td>-</td>
</tr>
<tr>
<td>L50₃</td>
<td>-</td>
<td>28.0 (0.1 - 30.0)</td>
<td>-</td>
<td>-</td>
</tr>
<tr>
<td>SR₃</td>
<td>-</td>
<td>15.1 (0.1 - 100.0)</td>
<td>-</td>
<td>-</td>
</tr>
<tr>
<td>Contact ratio</td>
<td>0.7 (0 - 1.0)</td>
<td>0.7 (0.1 - 0.9)</td>
<td>-</td>
<td>-</td>
</tr>
</tbody>
</table>
Contact ratio 2 - 0.2 (0 - 1.0) - - -  
Model Richard LogitS3 LogitS2 Logit 
P-value 0.8101 0.3499 0.8423 0.0000 
Deviance 45.70 21.92 26.69 71.67 
DOF 55 20 35 24 
$R^2$-value 0.93 1.00 1.00 0.33  

Table 4. Comparison of estimated selection factors (SFs) between this study and previous selectivity studies of Scottish seines and trawls. Data values are mean and range.

<table>
<thead>
<tr>
<th>Species</th>
<th>SF - present study</th>
<th>SF - former studies</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>Danish seine</td>
<td>Scottish seine</td>
</tr>
<tr>
<td>Cod</td>
<td>3.4</td>
<td>3.2 (2.0 – 3.8)</td>
</tr>
<tr>
<td></td>
<td></td>
<td>1.2</td>
</tr>
<tr>
<td>Plaice</td>
<td>2.3</td>
<td>-</td>
</tr>
<tr>
<td></td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

1Reeves et al., 1992; 2Isaksen and Løkkeborg, 1993; 3Graham et al., 2004; 4Madsen and Stæhr, 2005; 5Frandsen et al., 2009; 6Frandsen et al., 2010; 7Frandsen et al., 2011; 8Madsen et al., 2012